

Toxicity of clothianidin to common Eastern North American fireflies

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Background. Previous research suggests that fireflies (Coleoptera: Lampyridae) are susceptible to commonly used insecticides. In the United States, there has been a rapid and widespread adoption of neonicotinoid insecticides, predominantly used as seed coatings on large-acreage crops like corn, soy, and cotton. Neonicotinoid insecticides are persistent in soil yet mobile in water, so they have potential to contaminate firefly habitats both in and adjacent to application sites. As a result, fireflies may be at high risk of exposure to neonicotinoids, possibly jeopardizing this already at-risk group of charismatic insects.

Methods. To assess the sensitivity of fireflies to neonicotinoids, we exposed larvae of *Photuris versicolor* complex and *Photinus pyralis* to multiple levels of clothianidin-treated soil and monitored feeding behavior, protective soil chamber formation, intoxication, and mortality.

Results. *Pt. versicolor* and *Pn. pyralis* larvae exhibited long-term intoxication and mortality at concentrations above 1000 ng g⁻¹ soil (1 ppm). Under sub-lethal clothianidin exposure, firefly larvae fed less and spent less time in protective soil chambers, two behavioral changes that could decrease larval survival in the wild.

Discussion. Both firefly species demonstrated sub-lethal responses in the lab to clothianidin exposure at field-realistic concentrations, although *Pt. versicolor* and *Pn. pyralis* appeared to tolerate higher clothianidin exposure relative to other soil invertebrates and beetle species. While these two firefly species, which are relatively widespread in North America, appear somewhat tolerant of neonicotinoid exposure in a laboratory setting, further work is needed to extend this conclusion to wild populations, especially in rare or declining taxa.

1 **Toxicity of clothianidin to common Eastern North**
2 **American fireflies**

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13 **Abstract**

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33 conclusion to wild populations, especially in rare or declining taxa.

34

35 **Introduction**

36 In the United States alone, insects are estimated to provide over \$50 billion in ecological
37 services (Losey and Vaughan, 2006). Fireflies have great popular appeal and aesthetic and
38 cultural value, but fireflies also contribute biological control of some pest species, including
39 slugs and snails, which can be important agricultural pests (Godan, 1983; Lewis, 2016). Human
40 activities, however, have put these services at risk by triggering global insect declines (Sánchez-
41 Bayo and Wyckhuys, 2019). Some charismatic groups such as fireflies (Coleoptera: Lampyridae)
42 may be at elevated risk of at least localized extinction due to ongoing human activities such as
43 heavy pesticide use in and around their habitats (Reed et al., 2020).

44 Despite broad agreement that pesticides can pose a serious extinction threat to fireflies
45 (Lewis et al., 2020), there is a very poor understanding of the direct toxicity of insecticides on
46 fireflies. The most commonly applied classes of insecticides (neonicotinoids, pyrethroids, and
47 organophosphates) are broadly neurotoxic to most insect taxa (Sparks, 2013), so fireflies are
48 unlikely to be an exception. Indeed, full-strength organophosphate and neonicotinoid
49 formulations are toxic to the aquatic firefly larvae *Luciola cruciata* and *Luciola lateralis*,
50 respectively (Tabaru et al., 1970; Lee et al., 2008). Unfortunately, there have been no studies
51 assessing how terrestrial firefly larvae respond to residual concentrations of these insecticides in
52 soil, a likely route of exposure. Larvae of many common firefly species in North America are
53 soil-dwellers that intimately interact with soil as they forage for prey and form protective molting
54 chambers (Buschman, 1984; Lewis, 2016). These larvae inhabit forested, suburban, and
55 agricultural soils, where neonicotinoid insecticides are often applied directly, or via coatings on
56 crop seeds, to protect against pests (Knoepp et al., 2012; Douglas and Tooker, 2015; Simon-
57 Delso et al., 2015). In these habitats, neonicotinoid concentrations in soil can range from less
58 than 5 ng g⁻¹ to over 4,000 ng g⁻¹ (Knoepp et al., 2012; Schaafsma et al., 2015; Pearsons et al.,

59 2021), concentrations that could plausibly influence behavior and survival of firefly larvae (Lee
60 et al., 2008). Some indirect evidence suggests that firefly larvae are susceptible to neonicotinoids
61 because adult lampyrid densities have been found to be lower where neonicotinoid-coated seeds
62 were planted (Disque et al., 2019); however, to our knowledge, there have been no direct
63 evaluations of how terrestrial firefly larvae respond to neonicotinoid-treated soil.

64 To assess the direct sensitivity of fireflies to neonicotinoid insecticides, we measured
65 feeding behavior, development, and survival of larvae of two common North American firefly
66 species – *Photuris versicolor* species complex and *Photinus pyralis* (Linnaeus 1767) – exposed
67 to clothianidin-treated soil. We focused on clothianidin, one of the most widely-used seed- and
68 soil-applied neonicotinoid and the primary metabolite of another commonly applied
69 neonicotinoid, thiamethoxam (Douglas and Tooker, 2015). Generally applied to combat sucking
70 and chewing insects, clothianidin disrupts insect central nervous systems, leading to paralysis
71 and death (Simon-Delso et al., 2015). We exposed larvae to multiple field-realistic levels of
72 clothianidin-treated soil for 30 to 100 days with the expectation that they would be sensitive to
73 clothianidin at concentrations that have been detected in firefly habitats.

74

75 **Materials & Methods**

76 **Chemicals**

77 We acquired clothianidin from Chem Service (West Chester, PA, USA; purity $\geq 98\%$),
78 and prepared stock solutions of 2×10^2 , 2×10^3 , 2×10^4 , and 2×10^5 ng mL⁻¹ clothianidin in
79 acetone (Sigma Aldrich, St. Louis, MO, USA, ACS reagent, purity $\geq 99.5\%$). Pure acetone
80 served as a control. We stored stock solutions at 4 °C and allowed them to reach room
81 temperature (20 °C) before applying them to soils for the assays.

82

83 **Firefly Collection and Colony Care**

84 We ran toxicity assays on three separate cohorts of fireflies (Table 1): late-instar larvae from the
85 *Photuris versicolor* species complex (*Pt. Photuris*), early-instar *Pt. versicolor*, and early-instar
86 *Photinus pyralis* (*Pn. pyralis*). Both *Pt. versicolor* and *Pn. pyralis* are relatively large-bodied (6-
87 20 mm adult body length), widespread firefly species found throughout Eastern North America,
88 and their populations do not appear to be declining (Lewis, 2016). Because both species spend 1-
89 2 years in the soil as larvae and feed on soil invertebrates (*Pt. versicolor* are thought to feed on a
90 diversity of soil invertebrates whereas *Pn. pyralis* larvae are considered specialists on
91 earthworms; McLean et al., 1972; Buschman, 1984; Lewis, 2016), they likely experience chronic
92 contact and oral neonicotinoid exposure in contaminated habitats.

93 Five of the late-instar *Pt. versicolor* were reared from eggs laid by a mated female
94 collected in late July 2019 from the Bucknell University Chillisquaque Creek Natural Area
95 (Montour Co, PA; 41° 01' 15" N, 76° 44' 53" W), while the other 25 late-instar *Pt. versicolor*
96 were wild-collected in summer of 2019 from multiple locations throughout Pennsylvania: Bald
97 Eagle State Park (5 August; Centre Co, 41°00'44.0"N 77°12'54.3"W), Allegheny National Forest
98 (24-25 June; Forest Co, 41°31'29.8"N 79°17'33.9"W), and Bucknell University Forrest D. Brown
99 Conference Center (23-24 July; Union Co, PA; 40° 57' 28" N, 77° 00' 49" W). Larvae were
100 wild-collected at night by visually inspecting the ground for their faint glows. Larvae were
101 identified to genus by external morphology (McLean et al., 1972). After collection, we housed
102 individual larvae in 16-oz clear plastic deli containers (11.5-cm diameter × 8-cm tall) lined with
103 moist filter paper. Every 1-2 weeks, we provided each larva with one piece of cat food (Grain-
104 Free Real Chicken Recipe Dry Cat Food, Whole Earth Farm™, Merrick Pet Care Inc., Amarillo,

105 TX, USA), which had been softened in DI-water for 1 h (McLean et al., 1972). After 24 h, we
106 removed cat food and replaced the filter paper. Occasionally there was extensive fungal growth
107 on the cat food, which could be fatal to *Pt. versicolor* larvae; in these instances, we gently wiped
108 larvae with DI water and a delicate task wipe then transferred them to clean containers.

109 Early-instar *Pt. versicolor* and *Pn. pyralis* cohorts were reared from eggs laid in July
110 2020. On the evening of 10 July 2020, we collected 3 male and 2 female *Pt. versicolor* adults
111 and 3 mated *Pt. versicolor* females. Flying *Pn. pyralis* males were collected and identified based
112 on their characteristic “J” flash pattern (Lewis, 2016) while female *Pn. pyralis* were collected
113 from nearby patches of short grass and were identified based on their flash pattern and similar
114 morphology to the *Pn. pyralis* males (Lewis, 2016). Female *Pt. versicolor* were collected near
115 *Pn. pyralis* females and identified based on their green-shifted flash color and morphology
116 (Lewis, 2016). Additional *Pn. pyralis* males were collected to provision the mated *Pt. versicolor*
117 females. We collected *Pt. versicolor* and *Pn. pyralis* in a residential area (State College, Centre
118 Co, PA; 40° 47' 03" N, 77° 52' 25" W) into two separate 16-oz deli container “nurseries” kept at
119 ambient room temperature (20-22 °C); each nursery contained a handful of moist sphagnum
120 moss on top of moist soil (2-in deep; silt loam, collected from certified organic fields at the
121 Russell E. Larson Agricultural Research Center at Rock Springs, PA, U.S.A.; 40° 42' 52" N, 77°
122 56' 46" W). Both *Pn. pyralis* females mated within a few minutes of collection.

123 Female *Pt. versicolor* and *Pn. pyralis* laid eggs within the following 3 days (50+ *Pt.*
124 *versicolor* eggs and 100+ *Pn. pyralis* eggs; we did not attempt more accurate counts to avoid
125 damaging eggs). Under ambient temperatures (20-22 °C), first-instar larvae of both species began
126 to emerge three weeks after eggs were laid (5 August 2020). We kept *Pt. versicolor* larvae in the
127 nursery chambers for two weeks, and then, after we observed significant cannibalism among

128 larvae, moved them into individual soil-lined 1-oz polypropylene portion containers. As with
129 larvae collected and reared from 2019, developing *Pt. versicolor* were fed moistened cat food
130 (Grain-Free Real Chicken Recipe Dry Cat Food, Whole Earth Farm™, Merrick Pet Care Inc.,
131 Amarillo, TX, USA) in addition to pieces of freeze-killed *Lumbricus terrestris* (Josh's Frogs,
132 Owosso, MI). As evidence of the hypothesis that *Pn. pyralis* larvae are specialist on earthworms,
133 *Pn. pyralis* larvae did not feed on cat food but did feed gregariously on freeze-killed *L. terrestris*.
134 Unlike *Pt. versicolor*, *Pn. pyralis* failed to thrive in isolation, so they were kept in the nursery
135 chamber until starting the toxicity assay.

136

137 **Toxicity assay on Late-instar *Photuris versicolor***

138 We started the toxicity assay with late-instar *Photuris versicolor* on 22 June 2020. We
139 used 1-oz polypropylene portion containers containing 8 g of dry soil (same soil source as
140 nursery chambers) for our assay containers. To the soil in each assay container, we added 0.5 mL
141 of the appropriate clothianidin stock solution, allowed the acetone to completely evaporate, then
142 added 2 mL of DI water to moisten the soil and to achieve clothianidin concentrations of 0 ng g⁻¹,
143 10¹ ng g⁻¹ soil, 10² ng g⁻¹ soil, 10³ ng g⁻¹ soil, 10⁴ ng g⁻¹ soil. We chose this concentration
144 range (10¹ - 10⁴ ng g⁻¹ soil) to encompass the range of neonicotinoid concentrations in soil that
145 have been measured in potential firefly habitats (Knoepp et al., 2012; Schaafsma et al., 2015;
146 Pearsons et al., 2021).

147 After setting up assay containers, we weighed the late-instar *Pt. versicolor* and randomly
148 assigned each to a particular clothianidin concentration (ensuring all larvae in each dose-set were
149 sourced from the same location). All late-instar *Pt. versicolor* were over 12 months old at the
150 start of the assay, and were over 10-mm long and >50 mg (Table 1). Each clothianidin

151 concentration (0, 10¹ ng g⁻¹ soil, 10² ng g⁻¹ soil, 10³ ng g⁻¹ soil , 10⁴ ng g⁻¹ soil) was replicated
152 six times (30 late-instar *Pt. versicolor* larvae in total). We recorded firefly status at 1, 4, and 24
153 h, and every day for an additional 99 d. Fireflies were categorized as dead (D), exhibiting a toxic
154 response (T), or apparently healthy (A). A larva was assumed dead if it did not respond to gentle
155 prodding with forceps. If a larva was flipped on its back and/or demonstrating repetitive
156 twitching of its legs or head, it was recorded as exhibiting a toxic response (T). Fireflies were
157 recorded as apparently healthy (A) if they exhibited a usual response to prodding from blunt
158 forceps (Fig 1A; quickly curled up on its side, often glowing). During the toxicity assay, we fed
159 larvae once a week by carefully transferring individuals out of the assay containers into clean
160 containers lined with moistened filter paper and containing a piece of moistened cat food. After
161 24 h, we returned fireflies to the assay containers and noted if the cat food had obvious signs of
162 feeding (Fig 1B). Feeding activity for each week was measured as a simple binary (0 = no
163 obvious signs of feeding, 1 = obvious signs of feeding). At each status check, we noted if a
164 firefly had constructed a protective soil chamber, then carefully dismantled the chamber to check
165 larval status. Larvae often re-built soil chambers by the next day; if a larva built soil chambers on
166 multiple consecutive days (feeding days as an exception), we noted this behavior as a “period of
167 chamber formation.” Assay containers were kept in a dark drawer except when doing daily
168 checks, and we misted containers with DI water as needed to keep the soil from drying out.

169

170 **Toxicity assay on early-instar *Photuris versicolor***

171 The toxicity assay with early-instar *Photuris versicolor* was similar to the assay with late-
172 instar larvae, except we added half the amount of soil (4 g) and half the volume of clothianidin
173 stock solutions (0.25 mL) to each assay container to achieve the same clothianidin concentrations

174 (0, 10^1 ng g⁻¹ soil, 10^2 ng g⁻¹ soil, 10^3 ng g⁻¹ soil , 10^4 ng g⁻¹ soil). All early-instar *Pt. versicolor*
175 were less than 3 months old and weighed between 3 and 15 mg. On 17 Sept 2020, we started
176 trials with early-instar *Pt. versicolor* (three replicates at each concentration, 15 larvae in total),
177 feeding them cat food once a week and recording their status at 1, 4, and 24 h, and every day for
178 10 d, then twice a week for an additional 90 d. Unlike for late-instar *Pt. versicolor*, we fed early-
179 instars by directly placing moistened cat food in the assay containers (we removed the food 24 h
180 later after noting if food had been damaged [1] or not [0]).

181

182 **Toxicity assay on early-instar *Photinus pyralis***

183 As with the early-instar *Pt. versicolor* assay, the *Photinus pyralis* assay was run in 1-oz
184 polypropylene portion containers containing 4 g of soil with 0.25 mL doses of clothianidin stock
185 solutions (to achieve 0, 10^1 ng g⁻¹ soil, 10^2 ng g⁻¹ soil, 10^3 ng g⁻¹ soil , 10^4 ng g⁻¹ soil). All early-
186 instar *Pn. pyralis* were less than 3 months old and weighed between 0.6 and 2.4 mg. On 17 Sept
187 2020, we started the assay on early-instar *Pn. pyralis*, exposing larvae in sets of five (five larvae
188 per container, three replicates at each concentration, 75 larvae in total), recorded their status at 1,
189 4, and 24 h, and every day for 10 d, then at least twice a week for an additional 20 d. We
190 terminated the *Pn. pyralis* assay earlier than the *Pt. versicolor* assays due to an acarid mite
191 infestation, which rapidly increased larval mortality across all doses. During the assay, we fed
192 *Pn. pyralis* pieces of earthworm (*L. terrestris*) in the same manner that early-instar *Pt. versicolor*
193 were fed cat food.

194

195 **Statistical Analysis**

196 We performed all statistical analyses in R (v4.0.4) (R Core Team, 2021). For each firefly
197 cohort, we calculated median toxic concentrations (TC_{50}) and median lethal concentrations
198 (LC_{50}) at 24 h, 7 d, and 30 d of exposure using probit analysis (LC_PROBIT from the “ecotox”
199 package; Robertson et al., 2017; Hlina et al., 2019); for TC_{50} estimates, we included both sub-
200 lethal and lethal responses, while LC_{50} estimates were based on mortality alone. To assess long-
201 term survivorship across clothianidin levels, we used the Kaplan-Meier method (“survival”
202 functions SURVDIFF and PAIRWISE_SURVDIFF; Therneau, 2021; Therneau and Grambsch,
203 2000). To determine how clothianidin exposure affected firefly behavior, we used non-
204 parametric Mann-Whitney U tests (WILCOX.TEST) to compare feeding frequency and soil-
205 chamber construction across clothianidin doses; we made pairwise comparisons using Wilcoxon
206 rank sum tests with continuity corrections (PAIRWISE.WILCOX.TEST). As firefly larvae
207 reduce feeding before pupation (McLean et al. 1972), we excluded the two feeding events
208 preceding pupation for feeding assessments.

209

210 **Results**

211 **24 h, 7 d, and 30 d TC_{50} and LC_{50} estimates**

212 Dose-response curves and estimated TC_{50} and LC_{50} indicate that *Photuris versicolor* and
213 *Photinus pyralis* were surprisingly tolerant of exposure to clothianidin (Table 2 and Fig 2-4).
214 Reliable TC_{50} and LC_{50} estimates were limited by our small sample sizes and low acute mortality
215 within the tested concentration range. Overall, TC_{50} values ranged from 500 ng g⁻¹ to 2,000 ng g⁻¹
216 while LC_{50} values exceeded our test limit (above 10,000 ng g⁻¹).

217

218 **Firefly Survival**

219 Clothianidin exposure significantly reduced long-term firefly survival at high
220 concentrations (Fig 5). Between one and four hours after initial exposure, half of the late-instar
221 *Pt. versicolor* larvae and 87% of the early-instar *Pn. Pyralis* larvae exposed to the highest
222 clothianidin concentration (10,000 ng g⁻¹) began to exhibit toxic responses. By 24 h, all six late-
223 instar *Pt. versicolor* exposed to the highest clothianidin concentration (10,000 ng g⁻¹) exhibited a
224 toxic response (Fig 2A); these larvae never recovered and died by day 84. *Photuris* larvae were
225 somewhat tolerant to lower clothianidin concentrations (10 ng g⁻¹ or 100 ng g⁻¹) and neither late-
226 nor early-instar larvae exposed to low concentrations had significantly lower 100 d survival
227 probability compared to controls (Fig 5A-B). All *Pt. versicolor* in the control treatment either
228 pupated (2 out of 6 late-instar larvae) or survived through day 100 (4 out of 6 late-instar larvae,
229 all three early-instar larvae). Although the experiment was terminated at 30 d due to the mite
230 infestation, early-instar *Pn. Pyralis* exposed to 1,000 ng g⁻¹ clothianidin showed marginally non-
231 significant reduced survivorship ($P = 0.07$) while *Pn. pyralis* exposed to 10,000 ng g⁻¹
232 clothianidin showed significantly reduced survivorship ($P < 0.0001$) compared to controls (Fig.
233 5C).

234

235 **Feeding Behavior**

236 Clothianidin exposure significantly reduced the number of times firefly larvae fed (Fig
237 6). During the toxicity assays, no *Pn. pyralis* or *Pt. versicolor* larvae exposed to the highest
238 clothianidin concentration (10,000 ng g⁻¹ soil) fed. Late-instar *Pt. versicolor* exposed to 1,000 ng
239 g⁻¹ soil fed significantly less frequently than control larvae ($\chi^2_4 = 16.3$, $P = 0.003$), and early-
240 instar *Pn. pyralis* larvae fed significantly less at higher doses (1,000 ng g⁻¹ and 10,000 ng g⁻¹)
241 compared to the control or lower doses ($\chi^2_1 = 12.4$, $P = 0.0004$).

242

243 **Soil-Chambers, Molting, and Pupation of Late-instar *Photuris versicolor***

244 The 14 late-instar *Photuris* larvae that survived as larvae through day 100 went through 1
245 to 5 periods of consecutive days when they regularly formed protective soil chambers (median =
246 2 periods) and spent anywhere from 1 to 20 total days in soil chambers (median = 9 d). Larvae
247 exposed to 10,000 ng g⁻¹ clothianidin never constructed soil chambers while larvae exposed to 1
248 ppm clothianidin spent significantly fewer days in soil chambers than larvae exposed to 10 ng g⁻¹
249 ($P = 0.01$; Fig 7).

250 Formation of protective soil chambers did not correspond with molting or pupation, and
251 all recorded molting and pupation events occurred outside soil chambers, on the soil surface.
252 Late-instar *Pt. versicolor* larvae only molted once or twice, irrespective of how frequently or for
253 how long they built soil chambers (larvae that survived through 100 days; frequency: $R^2_{adj} = -$
254 0.09 , $F_{1,10} = 0.10$, $P = 0.76$; duration: $R^2_{adj} = -0.02$, $F_{1,10} = 0.81$, $P = 0.39$). Six of the thirty late-
255 instar *Pt. versicolor* larvae pupated; five of which successfully eclosed within 35 d of starting the
256 assay (two controls, one at 10 ng g⁻¹, two at 100 ng g⁻¹) and one which was unsuccessful (1,000
257 ng g⁻¹). The unsuccessful larva failed to shed its last-instar exoskeleton and died during the pupal
258 stage. At 35 d, three of the larvae exposed to the highest clothianidin concentration (10,000 ng g⁻¹
259 ¹) were still alive, but none of these larvae ever entered a pupal stage. Of the individuals that
260 successfully eclosed, three were lab-reared from eggs laid in 2019 (3 out of 5) while only two
261 were wild-collected (2 out of 25).

262

263 **Discussion**

264 *Photuris versicolor* complex and *Photinus pyralis* larvae did not significantly respond to
265 clothianidin concentrations at or below 100 ng g⁻¹ soil, but both firefly species exhibited
266 significant toxic responses to higher concentrations. Although some of the larvae exposed to
267 10,000 ng clothianidin g⁻¹ soil showed a toxic response within four hours of exposure, compared
268 to other soil invertebrates, larvae of these two firefly species were relatively tolerant to
269 clothianidin-treated soil. Twenty-four hour TC₅₀ values for *Pt. versicolor* and *Pn. pyralis* were
270 over 2× and 30× the TC₅₀ values for the earthworm *Eisenia andrei* and the collembolan *Folsomia*
271 *candida*, respectively (de Lima e Silva et al., 2020). Twenty-four hour LC₅₀ values exceeded our
272 maximum test concentration of 10⁴ ng clothianidin g⁻¹ soil (10 ppm), indicating higher tolerance
273 to clothianidin compared to other soil-dwelling beetles (*Agriotes* spp. [Elateridae] and *Atheta*
274 *coriaria* [Staphylinidae]; van Herk et al., 2007; Cloyd et al., 2009). The one other study which
275 tested neonicotinoid toxicity to fireflies observed 13% survival of aquatic *Luciola lateralis* larvae
276 after 24 h of exposure to 10⁵ ng thiamethoxam mL⁻¹ in water (Lee et al., 2008); these results
277 suggest that fireflies as a group may be somewhat tolerant to neonicotinoid exposure, although
278 this is likely a tenuous conclusion because it is based on just two studies that represent less than
279 0.2% of all described firefly species (Lewis et al., 2020). Tolerance to neonicotinoids may partly
280 explain why populations of *Pt. versicolor* and *Pn. pyralis* do not appear to be declining as fast as
281 rarer firefly species (Reed et al. 2020), which may be more sensitive to neonicotinoid exposure.
282 *Pt. versicolor* and *Pn. pyralis* may tolerate clothianidin exposure due to multiple behavioral,
283 morphological, and biochemical processes that could limit their sensitivity to clothianidin
284 (Alyokhin et al., 2008).

285 Behavioral avoidance of neonicotinoids has been observed across insect orders and beetle
286 families (Easton and Goulson, 2013; Fernandes et al., 2016; Pisa et al., 2021; Korenko et al.,

287 2019), and the results of this current study provide some support for behavioral avoidance of
288 neonicotinoids by Lampyridae. Although firefly larvae could not avoid dermal exposure to the
289 treated soil in our arenas, they may have decreased oral exposure by limiting construction of
290 their soil chambers. To form soil chambers, *Pt. versicolor* larvae manipulate soil with their
291 mouthparts (Buschman, 1984), providing a potentially more toxic pathway for neonicotinoid
292 exposure (Decourtye and Devillers, 2010). Because neonicotinoids are repellant to other beetle
293 species (Easton and Goulson, 2013), neonicotinoid-treated soil could have repulsed firefly
294 larvae, possibly explaining reduced chamber formation above 1000 ng clothianidin g⁻¹ soil.
295 Alternatively, sub-lethal neonicotinoid exposure may simply decrease the ability of fireflies to
296 construct soil chambers. Choice-based avoidance studies could be used to test if avoidance or
297 direct toxicity drove the decreased time late-instar *Pt. versicolor* spent constructing and
298 inhabiting soil chambers at high-clothianidin concentrations.

299 In addition to behavioral avoidance, specific morphological and metabolic characteristics
300 of fireflies may protect *Pt. versicolor* and *Pn. pyralis* larvae from toxic clothianidin exposure.
301 Unlike many other soil invertebrates (e.g., earthworms and mollusks), firefly larvae have a
302 comparably protective cuticle that may act as an effective barrier against neonicotinoid uptake
303 (Decourtye and Devillers, 2010; Wang et al., 2012). And even when clothianidin is absorbed,
304 insects can resist target-site exposure by quickly detoxify and/or excrete neonicotinoids (Olson et
305 al., 2000; Alyokhin et al., 2008). Although there has been no work on neonicotinoid metabolism
306 by fireflies, *Pt. versicolor* and *Pn. pyralis* may upregulate detoxification enzymes after
307 clothianidin exposure, similar to an aquatic firefly species after exposure to benzo[a]pyrene
308 (Zhang et al., 2021). Additionally, *Pt. versicolor* and *Pn. pyralis* may be tolerant to clothianidin
309 if neonicotinoids have a low binding affinity to target sites on firefly neurons. Neonicotinoids

310 primarily target nicotinic acetylcholine receptors (nAChRs), which regulate cation movement
311 and neuron firing in response to acetylcholine levels (Matsuda et al., 2020). Neonicotinoid
312 insecticides agonistically bind to these receptors, forcing ion channels open, leading to spasms
313 and eventual paralysis (Simon-Delso et al., 2015). As neonicotinoids have broad activity across
314 insect orders (Matsuda et al., 2020), it is unlikely that clothianidin has a low binding affinity for
315 nAChRs of *Pt. versicolor* and *Pn. pyralis*.

316 There is also the unlikely possibility that extensive neonicotinoid use has exerted
317 selection pressure on the firefly populations in central Pennsylvania to evolve resistance to
318 clothianidin. The way neonicotinoids are currently used is a perfect storm for developing
319 insecticide resistance (Tooker et al., 2017), and while most concern has focused on resistance-
320 development in herbivorous pest species, biocontrol agents and other predatory arthropods can
321 develop insecticide tolerance and resistance in response to heavy insecticide use (Bielza, 2016;
322 Mota-Sanchez and Wise, 2021). Although insecticide-resistance is thought to be rare among
323 biocontrol agents, lady beetles (Coleoptera: Coccinellidae) in particular, have been found to
324 develop resistance to a variety of broad-spectrum insecticides, including neonicotinoids (Tang et
325 al., 2015). Insecticide resistance has not been studied in many non-pest species (including
326 lampyrids), but if the selection pressure is high enough, firefly populations could evolve
327 increased tolerance or even resistance to neonicotinoid insecticides.

328 Differences among any of these potential mechanisms are likely driving differences in
329 tolerance between the two firefly species, namely, the dramatically reduced feeding response of
330 *Pn. pyralis* to clothianidin exposure. Although this difference could have been exacerbated by
331 mite pressure and the smaller body size of early-instar *Pn. pyralis*, it is possible that *Pn. pyralis*

332 has higher uptake, higher active-site affinity, or lower metabolism of clothianidin as compared to
333 *Pt. versicolor*.

334 Despite their relative tolerance to clothianidin exposure, field-realistic neonicotinoid
335 concentrations may still pose a chronic threat to *Pt. versicolor* and *Pn. pyralis*. Although residual
336 neonicotinoid concentrations in soil are often below 100 ng g⁻¹ (Schaafsma et al., 2016;
337 Radolinski et al., 2019; Pearsons et al., 2021), concentrations can regularly exceed these levels
338 after agricultural applications (as high as 594 ng g⁻¹ 23 days after planting neonicotinoid-coated
339 seeds; Radolinski et al., 2019), after turf applications (3 × higher than in agronomic settings;
340 Armbrust and Peeler, 2002) and after soil drenches to manage hemlock wooly adelgid (over
341 4,000 ng AI g⁻¹ soil; Knoepp et al., 2012). Such high concentrations are well within the acutely
342 toxic and chronically lethal range for *Pt. versicolor* and *Pn. pyralis* larvae (Table 2).

343 Encountering such high concentrations are likely to be even more lethal under field conditions,
344 as firefly larvae that exhibited toxic responses in the laboratory would be vulnerable to predation
345 and starvation, two risks that can increase mortality from insecticides (Kunkel et al., 2001).
346 Additionally, further work is needed to assess if neonicotinoid exposure can exacerbate other
347 stressors affecting firefly populations (i.e., light pollution) or if neonicotinoids pose a significant
348 risk to firefly eggs or adults.

349 As observed with other predatory beetle species (*Cycloneda sanguinea* [Coccinellidae]
350 and *Chauliognathus flavipes* [Cantharidae]; Fernandes et al., 2016), firefly larvae exhibited
351 reduced feeding activity in response to high neonicotinoid exposure. Firefly larvae that feed less
352 frequently may have less successful eclosion rates, and those that do eclose may have lower
353 reproductive success. Additionally, the prey that fireflies encounter in neonicotinoid-
354 contaminated environments likely provide an additional neonicotinoid exposure route. *Photinus*

355 larvae primarily feed on earthworms (Lewis et al., 2020), which have been found to contain
356 neonicotinoid concentrations above 200 ng g⁻¹ when collected from soybean fields that were
357 planted with neonicotinoid-coated seeds (Douglas et al., 2015) and 700 ng g⁻¹ when collected
358 from treated cereal fields (Pelosi et al., 2021). Firefly larvae of other species are known to feed
359 on slugs (Barker, 2004), which can also contain high doses of neonicotinoids (500 ng g⁻¹),
360 leading to disrupted biological control provided by carabid beetles (Douglas et al. 2015).
361 Compounded with reduced prey availability in habitats where neonicotinoids are used (Ritchie et
362 al., 2019; Tooker and Pearsons, 2021), decreased feeding activity and high risks of further
363 neonicotinoid exposure through contaminated prey may explain why adult lampyrid densities are
364 significantly lower where clothianidin has been used as a seed coating (Disque et al., 2019), even
365 if acute mortality is low. Adult fireflies may also encounter neonicotinoid residues while resting
366 on sprayed vegetation or during oviposition into soil (Pisa et al, 2021), although the risk of such
367 exposure does not appear to have been explored.

368 Despite low acute mortality, the sublethal effects of clothianidin were surprising, as some
369 *Pt. versicolor* larvae survived in a severely intoxicated state (not feeding, not building protective
370 soil chambers, only occasionally moving legs and/or mandibles) for over two months. A similar
371 phenomenon has been observed in European wireworms (*Agriotes* spp. [Coleoptera: Elateridae])
372 after exposure to clothianidin, with individuals surviving and even recovering from a severely
373 intoxicated state that can last months (van Herk et al., 2007; Vernon et al., 2007). For pests like
374 *Agriotes* spp., such sub-lethal effects of clothianidin exposure could still decrease crop damage
375 but may exacerbate the risk of *Agriotes* spp. developing neonicotinoid resistance. For predators
376 like *Pt. versicolor*, this long-term intoxication may limit their potential to provide biological
377 control beyond what would be expected based on population declines.

378

379

380 **Conclusions**

381 As larvae of the two firefly species that we studied appear to be somewhat tolerant to
382 clothianidin-treated soil, neonicotinoids alone may not be significant direct factors in firefly
383 declines in North America, at least for common species. Nevertheless, firefly populations around
384 the world appear to be suffering from other stressors (e.g., habitat loss, reduced prey availability,
385 light pollution), and ecological research has demonstrated that animal populations exposed to
386 multiple stressors can suffer disproportionately more than what is suffered from a single stress
387 (Relyea and Mills 2001). Therefore, continued widespread contamination of larval firefly
388 habitats with neonicotinoids may hold potential to exacerbate the influence of other stressors on
389 declining firefly populations (Lewis et al., 2020). We encourage researchers with access to other
390 species of fireflies, particularly those with declining populations in areas where neonicotinoids
391 are commonly used, to explore their toxicological responses to insecticides.

392

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399

400 **References**

401

402 Alyokhin A, Baker M, Mota-Sanchez D, Dively G, Grafius E. 2008. Colorado Potato Beetle

403 resistance to insecticides. *American Journal of Potato Research* 85:395–413. DOI:

404 [10.1007/s12230-008-9052-0](https://doi.org/10.1007/s12230-008-9052-0).

405 Armbrust KL, Peeler HB. 2002. Effects of formulation on the run-off of imidacloprid from turf.

406 *Pest Management Science* 58:702–706. DOI: [10.1002/ps.518](https://doi.org/10.1002/ps.518).

407 Barker GM. 2004. *Natural Enemies of Terrestrial Molluscs*. Oxford, United Kingdom: CABI

408 Publishing.

409 Bielza P. 2016. Insecticide Resistance in Natural Enemies. In: Horowitz AR, Ishaaya I eds.

410 *Advances in Insect Control and Resistance Management*. Switzerland: Springer

411 International Publishing, 313–329. DOI: [10.1007/978-3-319-31800-4](https://doi.org/10.1007/978-3-319-31800-4).

412 Buschman LL. 1984. Larval biology and ecology of *Photuris* Fireflies (Lampyridae: Coleoptera)

413 in Northcentral Florida. *Journal of the Kansas Entomological Society* 57:7–16.

414 Cloyd RA, Timmons NR, Goebel JM, Kemp KE. 2009. Effect of pesticides on adult rove beetle

415 *Atheta coriaria* (Coleoptera: Staphylinidae) survival in growing medium. *Journal of*

416 *Economic Entomology* 102:1750–1758. DOI: [10.1603/029.102.0504](https://doi.org/10.1603/029.102.0504).

417 Decourtye A, Devillers J. 2010. Ecotoxicity of Neonicotinoid Insecticides to Bees. In: Thany SH

418 ed. *Insect Nicotinic Acetylcholine Receptors*. Advances in Experimental Medicine and

419 Biology. New York, NY: Springer New York, 85–95. DOI: [10.1007/978-1-4419-6445-8_8](https://doi.org/10.1007/978-1-4419-6445-8_8).

420 Disque HH, Hamby KA, Dubey A, Taylor C, Dively GP. 2019. Effects of clothianidin-treated

421 seed on the arthropod community in a mid-Atlantic no-till corn agroecosystem. *Pest*

422 *Management Science* 75:969–978. DOI: [10.1002/ps.5201](https://doi.org/10.1002/ps.5201).

- 423 Douglas MR, Rohr JR, Tooker JF. 2015. Neonicotinoid insecticide travels through a soil food
424 chain, disrupting biological control of non-target pests and decreasing soya bean yield.
425 *Journal of Applied Ecology* 52:250–260. DOI: [10.1111/1365-2664.12372](https://doi.org/10.1111/1365-2664.12372).
- 426 Douglas MR, Tooker JF. 2015. Large-scale deployment of seed treatments has driven rapid
427 increase in use of neonicotinoid insecticides and preemptive pest management in U.S. field
428 crops. *Environmental Science & Technology* 49:5088–5097. DOI: [10.1021/es506141g](https://doi.org/10.1021/es506141g).
- 429 Easton AH, Goulson D. 2013. The neonicotinoid insecticide imidacloprid repels pollinating flies
430 and beetles at field-realistic concentrations. *PLoS ONE* 8:e54819. DOI:
431 [10.1371/journal.pone.0054819](https://doi.org/10.1371/journal.pone.0054819).
- 432 Fernandes MES, Alves FM, Pereira RC, Aquino LA, Fernandes FL, Zanuncio JC. 2016. Lethal
433 and sublethal effects of seven insecticides on three beneficial insects in laboratory assays
434 and field trials. *Chemosphere* 156:45–55. DOI: [10.1016/j.chemosphere.2016.04.115](https://doi.org/10.1016/j.chemosphere.2016.04.115).
- 435 Godan D. 1983. *Pest slugs and snails: biology and control*. New York: Springer.
- 436 van Herk WG, Vernon RS, Clodius M, Harding C, Tolman JH. 2007. Mortality of five wireworm
437 species (Coleoptera: Elateridae), following topical application of clothianidin and
438 chlorpyrifos. *J. Entomol. Soc. Brit. Columbia* 103:55–64.
- 439 Hlina BL, Birceanu O, Robinson CS, Dhiyebi H, Wilkie MP. 2019. Changes in the sensitivity of
440 piscicide in an invasive species. *North American Journal of Fisheries Management*.
- 441 Knoepp JD, Vose JM, Michael JL, Reynolds BC. 2012. Imidacloprid movement in soils and
442 impacts on soil microarthropods in southern Appalachian eastern hemlock stands. *Journal of*
443 *Environmental Quality* 41:469–478. DOI: [10.2134/jeq2011.0306](https://doi.org/10.2134/jeq2011.0306).

- 444 Korenko S, Saska P, Kysilková K, Řezáč M, Heneberg P. 2019. Prey contaminated with
445 neonicotinoids induces feeding deterrent behavior of a common farmland spider. *Scientific*
446 *Reports* 9:15895. DOI: [10.1038/s41598-019-52302-6](https://doi.org/10.1038/s41598-019-52302-6).
- 447 Kunkel BA, Held DW, Potter DA. 2001. Lethal and sublethal effects of bendiocarb,
448 halofenozide, and imidacloprid on *Harpalus pennsylvanicus* (Coleoptera: Carabidae)
449 following different modes of exposure in turfgrass. *Journal of Economic Entomology*
450 94:60–67. DOI: [10.1603/0022-0493-94.1.60](https://doi.org/10.1603/0022-0493-94.1.60).
- 451 Lee K-Y, Kim Y-H, Lee J-W, Song M-K, Nam S-H. 2008. Toxicity of firefly, *Luciola lateralis*
452 (Coleoptera Lampyridae) to commercially registered insecticides and fertilizers. *Korean*
453 *Journal of Applied Entomology* 47:265–272.
- 454 Lewis S. 2016. *Silent Sparks: The Wondrous World of Fireflies*. Princeton, NJ: Princeton
455 University Press. DOI: [10.2307/j.ctv7h0sxp](https://doi.org/10.2307/j.ctv7h0sxp).
- 456 Lewis SM, Wong CH, Owens ACS, Fallon C, Jepsen S, Thancharoen A, Wu C, De Cock R,
457 Novák M, López-Palafox T, Khoo V, Reed JM. 2020. A global perspective on firefly
458 extinction threats. *BioScience* 70:157–167. DOI: [10.1093/biosci/biz157](https://doi.org/10.1093/biosci/biz157).
- 459 de Lima e Silva C, Rooij W, Verweij RA, Gestel CAM. 2020. Toxicity in neonicotinoids to
460 *Folsomia candida* and *Eisenia andrei*. *Environmental Toxicology and Chemistry* 39:548–
461 555. DOI: [10.1002/etc.4634](https://doi.org/10.1002/etc.4634).
- 462 Losey JE, Vaughan M. 2006. The economic value of ecological services provided by insects.
463 *BioScience* 56:311. DOI: [10.1641/0006-3568\(2006\)56\[311:TEVOES\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2006)56[311:TEVOES]2.0.CO;2).
- 464 Matsuda K, Ihara M, Sattelle DB. 2020. Neonicotinoid insecticides: molecular targets, resistance,
465 and toxicity. *Annual Review of Pharmacology and Toxicology* 60:241–255. DOI:
466 [10.1146/annurev-pharmtox-010818-021747](https://doi.org/10.1146/annurev-pharmtox-010818-021747).

- 467 McLean M, Buck J, Hanson FE. 1972. Culture and Larval Behavior of Photurid Fireflies.
468 *American Midland Naturalist* 87:133. DOI: [10.2307/2423887](https://doi.org/10.2307/2423887).
- 469 Mota-Sanchez D, Wise JC. 2021. Arthropod pesticide resistance database.
- 470 Olson ER, Dively GP, Nelson JO. 2000. Baseline Susceptibility to Imidacloprid and Cross
471 Resistance Patterns in Colorado Potato Beetle (Coleoptera: Chrysomelidae) Populations.
472 *Journal of Economic Entomology* 93:447–458. DOI: [10.1603/0022-0493-93.2.447](https://doi.org/10.1603/0022-0493-93.2.447).
- 473 Pearsons KA, Rowen EK, Elkin KR, Wickings K, Smith RG, Tooker JF. 2021. Small-grain cover
474 crops have limited effect on neonicotinoid contamination from seed coatings.
475 *Environmental Science & Technology* 55:4679–4687. DOI: [10.1021/acs.est.0c05547](https://doi.org/10.1021/acs.est.0c05547).
- 476 Pelosi C, Bertrand C, Daniele G, Coeurdassier M, Benoit P, Néliu S, Lafay F, Bretagnolle V,
477 Gaba S, Vulliet E, Fritsch C. 2021. Residues of currently used pesticides in soils and
478 earthworms: A silent threat? *Agriculture, Ecosystems & Environment* 305:107167. DOI:
479 [10.1016/j.agee.2020.107167](https://doi.org/10.1016/j.agee.2020.107167).
- 480 Pisa L, Goulson D, Yang E-C, Gibbons D, Sánchez-Bayo F, Mitchell E, Aebi A, van der Sluijs J,
481 MacQuarrie CJK, Giorio C, Long EY, McField M, Bijleveld van Lexmond M, Bonmatin J-
482 M. 2021. An update of the Worldwide Integrated Assessment (WIA) on systemic
483 insecticides. Part 2: impacts on organisms and ecosystems. *Environmental Science and*
484 *Pollution Research* 28:11749–11797. DOI: [10.1007/s11356-017-0341-3](https://doi.org/10.1007/s11356-017-0341-3).
- 485 R Core Team. 2021. *R: A Language and Environment for Statistical Computing*. Vienna, Austria:
486 R Foundation for Statistical Computing.
- 487 Radolinski J, Wu J, Xia K, Hession WC, Stewart RD. 2019. Plants mediate precipitation-driven
488 transport of a neonicotinoid pesticide. *Chemosphere* 222:445–452. DOI:
489 [10.1016/j.chemosphere.2019.01.150](https://doi.org/10.1016/j.chemosphere.2019.01.150).

- 490 Reed JM, Nguyen A, Owens ACS, Lewis SM. 2020. Linking the seven forms of rarity to
491 extinction threats and risk factors: an assessment of North American fireflies. *Biodiversity
492 and Conservation* 29:57–75. DOI: [10.1007/s10531-019-01869-7](https://doi.org/10.1007/s10531-019-01869-7).
- 493 Relyea RA, Mills N. 2001. Predator-induced stress makes the pesticide carbaryl more deadly to
494 gray treefrog tadpoles (*Hyla versicolor*). *Proceedings of the National Academy of Sciences*
495 98:2491–2496.
- 496 Ritchie EE, Maisonneuve F, Scroggins RP, Princz JI. 2019. Lethal and sublethal toxicity of
497 thiamethoxam and clothianidin commercial formulations to soil invertebrates in a natural
498 soil. *Environmental Toxicology and Chemistry* 38:2111–2120. DOI: [10.1002/etc.4521](https://doi.org/10.1002/etc.4521).
- 499 Robertson JL, Jones MM, Olguin E, Alberts B. 2017. *Bioassays with Arthropods*. Boca Raton,
500 FL, USA: CRC Press, Inc.
- 501 Sánchez-Bayo F, Wyckhuys KAG. 2019. Worldwide decline of the entomofauna: A review of its
502 drivers. *Biological Conservation* 232:8–27. DOI: [10.1016/j.biocon.2019.01.020](https://doi.org/10.1016/j.biocon.2019.01.020).
- 503 Schaafsma A, Limay-Rios V, Baute T, Smith J, Xue Y. 2015. Neonicotinoid insecticide residues
504 in surface water and soil associated with commercial maize (corn) fields in southwestern
505 Ontario. *PLOS ONE* 10:e0118139. DOI: [10.1371/journal.pone.0118139](https://doi.org/10.1371/journal.pone.0118139).
- 506 Schaafsma A, Limay-Rios V, Xue Y, Smith J, Baute T. 2016. Field-scale examination of
507 neonicotinoid insecticide persistence in soil as a result of seed treatment use in commercial
508 maize (corn) fields in southwestern Ontario: Neonicotinoid persistence in cultivated field
509 soil. *Environmental Toxicology and Chemistry* 35:295–302. DOI: [10.1002/etc.3231](https://doi.org/10.1002/etc.3231).
- 510 Simon-Delso N, Amaral-Rogers V, Belzunces LP, Bonmatin JM, Chagnon M, Downs C, Furlan
511 L, Gibbons DW, Giorio C, Girolami V, Goulson D, Kreutzweiser DP, Krupke CH, Liess M,
512 Long E, McField M, Mineau P, Mitchell EAD, Morrissey CA, Noome DA, Pisa L, Settele J,

- 513 Stark JD, Tapparo A, Van Dyck H, Van Praagh J, Van der Sluijs JP, Whitehorn PR,
514 Wiemers M. 2015. Systemic insecticides (neonicotinoids and fipronil): trends, uses, mode of
515 action and metabolites. *Environmental Science and Pollution Research* 22:5–34. DOI:
516 [10.1007/s11356-014-3470-y](https://doi.org/10.1007/s11356-014-3470-y).
- 517 Sparks TC. 2013. Insecticide discovery: An evaluation and analysis. *Pesticide Biochemistry and*
518 *Physiology* 107:8–17. DOI: [10.1016/j.pestbp.2013.05.012](https://doi.org/10.1016/j.pestbp.2013.05.012).
- 519 Tabaru Y, Kouketsu T, Oba M, Okafuji S. 1970. Effects of some organophosphorus insecticides
520 against the larvae of Genji firefly, *Luciola cruciata* and their prey, Japanese melanical snail
521 *Semisulcospira bensoni*. *Medical Entomology and Zoology* 21:178–181.
- 522 Tang L-D, Qiu B-L, Cuthbertson AGS, Ren S-X. 2015. Status of insecticide resistance and
523 selection for imidacloprid resistance in the ladybird beetle *Propylaea japonica* (Thunberg).
524 *Pesticide Biochemistry and Physiology* 123:87–92. DOI: [10.1016/j.pestbp.2015.03.008](https://doi.org/10.1016/j.pestbp.2015.03.008).
- 525 Therneau TM. 2021. *A Package for Survival Analysis in R*.
- 526 Therneau TM, Grambsch PM. 2000. *Modeling Survival Data: Extending the Cox Model*. New
527 York: Springer.
- 528 Tooker JF, Douglas MR, Krupke CH. 2017. Neonicotinoid seed treatments: Limitations and
529 compatibility with integrated pest management. *Agricultural & Environmental Letters* 2:1–
530 5. DOI: [10.2134/ael2017.08.0026](https://doi.org/10.2134/ael2017.08.0026).
- 531 Tooker JF, Pearsons KA. 2021. Newer characters, same story: neonicotinoid insecticides disrupt
532 food webs through direct and indirect effects. *Current Opinion in Insect Science* 46:50–56.
533 DOI: [10.1016/j.cois.2021.02.013](https://doi.org/10.1016/j.cois.2021.02.013).

- 534 Vernon RS, Van Herk W, Moffat C, Harding C. 2007. European wireworms (*Agriotes* spp.) in
535 North America: toxicity and repellency of novel insecticides in the laboratory and field.
536 *IOBC/wprs Bulletin* 30:35–41.
- 537 Wang Y, Cang T, Zhao X, Yu R, Chen L, Wu C, Wang Q. 2012. Comparative acute toxicity of
538 twenty-four insecticides to earthworm, *Eisenia fetida*. *Ecotoxicology and Environmental*
539 *Safety* 79:122–128. DOI: [10.1016/j.ecoenv.2011.12.016](https://doi.org/10.1016/j.ecoenv.2011.12.016).
- 540 Zhang Q-L, Jiang Y-H, Dong Z-X, Li H-W, Lin L-B. 2021. Exposure to benzo[a]pyrene triggers
541 distinct patterns of microRNA transcriptional profiles in aquatic firefly *Aquatica wuhana*
542 (Coleoptera: Lampyridae). *Journal of Hazardous Materials* 401:123409. DOI:
543 [10.1016/j.jhazmat.2020.123409](https://doi.org/10.1016/j.jhazmat.2020.123409).

Table 1 (on next page)

Summary details of the firefly larvae (three cohorts of two species) used for toxicological assays.

Species	Instar group	Age (months)	Mean weight \pm SD (mg)	# used in assay	Source
<i>Pt. versicolor</i>	late	> 12	77 \pm 17	30	wild-collected
	early	< 3	9 \pm 3	15	lab-reared
<i>Pn. pyralis</i>	early	< 3	1.4 \pm 0.5	75	lab-reared

1

Table 2 (on next page)

Estimated median toxic concentrations (TC_{50}) and lethal concentrations (LC_{50}) for *Pt. versicolor* and *Pn. pyralis* exposure to clothianidin-contaminated soil.

95% confidence intervals (CI) are based on probit analyses. CIs are not shown where data did not fit a cumulative standard normal distribution. n.r. = no response in tested range.

Species	timeframe	TC ₅₀ (ng g ⁻¹ soil)	95% CI	LC ₅₀ (ng g ⁻¹ soil)	95% CI
<i>Pt. versicolor</i> , late-instar, 6 larvae / dose	24 h	1882	136–10,000+	> 10,000	-
	7 d	648	144–3047	> 10,000	-
	30 d	574	46–9895	> 10,000	-
<i>Pt. versicolor</i> , early-instar, 3 larvae / dose	24 h	> 10,000	-	n.r.	-
	7 d	1169	-	> 10,000	-
	30 d	1169	-	1169	-
<i>Pn. pyralis</i> , early-instar, 3 sets of 5 / dose	24 h	1726	836–3486	n.r.	-
	7 d	704	-	n.r.	-
	30 d	316	-	1591	246–10,000+

1

Figure 1

Healthy *Pt. versicolor* larvae (A) demonstrating a typical “curl and glow” response after being prodded with blunt forceps and (B) feeding on moistened cat food. (C) An intoxicated *Pt. versicolor* larva on its back, unable to right itself.

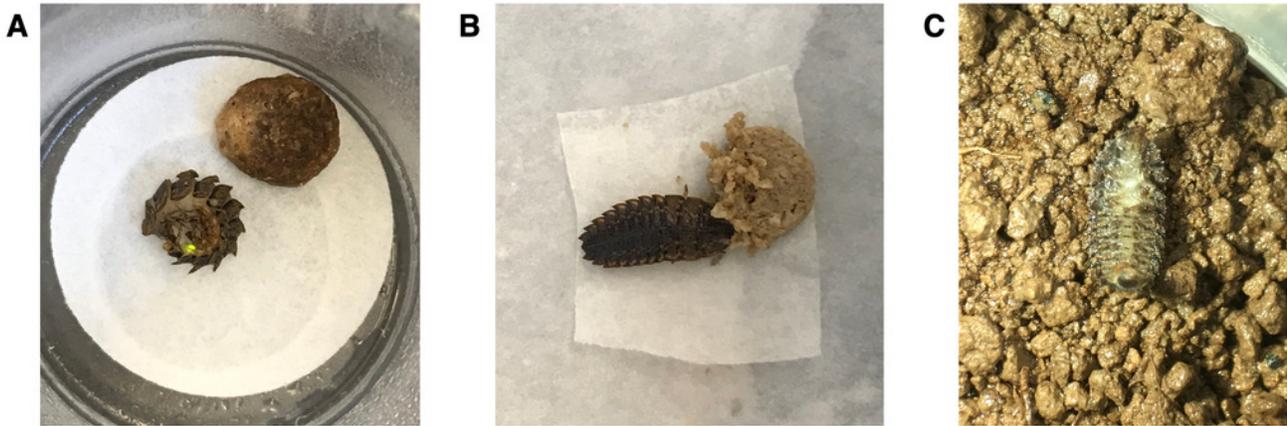


Figure 2

Dose-response curves for late-instar *Pt. versicolor* exposed to clothianidin-contaminated soil at 10, 100, 1000, and 10,000 ng clothianidin per gram of soil ($n = 6$ larvae for each concentration).

Toxic responses after (A) 24 h, (B) 7 d, and (C) 30 d, and lethal response after (D) 24 h, (E) 7 d, and (F) 30 d. Black dots in each panel represent mean responses at each insecticide concentration; the shaded area represents the 95% confidence interval for each curve. Blue diamonds represent the response of the control group. Dotted lines in each panel marks the 50% toxic response or mortality threshold.

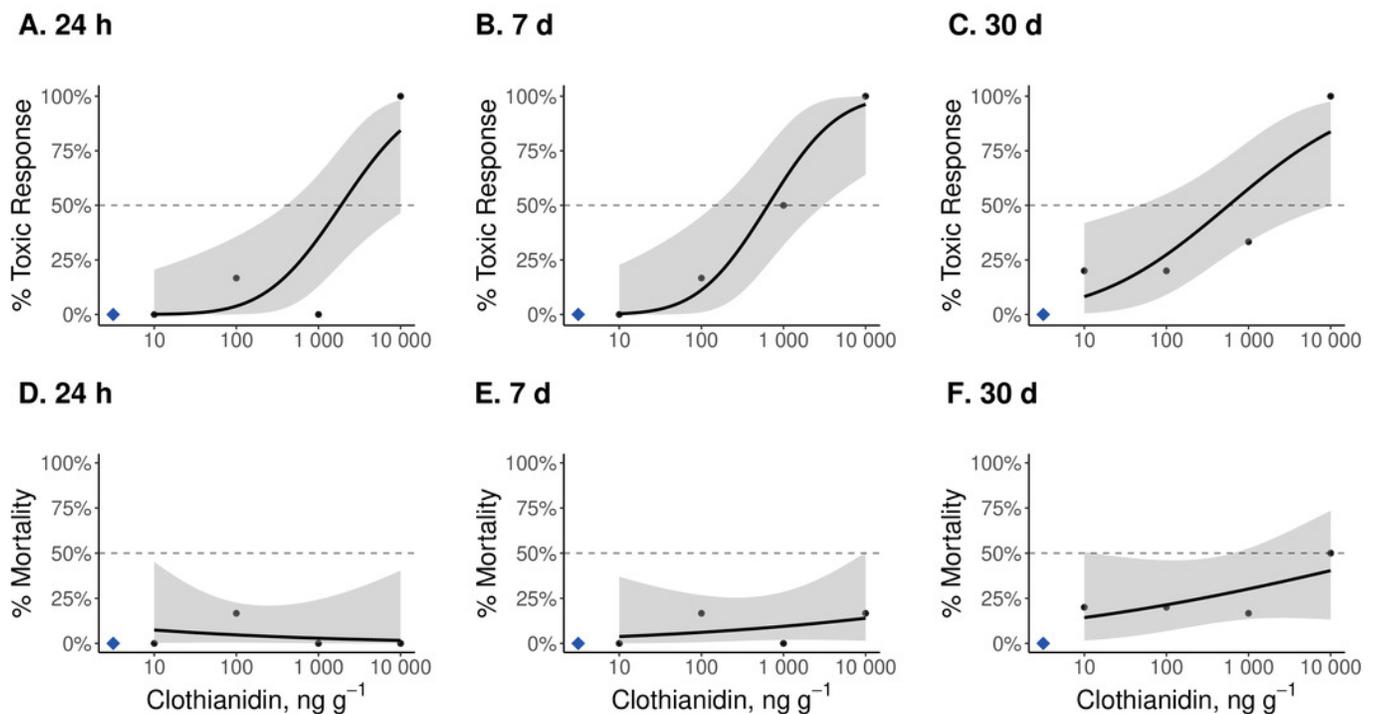


Figure 3

Dose-response curves for early-instar *Pt. versicolor* exposed to clothianidin-contaminated soil at 10, 100, 1000, and 10,000 ng clothianidin per gram of soil ($n = 3$ larvae for each concentration).

Toxic responses after (A) 24 h, (B) 7 d, and (C) 30 d, and lethal response after (D) 24 h, (E) 7 d, and (F) 30 d. Black dots in each panel represent mean responses at each insecticide concentration; the shaded area represents the 95% confidence interval for each curve. Blue diamonds represent the response of the control group. Dotted lines in each panel marks the 50% toxic response or mortality threshold.

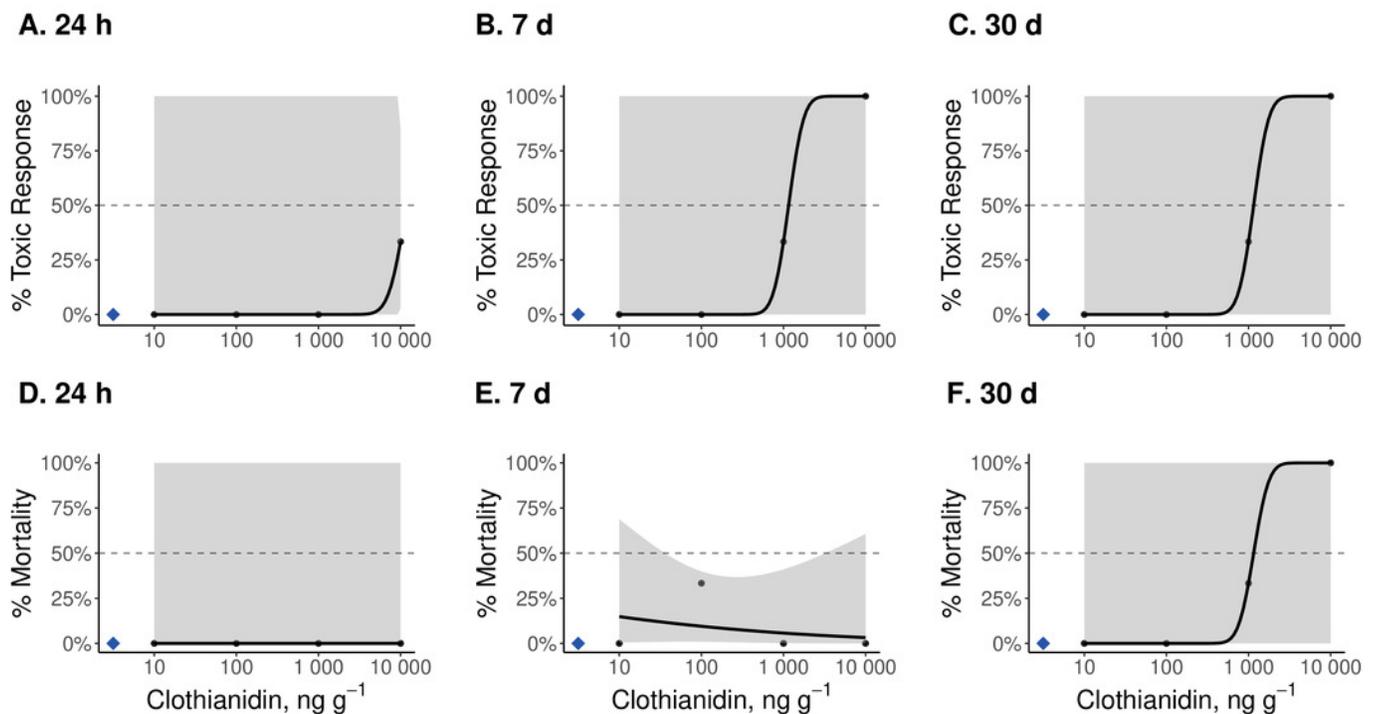


Figure 4

Dose-response curves for early-instar *Pn. pyralis* exposed to clothianidin-contaminated soil at 10, 100, 1000, and 10,000 ng clothianidin per gram of soil ($n = 3$ sets of 5 larvae for each concentration).

Toxic responses after (A) 24 h, (B) 7 d, and (C) 30 d, and lethal response after (D) 24 h, (E) 7 d, and (F) 30 d. Black dots in each panel represent mean responses at each insecticide concentration; the shaded area represents the 95% confidence interval for each curve. Blue diamonds represent the response of the control group. Dotted lines in each panel marks the 50% toxic response or mortality threshold.

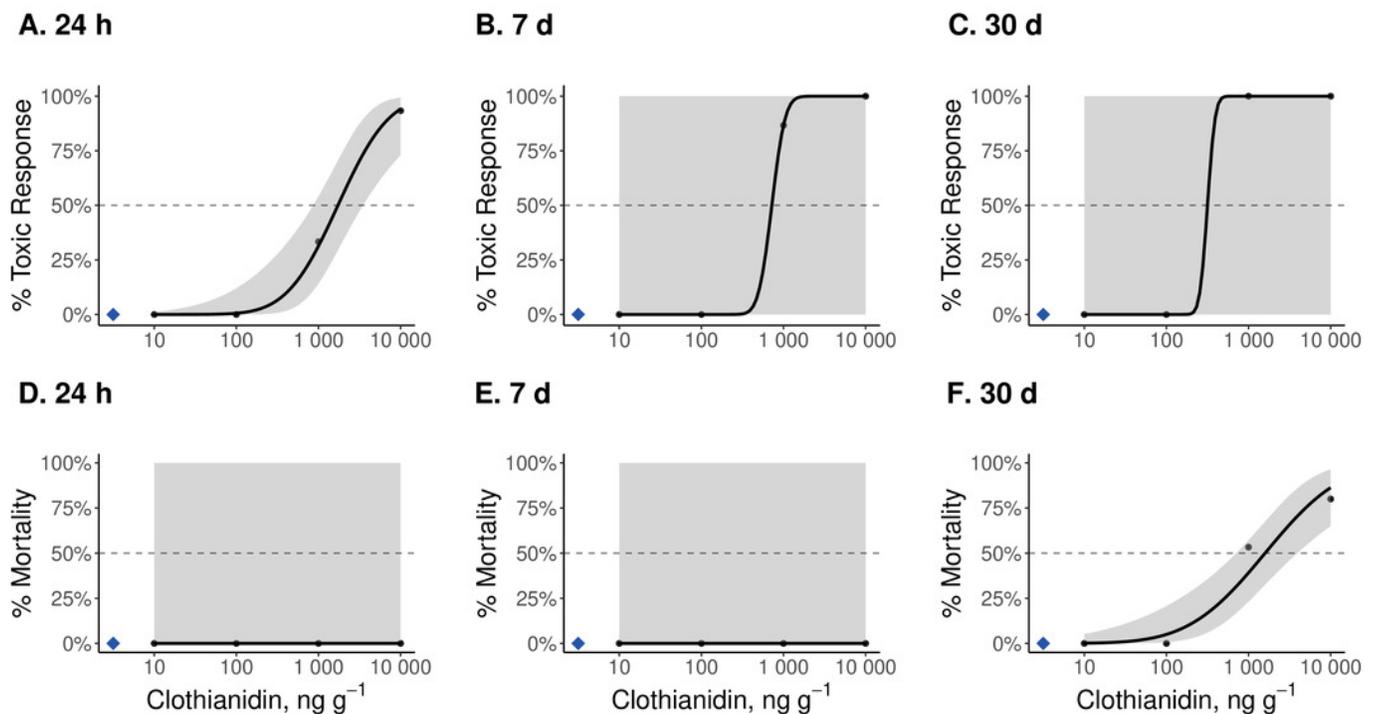


Figure 5

Survivorship curves.

A) late-instar *Pt. versicolor* (n=6 per concentration), B) early-instar *Pt. versicolor* (n=3 per concentration), and C) early-instar *Pn. pyralis* (n=15 per concentration) at different clothianidin concentrations. *P*-values next to each line indicate the significance of reduced survivorship compared to the control (with a Benjamini-Hochberg correction for multiple comparisons). *P*-values were excluded where survivorship was 100% and perfectly overlapped with control values (100 ng g⁻¹ in panel B, 10 and 100 ng g⁻¹ in panel C). Survival was significantly affected by clothianidin exposure (late-instar *Pt. versicolor*: $\chi^2_4 = 18$, $P = 0.001$; early-instar *Pt. versicolor*: $\chi^2_4 = 12.5$, $P = 0.01$; early-instar *Pn. pyralis*: $\chi^2_4 = 58.3$, $P < 0.0001$).

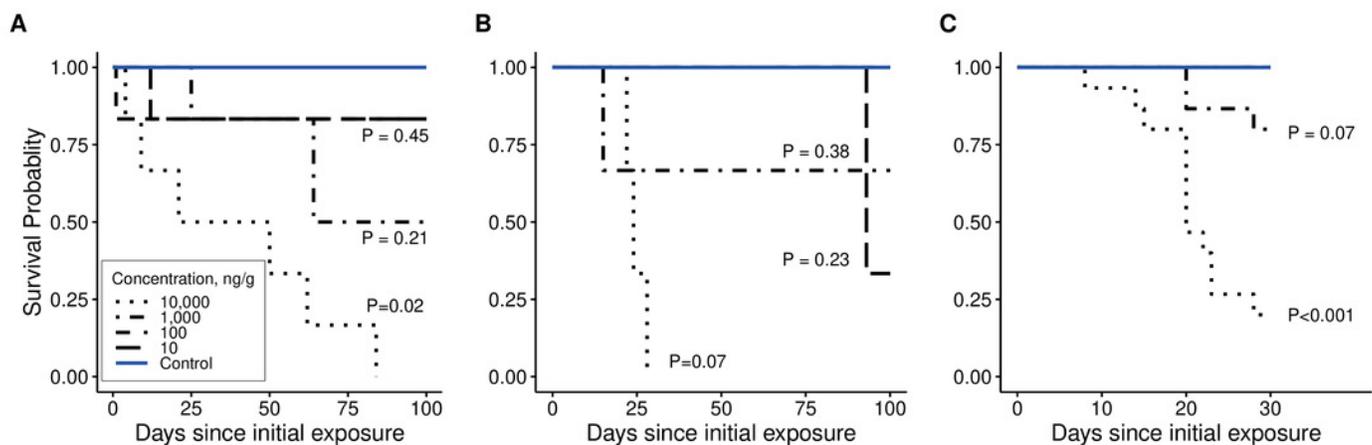


Figure 6

Percent of feeding opportunities taken by firefly larvae.

A) late-instar *Pt. versicolor* larvae ($\chi^2_4 = 15.8$, $P = 0.003$), (B) early-instar *Pt. versicolor* larvae ($\chi^2_4 = 8.2$, $P = 0.08$), and (C) early-instar *Pn. pyralis* larvae ($\chi^2_1 = 12.4$, $P = 0.0004$). Different letters indicate significant differences in feeding activity within each cohort at $P < 0.05$ (Benjamini-Hochberg correction for multiple comparisons).

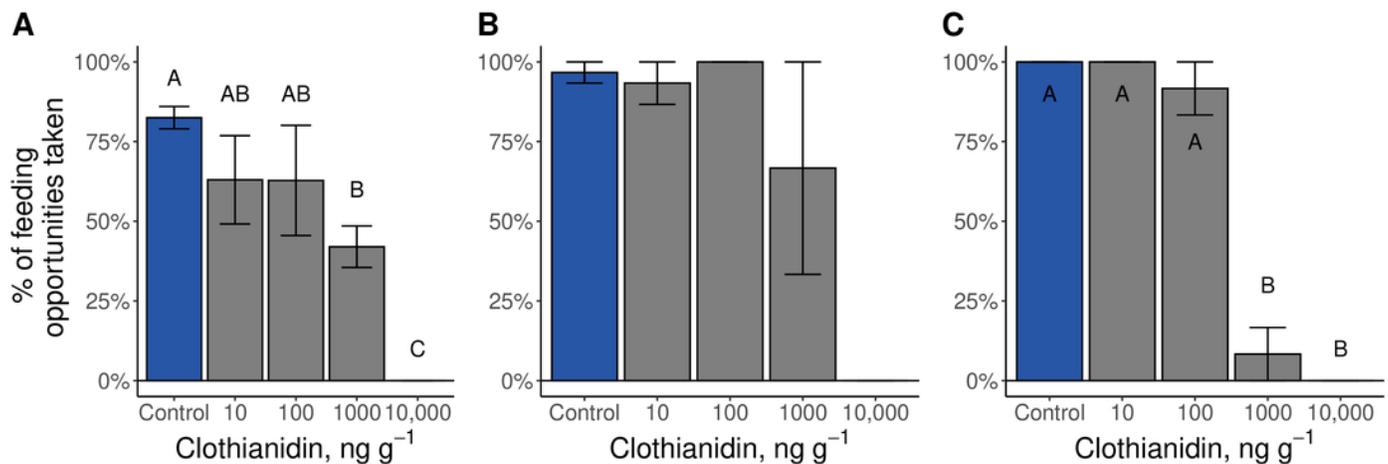


Figure 7

Amount of time that late-instar *Pt. versicolor* spent in soil chambers at different clothianidin-exposure levels ($\chi^2_4 = 18.4$, $P = 0.001$).

Different letters indicate significant differences at $P < 0.05$ (Benjamini-Hochberg correction for multiple comparisons).

