



Vertical and temporal variations of soil bacterial and archaeal communities in wheat-soybean rotation agroecosystem

Mika Yokota¹, Yupeng Guan², Yi Fan², Ximei Zhang² and Wei Yang²

¹Dulwich College Beijing, Beijing, China

²Institute of Environment and Sustainable Development in Agriculture, Chinese Academy of Agricultural Sciences, Beijing, China

ABSTRACT

Soil microbes are an essential component of terrestrial ecosystems and drive many biogeochemical processes throughout the soil profile. Prior field studies mainly focused on the vertical patterns of soil microbial communities, meaning their temporal dynamics have been largely neglected. In the present study, we investigated the vertical and temporal patterns of soil bacterial and archaeal communities in a wheat-soybean rotation agroecosystem at a depth of millions of sequences per sample. Our results revealed different vertical bacterial and archaeal richness patterns: bacterial richness was lowest in the deep soil layer and peaked in the surface or middle soil layer. In contrast, archaeal richness did not differ among soil layers. PERMANOVA analysis indicated that both bacterial and archaeal community compositions were significantly impacted by soil depth but unaffected by sampling time. Notably, the proportion of rare bacteria gradually decreased along with the soil profile. The rare bacterial community composition was the most important indicator for soil nutrient fertility index, as determined by random forest analysis. The soil prokaryotic co-occurrence networks of the surface and middle soil layers are more connected and harbored fewer negative links than that of the deep soil layer. Overall, our results highlighted soil depth as a more important determinant than temporal variation in shaping the soil prokaryotic community and interspecific interactions and revealed a potential role of rare taxa in soil biogeochemical function.

Submitted 30 August 2021

Accepted 10 January 2022

Published 10 February 2022

Corresponding author

Wei Yang, yangwei_85@163.com

Academic editor

Erica Lumini

Additional Information and
Declarations can be found on
page 13

DOI 10.7717/peerj.12868

© Copyright
2022 Yokota et al.

Distributed under
Creative Commons CC-BY 4.0

OPEN ACCESS

Subjects Agricultural Science, Microbiology, Molecular Biology, Soil Science

Keywords Bacteria, Archaea, Vertical distribution, Temporal variations, Soil

INTRODUCTION

Soil microbes are an essential component of terrestrial ecosystems and drive many biogeochemical processes throughout the soil profile. They are, therefore, closely linked to soil fertility and crop production (Cao *et al.*, 2012; Fierer, 2017; Jiao *et al.*, 2018; Chaudhary *et al.*, 2021). In agroecosystems, crop roots are mainly distributed in the surface soil layer, though they generally extend below 20 cm, even to depths as deep as 150 cm (Hao *et al.*, 2020). Deep soil is important because it harbors more than two-thirds of the total soil organic carbon and nearly equal amounts of phosphorus (Jobbágy & Jackson, 2000), serving as a key nutrient reservoir for crops. Alternatively, crop roots may absorb water from deep soil during dry seasons when surface soil dries out quickly (Hao *et al.*, 2020). Despite so,

previous investigations of soil microbial diversity and community composition are mainly restricted to the surface soil layer (top ~20 cm), limiting our understanding of microbes' potential functions in deep soil. Thus, determining the vertical patterns of soil microbial diversity and community composition would improve our understanding of the processes involved in agricultural soil nutrient cycling and contribute to agroecosystem productivity.

The vertical soil profile is highly heterogeneous, with drastic differences in soil texture, nutrient availability, and organic carbon levels among different soil depths (Kim *et al.*, 2016; Jiao *et al.*, 2018). Previous surveys of vertical microbial distributions revealed that soil depth could be a key determinant in shaping microbial diversity and community assembly in alpine meadows, forests, and agroecosystems (Eilers *et al.*, 2012; Hao *et al.*, 2020; Xu *et al.*, 2021). However, existing surveys of vertical microbial distributions have mainly employed one-time sampling, only capturing a specific status of soil microbes when soil microbial communities could be highly dynamic and show temporal patterns (Yang *et al.*, 2019; Yang *et al.*, 2020; Kivlin & Hawkes, 2020). In previous meta-analyses, soil microbial biomass and community composition were temporally variable across the globe (Shade *et al.*, 2013; Wardle, 1998). Therefore, incorporating temporal variations could be crucial to understanding the vertical distributions of soil microbes.

Although bacteria and archaea are abundant in soil, they occupy different ecological niches and are sometimes filtered by different soil variables (Wei *et al.*, 2020). Soil pH was identified as the primary ecological filter for bacterial communities, while soil C/N ratio or salinity were the primary ecological filters for archaeal communities (Auguet, Barberan & Casamayor, 2010; Bates *et al.*, 2011). These differences lead to contrasting vertical distribution patterns between archaea and bacteria, corroborated by previous reports. Bacterial diversity generally decreased towards deeper soil, while archaeal diversity increased along with soil depth in the desert (Wang *et al.*, 2021) and alpine ecosystems (Xu *et al.*, 2021). However, others reported that bacterial and archaeal richness negatively correlated with soil depth in paddy soils (Yuan *et al.*, 2020). The varying conclusions of these studies emphasize the need for further comparisons between the vertical distribution patterns of archaeal and bacterial communities.

Microbial communities typically show a skewed species abundance distribution, with relatively few abundant species co-existing with many rare species (Jousset *et al.*, 2017). Previous studies have indicated that abundant and rare species may possess different functional traits (Jiao & Lu, 2020). Although key roles of abundant microbial taxa in soil function have been well understood, relatively less attention has been paid to rare taxa (Chen *et al.*, 2020). Rare taxa are important reservoirs of genetic diversity, providing functional redundancy and maintaining ecosystem stability (Jousset *et al.*, 2017). In addition, rare taxa may be functionally dissimilar to the abundant members (Liang *et al.*, 2020), offering complementary functions or unique metabolic pathways to support the overall community functioning. Despite their importance, rare species in soil are often neglected due to low sequencing coverages. Therefore, deep sequencing with millions of reads would provide a higher resolution for the detection of rare species and a more comprehensive depiction of the overall soil microbial community along the soil profile.

In the present study, we investigated the vertical and temporal patterns of soil bacterial and archaeal communities in a wheat-soybean rotation agroecosystem at a scale of millions of sequences per sample. We aimed to determine the following: (1) the key determinant from soil depth and temporal fluctuations in shaping bacterial and archaeal communities; (2) differences in the vertical distribution patterns between soil bacteria and archaea; (3) whether rare taxa occupy a larger proportion in the surface soil layer than in the deep soil layer.

MATERIALS AND METHODS

Field description

Our field experiment was carried out at the Chinese Academy of Agricultural Sciences, located in the North China Plain (39.97N, 116.33E), an important food-producing region in China. This region has a warm temperate monsoon climate, with an average annual temperature of 11–13 °C and average annual precipitation of 500–700 mm (80% rainfall occurs from July to August). The soil has a loamy texture and belongs to the fluvo-aquic soil class. The field was fertilized before wheat planting (September 10) under the wheat-soybean rotation system; other agricultural practices, including weeding and tillage, were identical to those done by the local agricultural management.

Soil sampling and physicochemical analysis

Soil sampling was conducted on October 18 (wheat season) and December 18 (wheat season) of 2019, and on April 17 (wheat season) and August 17 of 2020 (soybean season). In brief, six plots (1 × 1 m) were randomly selected within an area of 20 × 20 m in our study site. For each plot, five soil cores were collected between two plant individuals from a 100 cm long vertical profile that corresponds to depths of 0–20 cm (surface soil layer), 20–50 cm (middle soil layer), and 50–100 cm (deep soil layer). Therefore, each soil sample was a mixture of five soil cores (3.8 cm diameter) for a given soil layer. 72 soil samples (6 plots × 4 sampling times × 3 soil depth) were collected from the six sampling sites. One part of the as-collected soil samples was filtered through a sieve of less than two mm and kept at –80 °C for subsequent extraction of DNA. The other part was sieved with less than one mm and less than 0.25 mm then stored at 4 °C or room temperature to determine the soil physicochemical properties. Soil physicochemical characteristics including soil total carbon (TC), total nitrogen (TN), dissolved organic carbon (DOC), dissolved organic nitrogen (DON), nitrate (NO₃⁻-N) and ammonium (NH₄⁺-N), moisture, pH, clay, silt, and sand content were determined as described by *Bao (2000)* and *Yang et al. (2017)*.

Miseq sequencing and bioinformatics

The soil DNA was extracted using a PowerSoil DNA Isolation Kit (MO BIO Laboratories, USA). The V4 region of the 16S rDNA was amplified using primer pairs 515F/806R (*Caporaso et al., 2012*). Primer 515F contained a 12 bp barcode unique to each sample for Miseq sequencing detection. All PCR reactions followed *Caporaso et al. (2011)* within a 25 µL reaction system and were amplified in triplicate. The PCR products were then pooled, purified, and sequenced on an Illumina MiSeq platform at Majorbio Biotech Co.,

Ltd. (Shanghai, China). The raw sequence data have been deposited on the NCBI SRA (accession No. [PRJNA766099](#)).

The raw sequences were trimmed to the shortest sequence length using QIIME 2 Pipeline Version 1.8.0 ([Bolyen et al., 2018](#)). Subsequently, the sequences were dereplicated, with all singletons discarded. Sequences were then error-filtered and grouped into amplicon sequence variants (ASVs) using the Deblur software ([Amir et al., 2017](#)), with ASVs containing less than two reads removed. Further, potential chimeras were discarded using the Vsearch software. The number of sequences per sample was rarefied to 1,000,000 using the “vegan” package in R. To annotate their taxonomy, the ASVs were blasted against the silva 16s database. Then, the ASV abundance tables were rarefied at 845,591 for bacteria and 10,413 for archaea to ensure even sampling depth within each prokaryote group.

Data analysis

ASVs with relative abundances below 0.01% were defined as “rare” taxa, while those with relative abundances above 0.1% were defined as “abundant” taxa ([Jiao & Lu, 2020](#)). The prokaryotic functional profiles were predicted using the FAPROTAX ([Louca, Parfrey & Doebeli, 2016](#)). Soil fertility index is a synthetic variable calculated from the sum of z-score transformation of TN, TC, DON, DOC, NH_4^+ -N, and NO_3^- -N. Bacterial and archaeal alpha-diversity indices, such as richness and Pielou evenness index, were calculated using the “vegan” package ([Oksanen et al., 2013](#)). Two-way analyses of variance (ANOVAs) were then performed to examine the effects of soil depth, sampling time, and their interaction on alpha-diversity indices and the proportion of rare ASVs. All data were tested for normality and homogeneity of variance before two-way ANOVAs. A post-hoc test was used to determine paired comparisons among the treatments at a 5% significance. Random forest analysis ([Breiman, 2001](#)) was used to explore the soil physicochemical drivers of bacterial and archaeal richness and evenness using the “randomForest” package ([Liaw & Wiener, 2002](#)). The “rfPermute” package ([Archer, 2016](#)) was then utilized to estimate the significance of important metrics for a random forest model by permuting the response variable.

The effects of soil depth, sampling time, and their interaction on soil bacterial and archaeal community compositions were evaluated using a permutational multivariate analysis of variance (PERMANOVA) with 999 permutations in the “vegan” package ([Oksanen et al., 2013](#)). The bacterial and archaeal community compositions were subsequently ordinated using principal coordinates analysis (PCoA) based on the Bray–Curtis dissimilarity matrices in the “vegan” package. A ternary plot was used to demonstrate the distribution of bacterial and archaeal ASVs along the vertical soil profile.

Soil prokaryotic co-occurrence networks for the surface, middle, and deep soil layers were built. Each network was based upon 24 prokaryotic soil communities. ASVs with relative abundances > 0.5% that also occurred in > 50% of the communities were included in the networks to focus solely on abundant ASVs. Spearman’s correlation coefficients were calculated between ASVs using the “Psych” package ([Revelle, 2015](#)). *P*-values for multiple tests were calculated using the false discovery rate (FDR), as described by [Benjamini & Hochberg \(1995\)](#). The correlations with a Spearman’s coefficient < 0.8 and a *P*-value > 0.001

Table 1 Two-way ANOVA examining the effects of soil depth and sampling time on bacterial and archaeal richness (S) and Pielou evenness (J) indices.

		Soil depth (D)		Sampling time (T)		D × T	
		F	P	F	P	F	P
Bacteria	S	11.97	<0.001	1.70	0.18	2.11	0.07
	J	13.60	<0.001	5.50	0.002	1.28	0.28
Archaea	S	1.10	0.34	1.60	0.20	1.27	0.28
	J	121.34	<0.001	0.13	0.94	0.63	0.71

were eliminated (*Widder et al., 2014*). Subsequently, the number of positive and negative correlations, average degree, connectedness, and modularity was calculated in each network using the “igraph” packages. Then, threshold values of Z_i and P_i were computed following the reference (*Guimera & Amaral, 2005*) for each node to determine their topological role. All analyses above were carried out in R (v.3.6.2).

RESULTS

Sequencing data analysis and prokaryotic diversity

A total of 185,093,268 reads were obtained after quality control and chimera checks, assigned as 225,608 ASVs. The read numbers were normalized to 1,000,000 for each sample, resulting in a normalized dataset containing 163,862 prokaryotic ASVs. Then, the read numbers of bacteria and archaea were normalized to 845,591 and 10,413 after species annotation, resulting in two normalized datasets containing 162,802 bacterial ASVs and 213 archaeal ASVs. Among these ASVs, 7,382 ASVs (4.5%) occurred in at least half of all samples.

The bacterial and archaeal diversities were assessed using the richness and Pielou evenness indices. One of the most interesting findings was that the vertical pattern of soil bacterial and soil archaeal diversity differed. Two-way ANOVA analysis indicated that both bacterial richness and evenness were significantly affected by soil depth (*Table 1*). Generally, bacterial richness and evenness were lowest in the deep soil layer and peaked in the surface or middle soil layer (*Figs. 1A* and *1C*). In contrast, archaeal richness was unaffected by soil depth, while its evenness showed a decreasing trend along with the soil profile (*Figs. 1B* and *1D*). We did not find an apparent temporal variation of archaeal richness or evenness (*Table 1*). Although bacterial richness did not show temporal variation (*Table 1*), its evenness index was significantly higher in August than in December (*Fig. 1C*). Random forest analysis indicated that TN, TC, and sand content were the key determinants for bacterial and archaeal diversities (except for archaeal richness, *Figs. 1E–1H*). The soil physicochemical variables are presented in *Table S1*.

Soil prokaryotic community composition

PERMANOVA analysis indicated that both bacterial and archaeal community compositions were significantly impacted by soil depth (Bacteria: $r^2 = 0.38$, $P < 0.001$; Archaea: $r^2 = 0.70$, $P < 0.001$), but unaffected by sampling time and their interaction. This pattern was further evidenced by PCoA ordination based on Bray–Curtis dissimilarity, which indicated

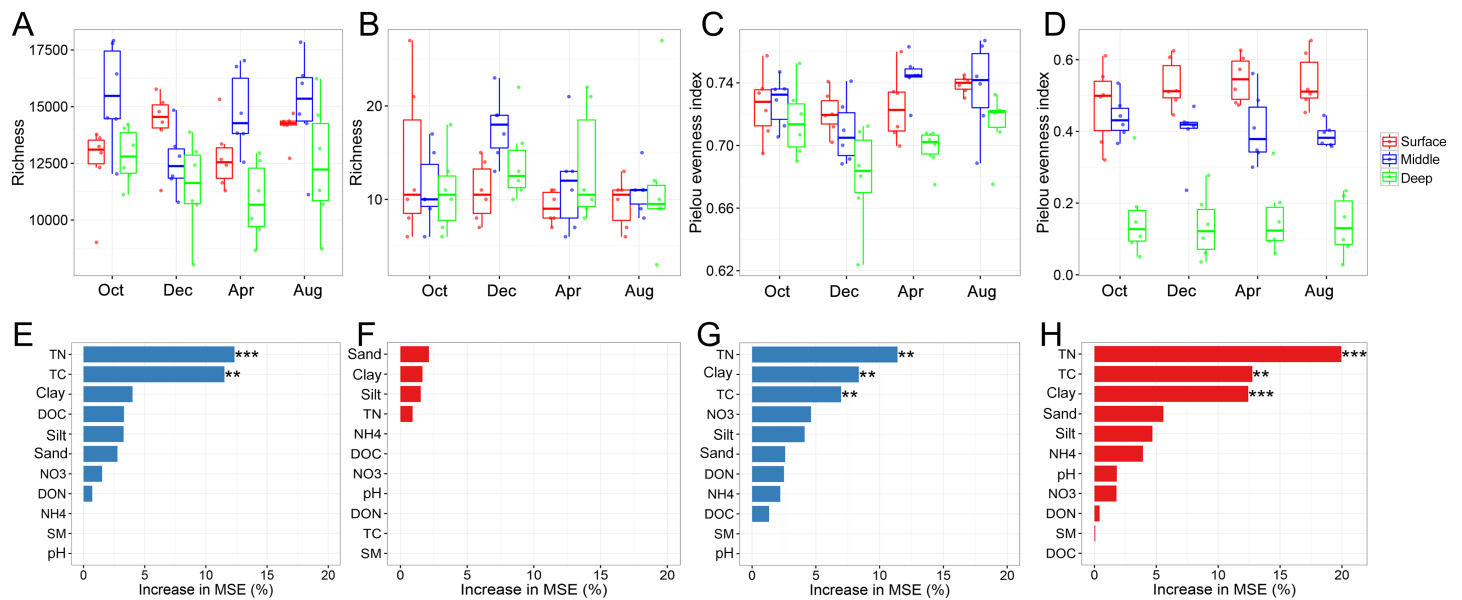


Figure 1 Bacterial and archaeal diversity indices. Bacterial richness (A), archaeal richness (B), bacterial Pielou evenness index (C), and archaeal evenness index (D) among soil layers in October, December, April, and October; random forest mean predictor importance of soil variables on bacterial richness (E), archaeal richness (F), bacterial Pielou evenness index (G), and archaeal Pielou evenness index (H). Significance level: $P < 0.05$, *; $P < 0.01$, **; $P < 0.001$, ***. Abbreviations: TC, total carbon; TN, total nitrogen; DOC, dissolved organic carbon; DON, dissolved organic nitrogen; NO₃, nitrate; NH₄, ammonium; SM, soil moisture.

Full-size [DOI: 10.7717/peerj.12868/fig-1](https://doi.org/10.7717/peerj.12868/fig-1)

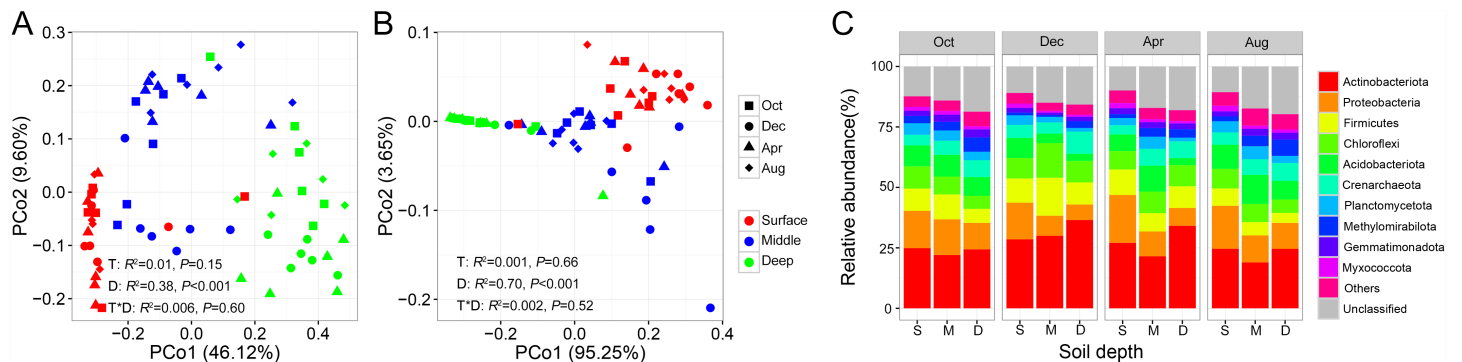


Figure 2 Principal coordinate analysis (PCoA) of soil bacterial community (A) and archaeal community (B); relative abundance of prokaryotic phyla shifts along soil profile (C). Abbreviations: S, surface soil layer; M, middle soil layer; D, deep soil layer.

Full-size [DOI: 10.7717/peerj.12868/fig-2](https://doi.org/10.7717/peerj.12868/fig-2)

that the bacterial and archaeal community compositions were separated by soil depth (Figs. 2A and 2B). We then tested for the temporal variations of the bacterial and archaeal community compositions at each soil depth. We observed that the archaeal soil community did not show any temporal variations at any soil depth (Fig. S1). In contrast, the bacterial soil community composition exhibited strong temporal variation in the middle soil layer (PERMANOVA: $r^2 = 0.13$, $P = 0.006$, Fig. S1).

At the phylum level, the bacterial community was mainly dominated by Actinobacteria, Proteobacteria, Firmicutes, Chloroflexi, and Acidobacteria, while the archaeal community was mainly dominated by Crenarchaeota in our study site, accounting for 69.19% of the total prokaryotic sequences (Fig. 2C). Next, we investigated the changes in relative abundance patterns of prokaryotic phyla associated with the soil profile. The relative abundance of nearly all phyla was significantly different among soil layers for all sampling times, except for Gemmatimonadetes and Chloroflexi. We observed that Proteobacteria, Firmicutes, Acidobacteria, Planctomycete, and Myxococcota, especially, decreased along with the soil profile, while Actinobacteria and Crenarchaeota exhibited the opposite trend (Fig. 2C, Table S2). Most strikingly, the proportion of unclassified ASVs showed an increasing trend along with the soil profile (Fig. 2C).

The shifts in bacterial and archaeal communities were also reflected in the abundant ASVs, with 50%–72.49% of total abundant ASVs unevenly distributed along the soil profile among sampling times (Figs. 3A–3D). ASVs enriched in the surface soil layer, mainly Solirubrobacterales and Propionibacterales, accounted for a more significant proportion than those in the middle and deep soil layers (Fig. 3E). ASVs enriched in the middle and deep soil layer, primarily classified as Rokubacterales and Gaiellales, differed from those in the surface soil layer (Fig. 3E). Meanwhile, abundant ASVs also exhibited temporal variations: there were less enriched ASVs (50%) in October than in other sampling times (71.5%–72.5%, Figs. 3A–3D). Interestingly, the proportion of unclassified ASVs was less abundant in the surface soil than in the middle and deep soil layers (Fig. 3E).

Predicted prokaryotic function

FAPROTAX was adopted to annotate and then screen the key ecological functions of the bacterial communities that contributed to soil biogeochemical cycling. A total of 48 functional groups were identified. Although a minor proportion of all ASVs (5.04%) was assigned to at least one predicted functional group, the predicted functional composition was significantly correlated with the soil prokaryotic community composition as confirmed by the Mantel test ($r = 0.64$, $P < 0.001$). Among these functional groups, we observed that 41 predicted functional groups exhibited significant differences among depths, while 21 functional groups displayed apparent temporal variations (Table S3).

Rare bacterial and archaeal ASVs

The proportion of rare bacterial richness and abundance was highest in the surface soil layer and showed a decreasing trend along with the soil profile (Figs. 4A and 4C). On the other hand, the proportion of rare archaeal richness and abundance were not impacted by soil depth (Figs. 4B and 4D). Random forest analysis further indicated that the rare bacterial community was the key determinant of the soil fertility index and functional diversity (Fig. S2). Moreover, the potential role of 43 archaeal and bacterial phyla on soil fertility was assessed. Among these phyla, Latescibacterota, Zixibacteria, Dadabacteria, Dependientiae, Firmicutes, Micrarchaeota, and Calditrichota were the main predictors of the soil fertility index (Fig. S3); these phyla, except Firmicutes, were relatively rare taxa, accounting for 0.00034%–0.31% of the total reads. We further explored the potential

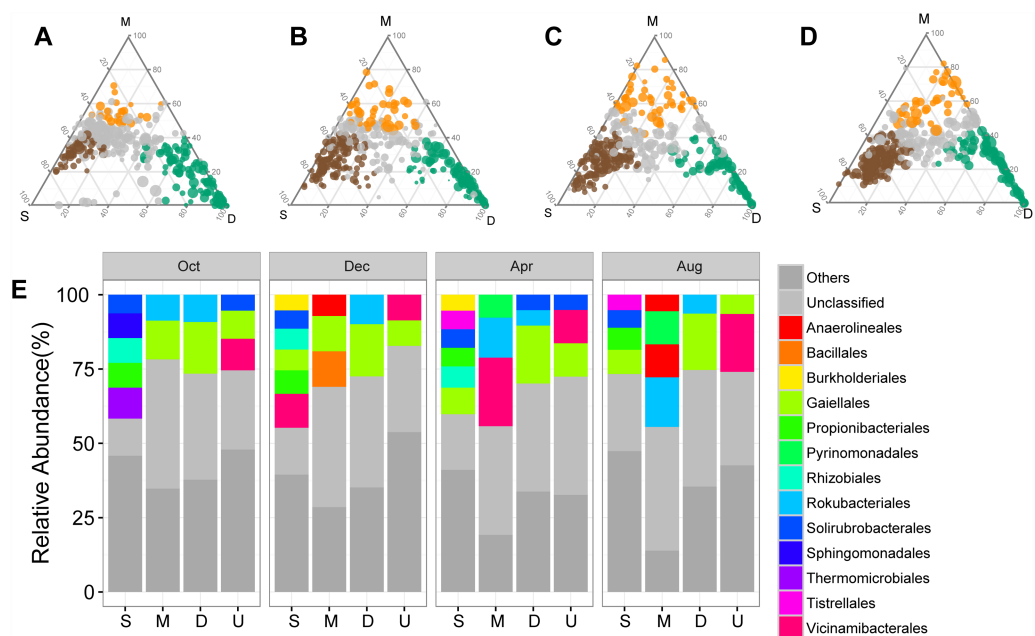


Figure 3 Ternary plots showing the distribution of enriched prokaryotic ASVs in surface (brown), middle (orange), and deep (green) soil layers in October (A), December (B), April (C), and August (D). The size of each circle is equivalent to its relative abundance. The orders of enriched ASVs are displayed in bar plots. Abbreviations: S, ASVs enriched in surface soil layer; M, ASVs enriched in middle soil layer; D, ASVs enriched in deep soil layer; U, unenriched ASVs.

Full-size [DOI: 10.7717/peerj.12868/fig-3](https://doi.org/10.7717/peerj.12868/fig-3)

functions of rare phyla through FAPROTAX: Desulfobacterota accounted for 90.65% of sulfate respiration and 46.26% of iron respiration; Cyanobacteria accounted for 6.10% of nitrogen fixation and 97.50% of phototrophy; Halanaerobiaeota accounted for 62.90% of cellulolysis; and Euryarchaeota and Halobacterota accounted for 14.33% and 0.12% of dark hydrogen oxidation, respectively.

Prokaryotic co-occurrence network

We constructed a soil prokaryotic co-occurrence network for each soil layer. The network size generally became larger along with increasing soil depth, with 214, 247, and 274 nodes in the surface, middle, and deep soil layers, respectively (Figs. 5A–5C). Additionally, we observed that only a small proportion of nodes (54) were shared among these three networks. The network of the surface soil layer was analogous to the network of the middle soil layer (153 nodes shared) but was quite dissimilar to that of the deep soil layer (65 nodes shared). The surface and middle soil layers' networks were more connected than that of the deep soil layer (Figs. 5A–5C). This pattern was affirmed by their topological characteristics (Table S4). Furthermore, the number of positive/negative links gradually decreased along with the soil profile (Table S4).

From the plot of Zi (a value measuring within-module connectivity) against Pi (a value measuring among-module connectivity), the different roles of each node in the network were identified (Figs. 5D–5F). Notably, there were more keystone species observed in the

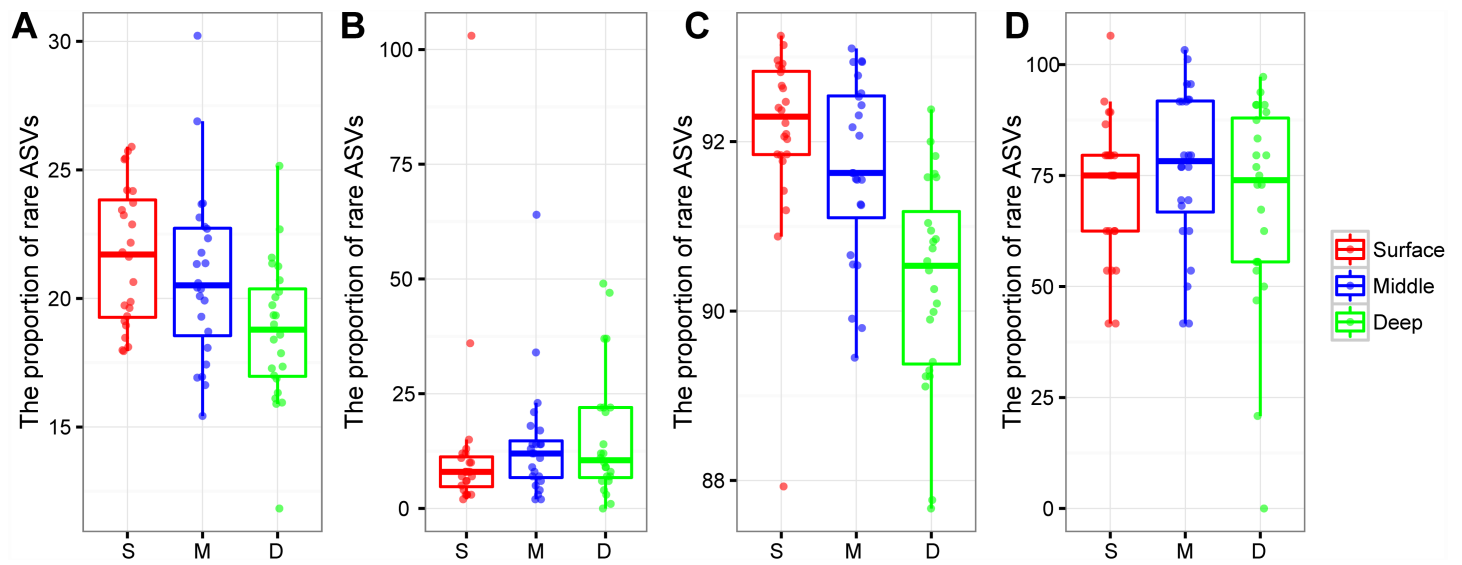


Figure 4 Box plots showing the proportion of rare bacterial richness (A) and rare archaeal richness (B), the proportion of rare bacterial abundance (C), and rare archaeal abundance (D) among soil layers. Abbreviations: S, surface soil layer; M, middle soil layer; D, deep soil layer.

Full-size [DOI: 10.7717/peerj.12868/fig-4](https://doi.org/10.7717/peerj.12868/fig-4)

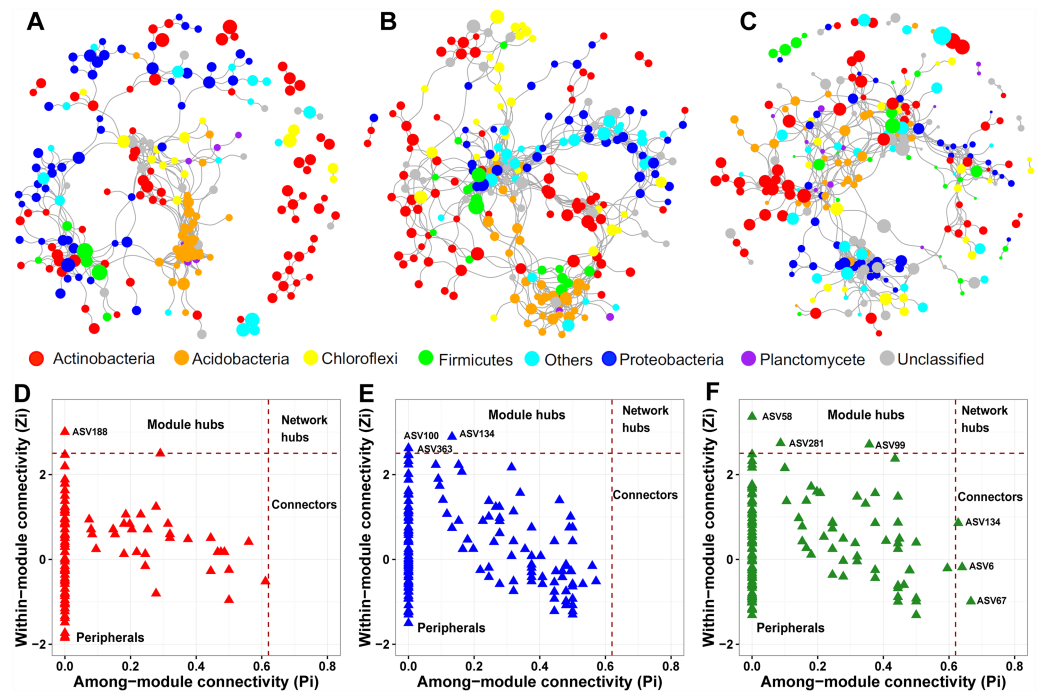


Figure 5 Prokaryotic co-occurrence networks in surface (A), middle (B), and deep (C) soil layers. The size of each node is proportional to its relative abundance. Pi-Zi plots showing the distribution of prokaryotic soil ASVs based on their topological roles in surface (D), middle (E), and deep (F) soil layers.

Full-size [DOI: 10.7717/peerj.12868/fig-5](https://doi.org/10.7717/peerj.12868/fig-5)

deep soil layer (six) than in the surface (one) and middle (three) soil layers (Figs. 5D–5F). These keystone species were not shared among the three networks. The classification of these keystone species is listed in Table S5.

DISCUSSION

Different vertical patterns between bacterial and archaeal diversities

We observed that the deep soil layer harbors the lowest bacterial diversity, which is consistent with previous studies from the alpine meadow (Xu *et al.*, 2021), soybean cropland (Hao *et al.*, 2020), and deciduous birch forest (Mundra *et al.*, 2021). This finding confirmed the ecological principle that more extreme environments are expected to be inhabited by a less diverse community of microbes (Gaston, 2000). In contrast to previous studies, we observed that bacterial richness and evenness in the middle soil layer were roughly equal to that in the surface. This observation may have the following two explanations. First, this is likely attributable to the soil carbon content which did not differ between the surface and middle soil layer. Soil organic carbon content is the most important factor that influences soil microorganisms in addition to pH (Fierer, 2017; Kukreti *et al.*, 2020) and was observed to be the best predictor of bacterial diversity, as indicated by random forest analysis in this study. Second, on a depth-weighted basis, the microbial biomass in the middle soil layer may not be less than that in the surface soil layer, thus, containing an equal diversity to the surface soil layer.

For archaea, its richness was unaffected by soil depth and cannot be explained by any measured soil variable. Archaea are reported to be highly adaptable to environmental stress (Cao *et al.*, 2012), so may survive in the nutrient-poor and oxygen-deficient deep soil layer. In addition, we detected that archaeal communities are highly uneven in deep soil, meaning that there is extreme dominance by one or a few archaeal taxa. Therefore, the archaeal communities in deep soil would be less resistant to environmental stress (Wittebolle *et al.*, 2009).

Vertical and temporal patterns of bacterial and archaeal communities

Both bacterial and archaeal community compositions exhibited distinct distribution patterns with the vertical soil profile, mainly due to the drastic differences in soil carbon and nitrogen content and soil texture (Jobbágy & Jackson, 2000; Eilers *et al.*, 2012; Gu *et al.*, 2017). We also detected different ecological preferences for certain phyla. For instance, Actinobacteria and Crenarchaeota showed increasing abundance along with the vertical soil profile. This result corresponds with the conclusions of previous studies that archaea occupy a larger proportion in the deep soil layer due to their anaerobic characteristic. However, some bacterial phyla, including Proteobacteria, Firmicutes, and Acidobacteria, displayed a sharply decreasing trend in the deep soil layer. This trend may be owing to the copiotrophic-oligotrophic trade-off theory (Ramirez, Craine & Fierer, 2012) and/or the deficiency in oxygen availability. For instance, Proteobacteria are typically copiotrophic and capable of degrading various organic materials (Eilers *et al.*, 2012). Thus, the higher abundances of Proteobacteria in the surface layer may, in part, reflect increased organic carbon availability.

We only detected a slight temporal variation in bacterial and archaeal community compositions. In a meta-analysis, *Shade et al. (2013)* found that soil microbial communities were consistently less temporally variable than other ecosystems. For time-series field studies, it is impossible to collect samples in the exact same location due to destructive soil sampling. Therefore, soil microbial community dynamics may be masked by the high small-scale spatial heterogeneity (*Docherty et al., 2015*). Although the use of the soybean in a crop rotation would increase soil total nitrogen, there is no apparent difference in soil TN and TC among seasons (*Table S6*) in the present study, perhaps due to the removal of crop residues after harvest. As proposed by previous studies, temporal variations of soil microbial communities are driven by soil physicochemical characteristics (*Yang et al., 2017*). Consequently, the minimal temporal variations might be explained, in part, by the unapparent difference in soil TC and TN. Alternatively, soil contains many dormant microbes (*Lennon & Jones, 2011*), which were indiscriminately captured using the 16s rRNA sequencing approach (*Shade et al., 2013*). For the reasons above, the soil prokaryotic communities only fluctuated within a narrow range over the given time period.

Vertical pattern of rare ASVs and their potential function

Deep sequencing enabled us to explore reliable rare taxa in soil. For the first time, the vertical patterns of rare bacterial and archaeal taxa were evaluated. We observed that the proportion of rare bacterial ASVs decreased along with the soil profile. The root system in the surface soil layer was more abundant than that in the middle or deep soil layers (*Chaparro, Badri & Vivanco, 2014*) and may provide more niches for rare species with specialized functions. Alternatively, it was proposed that rare species may occur from stochastic processes (*Jousset et al., 2017*), whereas deterministic processes would reduce rare species. Since soil nutrient availability generally decreased along with the vertical soil profile, stochastic processes would play a less predominant role in the deep soil layer.

Microbial rare taxa constitute an important “genomic reservoir” and could be used for a variety of functions (*Jousset et al., 2017*). It was evidenced that rare taxa contributed to sulfate reduction (*Pester et al., 2012*) and have the potential for bioremediation (*Pascoal, Magalhães & Costa, 2020*). Recently, the ecological importance of rare taxa in terrestrial and aquatic ecosystems has been increasingly recognized (*Jousset et al., 2017; Chen et al., 2020*). In the present study, the community of rare taxa was observed to be a key determinant of soil fertility and ecosystem multifunctionality, indicating that rare taxa contribute greatly to agroecosystem functioning. Likewise, *Chen et al. (2020)* also observed that rare bacterial and fungal taxa (less than 3% of total reads) were the major drivers of ecosystem multifunctionality in long-term fertilized soils. As revealed by Random Forest analysis, rare prokaryotic phyla (*e.g.*, Latescibacterota, Zixibacteria, and Dadabacteria) had an over-proportional role in the soil biogeochemical cycling process. In contrast, the dominant bacterial and archaeal phyla (*e.g.*, Proteobacteria, Actinobacteria, and Acidobacteria) showed little control over soil fertility. Our analysis further revealed that rare phyla influence narrow-range functions, such as dark hydrogen oxidation, sulfate respiration, and cellulolysis. Other studies highlighted that rare taxa exert great influence over sulfate reduction (*Pester et al., 2010; Dell’Anno et al., 2012*) and organic pollutant

degradations, where their removal greatly impacted ecosystem function (Jousset *et al.*, 2017). Taken together, our knowledge of the ecological role of the rare microbial taxa is still obscure with more attention needed.

Vertical patterns of prokaryotic co-occurrence networks

Our results emphasized that the prokaryotic co-occurrence network was more complex in the surface and middle soil layers than in the deep soil layer, reflected by the higher connectedness and average degree. Likewise, Xu *et al.* (2021) observed a more complex and more extensive microbial network in the surface than in the deep soil layer of meadows and shrubland. Higher network complexity in the surface and middle soil layer might be explained, in part, by greater amounts of substrate and nutrient in the surface soil layer (Creamer *et al.*, 2016), as microbial interactions feeding on these substrates would be largely strengthened. In fact, a large number of studies have found that soil microbial networks would be more connected after nutrient amendments (Yang *et al.*, 2019; Yang *et al.*, 2020). Moreover, the microbial richness and biomass were extremely low in the deep soil layer (Eilers *et al.*, 2012; Xu *et al.*, 2021), reducing the opportunity for different species to interact. It has been proposed that complex networks contribute to better environmental adaptation and higher efficiency of resources transfer than simple networks (Morriën *et al.*, 2017). In this sense, the complex prokaryotic networks in the surface and middle soil layers would lead to better biogeochemical functions and provide a stable soil environment for plant growth.

We observed a gradual increase in proportion of negative links along the soil profile by assessing the correlations among microbial species, which indicates that microbial competition for the same resources increases with soil depth. The niche differentiation would be a possible explanation for this result. The surface and middle soil layers are easily impacted by external disturbance; the roots within these layers also create heterogeneous environments (Morriën *et al.*, 2017), leading to significant niche differentiation and reduction of competition. In contrast, the deep soil layer is deficient in roots and is less impacted by external disturbance, possessing a more homogeneous environment and weak niche differentiation (Faust & Raes, 2012). The nutrient-limited environment in the deep soil layer would also induce intense competition among species.

In conclusion, we fully evaluated the vertical and temporal patterns of soil bacterial and archaeal communities and co-occurrence networks in a wheat-soybean rotation agroecosystem. Our results showed that soil depth as a more predominant factor than sampling time in shaping soil prokaryotic communities and co-occurrence networks and revealed different vertical distribution patterns of bacteria and archaea. With deep sequencing at a depth of millions of sequences per sample, we can investigate the rare microbial taxa and reveal their potential in soil biogeochemical function. Overall, our better understanding of vertical and temporal patterns of soil prokaryotic communities should enable more strategic agricultural management to mitigate adverse environmental impacts and improve crop productivity.

ADDITIONAL INFORMATION AND DECLARATIONS

Funding

This research was supported by the National Natural Science Foundation of China (32071547). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Grant Disclosures

The following grant information was disclosed by the authors:
National Natural Science Foundation of China: 32071547.

Competing Interests

The authors declare there are no competing interests.

Author Contributions

- Mika Yokota conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Yupeng Guan performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Yi Fan performed the experiments, analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.
- Ximei Zhang conceived and designed the experiments, authored or reviewed drafts of the paper, and approved the final draft.
- Wei Yang conceived and designed the experiments, prepared figures and/or tables, and approved the final draft.

Data Availability

The following information was supplied regarding data availability:
The raw sequence data is available at NCBI SRA: [PRJNA766099](https://www.ncbi.nlm.nih.gov/sra/PRJNA766099).

Supplemental Information

Supplemental information for this article can be found online at <http://dx.doi.org/10.7717/peerj.12868#supplemental-information>.

REFERENCES

- Amir A, McDonald D, Navas-Molina JA, Kopylova E, Morton JT, Xu ZZ, Kightley EP, Thompson LR, Hyde ER, Gonzalez A, Knight R, Gilbert JA. 2017. Deblur rapidly resolves single-nucleotide community sequence patterns. *mSystems* 2:e00191-00116.
- Archer E. 2016. RfPermute: estimate permutation p -values for random forest importance metrics. Available at <https://CRAN.R-project.org/package=rfpermute>.
- Auguet JC, Barberan A, Casamayor EO. 2010. Global ecological patterns in uncultured Archaea. *The ISME Journal* 4:182–190 DOI 10.1038/ismej.2009.109.

- Bao SD. 2000.** *Soil and agricultural chemistry analysis*. Beijing: Agricultural Science and Technology Press (in Chinese).
- Bates ST, Berg-Lyons D, Caporaso JG, Walters WA, Knight R, Fierer N. 2011.** Examining the global distribution of dominant archaeal populations in soil. *The ISME Journal* 5:908–917 DOI [10.1038/ismej.2010.171](https://doi.org/10.1038/ismej.2010.171).
- Benjamini Y, Hochberg Y. 1995.** Controlling the false discovery rate: a practical and powerful approach to multiple testing. *Journal of the Royal Statistical Society Series B* 57:289–300.
- Bolyen E, Rideout JR, Dillon MR, Bokulich NA, Abnet C, Al-Ghalith GA, Alexander H, Alm EJ, Arumugam M, Asnicar F, Bai Y, Bisanz JE, Bittinger K, Brejnrod A, Brislawn CJ, Brown CT, Callahan BJ, Caraballo-Rodríguez AM, Chase J, Cope E, Da Silva R, Dorrestein PC, Douglas GM, Durall DM, Duvallet C, Edwardson CF, Ernst M, Estaki M, Fouquier J, Gauglitz JM, Gibson DL, Gonzalez A, Gorlick K, Guo J, Hillmann B, Holmes S, Holste H, Huttenhower C, Huttley G, Janssen S, Jarmusch AK, Jiang L, Kaehler B, Kang KB, Keefe CR, Keim P, Kelley ST, Knights D, Koester I, Kosciulek T, Kreps J, Langille MG, Lee J, Ley R, Liu Y, Loftfield E, Lozupone C, Maher M, Marotz C, Martin BD, McDonald D, McIver LJ, Melnik AV, Metcalf JL, Morgan SC, Morton J, Naimey AT, Navas-Molina JA, Nothias LF, Orchanian SB, Pearson T, Peoples SL, Petras D, Preuss ML, Pruesse E, Rasmussen LB, Rivers A, Robeson II MS, Rosenthal P, Segata N, Shaffer M, Shiffer A, Sinha R, Song SJ, Spear JR, Swafford AD, Thompson LR, Torres PJ, Trinh P, Tripathi A, Turnbaugh PJ, Ul-Hasan S, Vander Hooft JJ, Vargas F, Vázquez-Baeza Y, Vogtmann E, Von Hippel M, Walters W, Wan Y, Wang M, Warren J, Weber KC, Williamson CH, Willis AD, Xu ZZ, Zaneveld JR, Zhang Y, Zhu Q, Knight R, Caporaso JG. 2018.** QIIME 2: Reproducible, interactive, scalable, and extensible microbiome data science (No. e27295v1). *PeerJ Preprints* 6:e27295v2 DOI [10.1038/s41587-019-0209-9](https://doi.org/10.1038/s41587-019-0209-9).
- Breiman L. 2001.** Random forest. *Machine Learning* 45:5–32 DOI [10.1023/A:1010933404324](https://doi.org/10.1023/A:1010933404324).
- Cao P, Zhang LM, Shen JP, Zheng YM, Di HJ, He JZ. 2012.** Distribution and diversity of archaeal communities in selected Chinese soils. *FEMS Microbiology Ecology* 80:146–158 DOI [10.1111/j.1574-6941.2011.01280.x](https://doi.org/10.1111/j.1574-6941.2011.01280.x).
- Caporaso JG, Lauber CL, Walters WA, Berg-Lyons D, Huntley J, Fierer N, Owens SM, Betley J, Fraser L, Bauer M, Gormley N, Gilbert JA, Smith G, Knight R. 2012.** Ultra-high-throughput microbial community analysis on the Illumina HiSeq and MiSeq platforms. *The ISME Journal* 6:1621–1624 DOI [10.1038/ismej.2012.8](https://doi.org/10.1038/ismej.2012.8).
- Caporaso JG, Lauber CL, Walters WA, Berg-Lyons D, Lozupone CA, Turnbaugh PJ, Fierer N, Knight R. 2011.** Global patterns of 16S rRNA diversity at a depth of millions of sequences per sample. *Proceedings of National Academy Sciences of the United States of America* 108:4516–4522 DOI [10.1073/pnas.1000080107](https://doi.org/10.1073/pnas.1000080107).
- Chaparro JM, Badri D, Vivanco J. 2014.** Rhizosphere microbiome assemblage is affected by plant development. *The ISME Journal* 8:790–803 DOI [10.1038/ismej.2013.196](https://doi.org/10.1038/ismej.2013.196).

- Chaudhary P, Khati P, Gangola S, Kumar A, Kumar R, Sharma A. 2021.** Impact of nanochitosan and *Bacillus* spp. on health, productivity and defence response in *Zea mays* under field condition. *3Biotech* **11**:237.
- Chen QL, Ding J, Zhu D, Hu HW, Delgado-Baquerizo M, Ma YB, He JZ, Zhu YG. 2020.** Rare microbial taxa as the major drivers of ecosystem multifunctionality in long-term fertilized soils. *Soil Biology and Biochemistry* **141**:107686 DOI [10.1016/j.soilbio.2019.107686](https://doi.org/10.1016/j.soilbio.2019.107686).
- Creamer RE, Hannula SE, Leeuwen JP, Stone D, Rutgers M, Schmelz RM, Ruiter PCD, Hendriksen NB, Bolger T, Bouffaud ML, Buee M, Carvalho F, Costa D, Dirilgen T, Francisco R, Griffiths BS, Griffiths R, Martin F, Silva PMD, Mendes S, Morais PV, Pereira C, Philippot L, Plassart P, Redecker D, Römbke J, Sousa JP, Wouterse M, Lemanceau P. 2016.** Ecological network analysis reveals the interconnection between soil biodiversity and ecosystem function as affected by land use across Europe. *Applied Soil Ecology* **97**:112–124 DOI [10.1016/j.apsoil.2015.08.006](https://doi.org/10.1016/j.apsoil.2015.08.006).
- Dell'Anno A, Beolchini F, Rocchetti L, Luna GM, Danavaro R. 2012.** High bacterial biodiversity increases degradation performance of hydrocarbons during bioremediation of contaminated harbor marine sediments. *Environmental Pollution* **167**:85–92 DOI [10.1016/j.envpol.2012.03.043](https://doi.org/10.1016/j.envpol.2012.03.043).
- Docherty KM, Borton HM, Espinosa N, Gebhardt M, Gil-Loaiza J, Gutknecht JL, Maes PW, Mott BM, Parnell JJ, Purdy G, Rodrigues PA, Stanish LF, Walser ON, Gallery RE. 2015.** Key edaphic properties largely explain temporal and geographic variation in soil microbial communities across four biomes. *PLOS ONE* **10**:e0135352 DOI [10.1371/journal.pone.0135352](https://doi.org/10.1371/journal.pone.0135352).
- Eilers KG, Debenport S, Anderson S, Fierer N. 2012.** Digging deeper to find unique microbial communities: the strong effect of depth on the structure of bacterial and archaeal communities in soil. *Soil Biology and Biochemistry* **50**:58–65 DOI [10.1016/j.soilbio.2012.03.011](https://doi.org/10.1016/j.soilbio.2012.03.011).
- Faust K, Raes J. 2012.** Microbial interactions: from networks to models. *Nature Reviews Microbiology* **10**:538–550 DOI [10.1038/nrmicro2832](https://doi.org/10.1038/nrmicro2832).
- Fierer N. 2017.** Embracing the unknown: disentangling the complexities of the soil microbiome. *Nature Reviews Microbiology* **15**:579–590 DOI [10.1038/nrmicro.2017.87](https://doi.org/10.1038/nrmicro.2017.87).
- Gaston KJ. 2000.** Global patterns in biodiversity. *Nature* **405**:220–227 DOI [10.1038/35012228](https://doi.org/10.1038/35012228).
- Gu Y, Wang Y, Lu S, Xiang Q, Yu X, Zhao K, Zou L, Chen Q, Tu S, Zhang X. 2017.** Long-term fertilization structures bacterial and archaeal communities along soil depth gradient in a paddy soil. *Frontiers in Microbiology* **8**:1516 DOI [10.3389/fmicb.2017.01516](https://doi.org/10.3389/fmicb.2017.01516).
- Guimera R, Amaral L. 2005.** Functional cartography of complex metabolic networks. *Nature* **433**:895–900 DOI [10.1038/nature03288](https://doi.org/10.1038/nature03288).
- Hao J, Chai YN, Lopes LD, Ordóñez RA, Wright EE, Archontoulis S, Schachtman DP. 2020.** The effects of soil depth on the structure of microbial communities in agricultural soils in Iowa, USA. *Applied and Environmental Microbiology* **87**:e02673-20.

- Jiao S, Chen W, Wang J, Du N, Li Q, Wei G. 2018.** Soil microbiomes with distinct assemblies through vertical soil profiles drive the cycling of multiple nutrients in reforested ecosystems. *Microbiome* **6**:146 DOI [10.1186/s40168-018-0526-0](https://doi.org/10.1186/s40168-018-0526-0).
- Jiao S, Lu Y. 2020.** Abundant fungi adapt to broader environmental gradients than rare fungi in agricultural fields. *Global Change Biology* **26**:4506–4520 DOI [10.1111/gcb.15130](https://doi.org/10.1111/gcb.15130).
- Jobbágy EG, Jackson RB. 2000.** The vertical distribution of soil organic carbon and its relation to climate and vegetation. *Ecological Applications* **10**:423–436 DOI [10.1890/1051-0761\(2000\)010\[0423:TVDOSO\]2.0.CO;2](https://doi.org/10.1890/1051-0761(2000)010[0423:TVDOSO]2.0.CO;2).
- Jousset A, Bienhold C, Chatzinotas A, Gallien L, Gobet A, Kurm V, Küsel K, Rillig MC, Rivett DW, Salles JF, Van der Heijden MGA, Youssef NH, Zhang X, Wei Z, Hol WHG. 2017.** Where less may be more: how the rare biosphere pulls ecosystems strings. *The ISME Journal* **11**:853–862 DOI [10.1038/ismej.2016.174](https://doi.org/10.1038/ismej.2016.174).
- Kim HM, Lee MJ, Jung JY, Hwang CY, Kim M, Ro HM, Chun J, Lee YK. 2016.** Vertical distribution of bacterial community is associated with the degree of soil organic matter decomposition in the active layer of moist acidic tundra. *Journal of Microbiology* **54**:713–723 DOI [10.1007/s12275-016-6294-2](https://doi.org/10.1007/s12275-016-6294-2).
- Kivlin SN, Hawkes CV. 2020.** Spatial and temporal turnover of soil microbial communities is not linked to function in a primary tropical forest. *Ecology* **101**:e02985.
- Kukreti B, Sharma A, Chaudhary P, Agri U, Maithani D. 2020.** Influence of nanosilicon dioxide along with bioinoculants on *Zea mays* and its rhizospheric soil. *3Biotech* **10**:345.
- Lennon JT, Jones SE. 2011.** Microbial seed banks: the ecological and evolutionary implications of dormancy. *Nature Reviews Microbiology* **9**:119–130 DOI [10.1038/nrmicro2504](https://doi.org/10.1038/nrmicro2504).
- Liang Y, Xiao X, Nuccio EE, Yuan M, Zhang N, Xue K, Cohan FM, Zhou J, Sun B. 2020.** Differentiation strategies of soil rare and abundant microbial taxa in response to changing climatic regimes. *Environmental Microbiology* **22**:1327–1340 DOI [10.1111/1462-2920.14945](https://doi.org/10.1111/1462-2920.14945).
- Liaw A, Wiener M. 2002.** Classification and regression by randomForest. *R News* **2**:18–22.
- Louca S, Parfrey LW, Doebeli M. 2016.** Decoupling function and taxonomy in the global ocean microbiome. *Science* **353**:1272–1277.
- Morriën E, Hannula SE, Snoek LB, Helmsing NR, Zweers H, De Hollander M, Soto RL, Bouffaud L, Buée M, Dimmers W, Duyts H, Geisen S, Girlanda M, Griffiths RI, Jørgensen HB, Jensen J, Plassart P, Redecker D, Schmelz RM, Schmidt O, Thomson BC, Tisserant E, Uroz S, Winding A, Bailey MJ, Bonkowski M, Faber JH, Martin F, Lemanceau P, DeBoer W, VanVeen JA, Van der Putten WH. 2017.** Soil networks become more connected and take up more carbon as nature restoration progresses. *Nature Communications* **8**:14349 DOI [10.1038/ncomms14349](https://doi.org/10.1038/ncomms14349).
- Mundra S, Kjønaas OJ, Morgado LN, Krabberød AK, Ransedokken Y, Kauserud H. 2021.** Soil depth matters: shift in composition and inter-kingdom co-occurrence

- patterns of microorganisms in forest soils. *FEMS Microbiology Ecology* **97**:fiab022 DOI [10.1093/femsec/fiab022](https://doi.org/10.1093/femsec/fiab022).
- Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara RB, Simpson GL, Solymos P. 2013.** Vegan: community ecology. Package Version 2.0-6. Available at <http://CRAN.R-project.org/package=vegan>.
- Pascoal F, Magalhães C, Costa R. 2020.** The link between the ecology of the prokaryotic rare biosphere and its biotechnological potential. *Frontiers in Microbiology* **11**:231 DOI [10.3389/fmicb.2020.00231](https://doi.org/10.3389/fmicb.2020.00231).
- Pester M, Bittner N, Deevong P, Wagner M, Loy A. 2010.** A 'rare biosphere' microorganism contributes to sulfate reduction in a peatland. *The ISME Journal* **4**:1591–1602 DOI [10.1038/ismej.2010.75](https://doi.org/10.1038/ismej.2010.75).
- Pester M, Knorr K-H, Friedrich MW, Wagner M, Loy A. 2012.** Sulfate-reducing microorganisms in wetlands—fameless actors in carbon cycling and climate change. *Frontiers in Microbiology* **3**:72 DOI [10.3389/fmicb.2012.00072](https://doi.org/10.3389/fmicb.2012.00072).
- Ramirez KS, Craine JM, Fierer N. 2012.** Consistent effects of nitrogen amendments on soil microbial communities and processes across biomes. *Global Change Biology* **18**:1918–1927 DOI [10.1111/j.1365-2486.2012.02639.x](https://doi.org/10.1111/j.1365-2486.2012.02639.x).
- Revelle W. 2015.** Psych: procedures for psychological, psychometric, and personality research. Package Version 1.5.1. Available at <https://CRAN.R-project.org/package=psych>.
- Shade A, Caporaso JG, Handelsman J, Knight R, Fierer N. 2013.** A meta-analysis of changes in bacterial and archaeal communities with time. *The ISME Journal* **7**:1493–1506 DOI [10.1038/ismej.2013.54](https://doi.org/10.1038/ismej.2013.54).
- Wang H, Bu L, Tian J, Hu Y, Song F, Chen C, Zhang Y, Wei G. 2021.** Particular microbial clades rather than total microbial diversity best predict the vertical profile variation in soil multifunctionality in desert ecosystems. *Land Degradation and Development* **32**:2157–2168 DOI [10.1002/ldr.3873](https://doi.org/10.1002/ldr.3873).
- Wardle DA. 1998.** Controls of temporal variability of the soil microbial biomass: a global-scale synthesis. *Soil Biology and Biochemistry* **30**:1627–1637 DOI [10.1016/S0038-0717\(97\)00201-0](https://doi.org/10.1016/S0038-0717(97)00201-0).
- Wei G, Li M, Shi W, Tian R, Chang C, Z. Wang Z, Wang N, Zhao G, Gao Z. 2020.** Similar drivers but different effects lead to distinct ecological patterns of soil bacterial and archaeal communities. *Soil Biology and Biochemistry* **144**:107759 DOI [10.1016/j.soilbio.2020.107759](https://doi.org/10.1016/j.soilbio.2020.107759).
- Widder S, Besemer K, Singer GA, Ceola S, Bertuzzo E, Quince C, Sloan TW, Rinaldo A, Battin TJ. 2014.** Fluvial network organization imprints on microbial co-occurrence networks. *Proceedings of National Academy Sciences of the United States of America* **111**:12799–12804 DOI [10.1073/pnas.1411723111](https://doi.org/10.1073/pnas.1411723111).
- Wittebolle L, Marzorati M, Clement L, Balloi A, Daffonchio D, Heylen K, De Vos P, Verstraete W, Boon N. 2009.** Initial community evenness favours functionality under selective stress. *Nature* **458**:623–626 DOI [10.1038/nature07840](https://doi.org/10.1038/nature07840).

- Xu T, Chen X, Hou Y, Zhu B. 2021.** Changes in microbial biomass, community composition and diversity, and functioning with soil depth in two alpine ecosystems on the Tibetan plateau. *Plant and Soil* **459**:137–153 DOI [10.1007/s11104-020-04712-z](https://doi.org/10.1007/s11104-020-04712-z).
- Yang W, Guo Y, Wang X, Chen C, Hu Y, Cheng L, Gu S, Xu X. 2017.** Temporal variations of soil microbial community under compost addition in black soil of Northeast China. *Applied Soil Ecology* **121**:214–222 DOI [10.1016/j.apsoil.2017.10.005](https://doi.org/10.1016/j.apsoil.2017.10.005).
- Yang W, Jing X, Guan Y, Zhai C, Wang T, Shi D, Sun W, Gu S. 2019.** Response of fungal communities and co-occurrence network patterns to compost amendment in black soil of Northeast China. *Frontiers in Microbiology* **10**:1562 DOI [10.3389/fmicb.2019.01562](https://doi.org/10.3389/fmicb.2019.01562).
- Yang W, Yang Z, Guan Y, Zhai C, Shi D, Chen J, Wang T, Gu S. 2020.** Dose-dependent effect of compost amendment on soil bacterial community composition and co-occurrence network patterns in soybean agroecosystem. *Archives of Agronomy and Soil Science* **66**:1027–1041 DOI [10.1080/03650340.2019.1651450](https://doi.org/10.1080/03650340.2019.1651450).
- Yuan CL, Zhang LM, Wang JT, Teng WK, Hu HW, Shen JP, He JZ. 2020.** Limited effects of depth (0–80 cm) on communities of archaea, bacteria and fungi in paddy soil profiles. *European Journal of Soil Science* **71**:955–966.