

Genomics versus mtDNA for resolving stock structure in the silky shark (*Carcharhinus falciformis*) (#51011)

1

First submission

Guidance from your Editor

Please submit by **13 Aug 2020** for the benefit of the authors (and your \$200 publishing discount) .



Structure and Criteria

Please read the 'Structure and Criteria' page for general guidance.



Custom checks

Make sure you include the custom checks shown below, in your review.



Author notes

Have you read the author notes on the [guidance page](#)?



Raw data check

Review the raw data.



Image check

Check that figures and images have not been inappropriately manipulated.

Privacy reminder: If uploading an annotated PDF, remove identifiable information to remain anonymous.

Files

Download and review all files from the [materials page](#).

4 Figure file(s)

2 Table file(s)

1 Raw data file(s)

! Custom checks

DNA data checks

- ! Have you checked the authors [data deposition statement](#)?
- ! Can you access the deposited data?
- ! Has the data been deposited correctly?
- ! Is the deposition information noted in the manuscript?

Vertebrate animal usage checks

- ! Have you checked the authors [ethical approval statement](#)?
- ! Were the experiments necessary and ethical?
- ! Have you checked our [animal research policies](#)?



Structure and Criteria

Structure your review

The review form is divided into 5 sections. Please consider these when composing your review:

1. BASIC REPORTING
2. EXPERIMENTAL DESIGN
3. VALIDITY OF THE FINDINGS
4. General comments
5. Confidential notes to the editor

You can also annotate this PDF and upload it as part of your review

When ready [submit online](#).

Editorial Criteria

Use these criteria points to structure your review. The full detailed editorial criteria is on your [guidance page](#).

BASIC REPORTING

- Clear, unambiguous, professional English language used throughout.
- Intro & background to show context. Literature well referenced & relevant.
- Structure conforms to [Peerj standards](#), discipline norm, or improved for clarity.
- Figures are relevant, high quality, well labelled & described.
- Raw data supplied (see [Peerj policy](#)).

EXPERIMENTAL DESIGN

- Original primary research within [Scope of the journal](#).
- Research question well defined, relevant & meaningful. It is stated how the research fills an identified knowledge gap.
- Rigorous investigation performed to a high technical & ethical standard.
- Methods described with sufficient detail & information to replicate.

VALIDITY OF THE FINDINGS

- Impact and novelty not assessed. Negative/inconclusive results accepted. *Meaningful* replication encouraged where rationale & benefit to literature is clearly stated.
- All underlying data have been provided; they are robust, statistically sound, & controlled.
- Speculation is welcome, but should be identified as such.
- Conclusions are well stated, linked to original research question & limited to supporting results.



The best reviewers use these techniques

Tip

Example

Support criticisms with evidence from the text or from other sources

Smith et al (J of Methodology, 2005, V3, pp 123) have shown that the analysis you use in Lines 241-250 is not the most appropriate for this situation. Please explain why you used this method.

Give specific suggestions on how to improve the manuscript

Your introduction needs more detail. I suggest that you improve the description at lines 57- 86 to provide more justification for your study (specifically, you should expand upon the knowledge gap being filled).

Comment on language and grammar issues

The English language should be improved to ensure that an international audience can clearly understand your text. Some examples where the language could be improved include lines 23, 77, 121, 128 - the current phrasing makes comprehension difficult.

Organize by importance of the issues, and number your points

- 1. Your most important issue*
- 2. The next most important item*
- 3. ...*
- 4. The least important points*

Please provide constructive criticism, and avoid personal opinions

I thank you for providing the raw data, however your supplemental files need more descriptive metadata identifiers to be useful to future readers. Although your results are compelling, the data analysis should be improved in the following ways: AA, BB, CC

Comment on strengths (as well as weaknesses) of the manuscript

I commend the authors for their extensive data set, compiled over many years of detailed fieldwork. In addition, the manuscript is clearly written in professional, unambiguous language. If there is a weakness, it is in the statistical analysis (as I have noted above) which should be improved upon before Acceptance.

Genomics versus mtDNA for resolving stock structure in the silky shark (*Carcharhinus falciformis*)

Derek W. Kraft ^{Corresp., 1}, Emily Conklin ¹, Evan Barba ¹, Melanie Hutchinson ^{1,2}, Robert J. Toonen ¹, Zac H. Forsman ¹, Brian W. Bowen ¹

¹ Hawai'i Institute of Marine Biology, University of Hawai'i, Kaneohe, Hawaii, USA

² Joint Institute of Marine and Atmospheric Research, Pacific Islands Fisheries Science Center, NOAA, University of Hawai'i, Honolulu, Hawaii, USA

Corresponding Author: Derek W. Kraft
Email address: Kraftd@hawaii.edu

Conservation genetic approaches for elasmobranchs have focused on regions of the mitochondrial genome or a handful of nuclear microsatellites. High-throughput sequencing offers a powerful alternative for examining population structure using many loci distributed across the nuclear and mitochondrial genomes. These single nucleotide polymorphisms are expected to provide finer scale and more accurate population level data; however, there have been few genomic studies applied to elasmobranch species. The desire to apply next-generation sequencing approaches is often tempered by the costs, which can be offset by pooling specimens prior to sequencing (pool-seq). In this study, we assess the utility of pool-seq by applying this method to the same individual silky sharks, *Carcharhinus falciformis* (Müller & Henle, 1839), previously surveyed with the mtDNA control region in the Atlantic and Indian Oceans (Clarke *et al.*, 2015). Pool-seq methods were able to recover the entire mitochondrial genome as well as thousands of nuclear markers. This volume of sequence data enabled the detection of population structure between regions of the Atlantic Ocean populations, undetected in the previous study (inter-Atlantic mitochondrial SNPs F_{ST} values comparison ranging from 0.029 to 0.135 and nuclear SNPs from 0.015 to 0.025). Our results lend further support that sampling the mitochondrial control region alone may fail to detect fine-scale population structure, and additional sampling across the genome may increase resolution for some species. Additionally, this study shows that the costs of analyzing thousands of loci using pool-seq methods are equivalent to the standard Sanger-sequenced markers and become less expensive when large numbers of individuals (>300) are analyzed.

1 **Genomics versus mtDNA for Resolving Stock Structure in the Silky Shark**
2 **(*Carcharhinus falciformis*)**

3

4 Derek W. Kraft^{1*}, Emily Conklin¹, Evan Barba¹, Melanie Hutchinson^{1,2}, Robert J. Toonen¹,

5 Zac H. Forsman¹, Brian W. Bowen¹

6

7 ¹Hawai‘i Institute of Marine Biology, University of Hawai‘i, 46-007 Lilipuna Road, Kāne‘ohe,

8 HI 96744

9 ²Joint Institute for Marine and Atmospheric Research, University of Hawai‘i, Pacific Islands

10 Fisheries Science Center, NOAA

11

12 Corresponding Author:

13 Derek W. Kraft¹

14 46-007 Lilipuna Rd, Kaneohe, HI 96744, USA

15

16 Email address: Kraftd@hawaii.edu

17

18

19

20

21

22

23

24

25

26 Abstract

27 Conservation genetic approaches for elasmobranchs have focused on regions of the
28 mitochondrial genome or a handful of nuclear microsatellites. High-throughput sequencing
29 offers a powerful alternative for examining population structure using many loci distributed
30 across the nuclear and mitochondrial genomes. These single nucleotide polymorphisms are
31 expected to provide finer scale and more accurate population level data; however, there have
32 been few genomic studies applied to elasmobranch species. The desire to apply next-generation
33 sequencing approaches is often tempered by the costs, which can be offset by pooling specimens
34 prior to sequencing (pool-seq). In this study, we assess the utility of pool-seq by applying this
35 method to the same individual silky sharks, *Carcharhinus falciformis* (Müller & Henle, 1839),
36 previously surveyed with the mtDNA control region in the Atlantic and Indian Oceans (Clarke *et*
37 *al.*, 2015). Pool-seq methods were able to recover the entire mitochondrial genome as well as
38 thousands of nuclear markers. This volume of sequence data enabled the detection of population
39 structure between regions of the Atlantic Ocean populations, undetected in the previous study
40 (inter-Atlantic mitochondrial SNPs F_{ST} values comparison ranging from 0.029 to 0.135 and
41 nuclear SNPs from 0.015 to 0.025). Our results lend further support that sampling the
42 mitochondrial control region alone may fail to detect fine-scale population structure, and
43 additional sampling across the genome may increase resolution for some species. Additionally,
44 this study shows that the costs of analyzing thousands of loci using pool-seq methods are
45 equivalent to the standard Sanger-sequenced markers and become less expensive when large
46 numbers of individuals (>300) are analyzed.

47

48 Introduction

49 Many elasmobranchs throughout the ocean have experienced devastating population declines due
50 to overfishing in both target and non-target fisheries (Musick et al., 2000; Clarke et al., 2006; F.
51 Ferretti et al., 2010; Heupel et al., 2014; Dulvy et al., 2014; Oliver et al., 2015; Dulvy &
52 Trebilco, 2018). These species are especially vulnerable to overfishing due to life history traits
53 such as late maturity, slow growth, low fecundity, and high juvenile mortality, which collectively
54 result in low intrinsic rate of population increase (Baum et al., 2003; Dulvy et al., 2008).
55 Elasmobranch populations take decades to recover from overfishing, and only if fishing pressure
56 is relieved for an extended period (Stevens et al., 2000). Furthermore, many threatened and
57 endangered elasmobranchs have little to no population genetic data that would assist in the
58 resolution of management units (reviewed in Rodrigo Rodrigues Domingues, Hilsdorf, & Gadig,
59 2018).



60
61 Genetically distinct populations are isolated management units known as stocks; however, stocks
62 can be defined on a smaller scale than genetic populations through other criteria, such as an
63 exclusive economic zone boundary. (Carvalho & Hauser, 1994; Ovenden et al., 2015). Reduced
64 gene flow indicates that if a population is overfished it will not be replenished by immigrants
65 from surrounding populations. This is why managing on a genetic stock-by-stock basis is
66 essential for successful maintenance of exploited species and is sorely needed for over-harvested
67 elasmobranchs (Dizon et al., 1993; Heist, 2004; Tallmon et al., 2010).

68
69 For the past two decades the standard for examining population structure in elasmobranchs has
70 been a section of the mitochondrial genome, usually the control region (mtCR) (e.g. Duncan et
71 al. 2006; Hoelzel et al. 2006; Keeney & Heist 2006; Castro et al. 2007; Whitney et al. 2012;

72 Clarke et al. 2015; reviewed in Domingues et al. 2017a). Though recent studies are moving
73 towards multi-marker approaches (Momigliano et al., 2017; Pazmiño et al., 2018; Green et al.,
74 2019), there is still a large body of literature focusing on mtCR. The mitochondrial genome has a
75 higher rate of mutation than most of the nuclear genome (Brown et al., 1979; Charlesworth &
76 Wright, 2001; Neiman & Taylor, 2009) and this rate of mutation is a key advantage in
77 vertebrates with slowly-evolving genomes (Avisé et al., 1992; Martin et al., 1992).
78 Elasmobranch mtDNA studies to date have been successful in elucidating population partitions
79 and evolutionary divergences, but the maternal inheritance of mtDNA can limit conclusions
80 about gene flow. Both mtDNA and nuclear markers often have concordant results in sedentary
81 species (e.g., Lavery et al., 1996; Avisé, 2004; Zink & Barrowclough, 2008; DiBattista et al.,
82 2015) but, when examined alone, may miss key components of population structure, particularly
83 in migratory fauna (Pardini et al., 2001; Bowen, Bass, Soares, & Toonen, 2005; Toews &
84 Brelsford, 2012). When highly mobile elasmobranchs are examined with both mtDNA and
85 nuclear markers (usually microsatellites), a different picture often emerges, in which females are
86 more resident and males are dispersive (Pardini, Jones, Noble, Kreiser, & Malcolm, 2001;
87 Schultz et al., 2008; Portnoy, McDowell, Heist, Musick, & Graves, 2010; S. A. Karl, Castro,
88 Lopez, Charvet, & Burgess, 2011; Daly-Engel, Randall, & Bowen, 2012; Bernard et al., 2017;
89 Domingues et al., 2018). Identifying outlier SNPs in the nuclear genome can highlight genes
90 possibly under selection, or show functional responses to environmental changes that have
91 important management consequences (Barrio et al., 2016; Fischer et al., 2013; Guo et al., 2016;
92 Jones et al., 2012). Therefore, the combination of mitochondrial and nuclear markers can yield
93 fundamental ecological and evolutionary insights.
94



95 High-throughput sequencing is a powerful tool for revealing fine-scale population structure that
96 may be missed by single locus studies (Andrews et al., 2016; Hohenlohe et al., 2018). However,
97 this method can be costly, especially when examining many individuals as is typical of
98 population genetic or phylogeography studies. For population genetics approaches based on
99 differences in allele frequencies among populations, equimolar pooling of samples before
100 sequencing is an affordable strategy for large scale genetic analysis (Schlötterer et al., 2014).
101 Several studies have successfully resolved population structure using a pooled site-associated
102 DNA approach known as pool-seq, including some in commercially valuable marine species
103 (e.g. Gautier et al., 2013; Mimee et al., 2015). Pool-seq provides allele frequencies for thousands
104 of loci distributed across the genome simultaneously, which gives greater statistical power that
105 can actually exceed the accuracy of allele frequency estimates based on individual sequencing
106 (Futschik & Schlötterer, 2010). Therefore, a comparison of results between the standard mtCR
107 analysis and high-throughput pool-seq is informative in evaluating the relative power and cost of
108 the two approaches for examining population structure in wild populations.



109

110 The silky shark (*Carcharhinus falciformis*) is the second most commonly harvested shark on
111 Earth (Oliver et al., 2015; Rice & Harley, 2013). They are one of the top contributors to the
112 shark fin trade and the most common elasmobranch bycatch species in tuna purse-seine fisheries
113 around the world (Cardenosa et al., 2018; Clarke et al., 2006; Oliver et al., 2015). This pelagic
114 shark, formerly abundant in all tropical oceans, has declined by an estimated 85% in the last 20
115 years, and is now listed as vulnerable and declining by the International Union for the
116 Conservation of Nature (Rice & Harley, 2013; IUCN, 2017). Currently silky shark population
117 assessments are conducted at the scale of regional fishery management organization, and

118 conservation management measures are implemented at this scale in the absence of genetic or
119 movement data to define population boundaries. Clarke et al. (2015) surveyed silky sharks across
120 these regional management regions and found the western Atlantic was strongly differentiated
121 from the Indian Ocean, but the North Atlantic, Gulf of Mexico, and Brazil could not be
122 differentiated and appeared to comprise a single population. In contrast, using the same mtCR
123 marker, Domingues et al. (2017b) examined five regions across the Western Atlantic and found
124 the North Western Atlantic was distinct from the South Western Atlantic.

125

126 In an era where wildlife management needs far exceed the financial resources to address them,
127 many seek to find the most accessible, robust, and economical means to define management
128 units. In this study, we provide a direct comparison of population genetic analysis methods
129 between Sanger sequencing of the mtCR region and high-throughput sequencing of regional
130 pools of individuals. The same individuals from Clarke et al. (2015) were re-sequenced using
131 pool-seq approaches. Regions re-sequenced included Gulf of Mexico, North West Atlantic, and
132 Brazil, as well as one geographically distant location in the Red Sea (*Fig 1*). We then evaluate
133 the economics of conducting pool-seq relative to conventional Sanger sequencing of these same
134 individuals. Ecological and management implications will be addressed in a subsequent
135 companion paper.

136

137 **Materials & Methods**

138 *Sampling and sequencing*

139 A total of 143 silky shark fin clips or muscle sections were sampled from commercial or artisanal
140 fisheries across four geographic regions and are the same samples examined in Clarke et al.

141 (2015). Specifically, we sampled the Gulf of Mexico (GM, n =39), the North Atlantic (NA, n =
142 33), Brazil (BR, n = 34), and the Red Sea (RS, n = 37).

143

144 DNA was extracted using Qiagen DNeasy Blood & Tissue kit (Qiagen, Mississauga, ON,
145 Canada), following manufacturer protocols. Extracted DNA quality was assessed visually by gel
146 electrophoresis and imaged using Gel Doc E-Z System (BIO RAD, Hercules, California, USA).

147 Only DNA aliquots with strong genomic DNA bands were further processed, while degraded or
148 overly digested DNA was discarded. Aliquots of high-quality DNA were quantified using an

149 AccuClear Ultra high sensitivity dsDNA quantitation kit (Biotium, Fremont CA, USA) and a
150 SpectroMax M2 (Molecular Devices, Sunnyvale, CA, USA). Libraries were pooled with an

151 equal amount of DNA (ng/ μ l) contributed per individual to minimize individual contribution

152 bias, totaling 2000 ng of DNA per library. Number of individuals per samples are displayed in

153 *Fig 1*. No PCR was performed to ensure individual DNA contribution was kept equal across the

154 library (Anderson et al., 2014). The rest of the library preparation followed the ezRAD library

155 preparation protocol (Toonen et al. 2013; Knapp et al., 2016). This included DNA digested with

156 DpnII restriction enzyme, and adapters ligated using a Kapa hyper Prep Kit (Kapa Biosystems,

157 Wilmington, MA, USA). Pooled libraries were sequenced using Illumina MiSeq (v3 PE300bp) at

158 the Hawai'i Institute of Marine Biology EPSCoR Core sequencing facility.

159 ***Genetic analyses***

160 MultiQC was used to assess sequence quality scores, sequence length distributions, duplication

161 levels, and overrepresented sequences (Ewels et al., 2016). To analyze the mitochondrial

162 genome, a previously published mitochondrial genome from *Carcharhinus falciformis* was used

163 as a reference (GeneBank accession number KF801102). Raw paired-end reads were trimmed

164 with TRIMMOMATIC, mapped to the mitochondrial genome reference BWA (mem algorithm),
165 and variants called using the dDocent bioinformatics pipeline, modified for pool-seq (Puritz et al.
166 2014, see below for details). Called SNPs were then analyzed with AssessPool
167 (github.com/ToBoDev/assessPool, see below for details).

168

169 The bioinformatics **pipeline** utilized the dDocent followed by AssessPool. Given no reference
170 genome was available, a reference was constructed using the dDocent *de novo* assembly and
171 optimized utilizing the reference optimization steps provided on the dDocent assembly tutorial
172 (<http://ddocent.com/assembly/>). Before assembly reads were trimmed using default settings and
173 then an overlap (OL) assembly was performed, following by clustering with CD-HIT with a $-c$
174 parameter of 90% similarity. For mapping using BWA (mem algorithm) all match, mismatch,
175 and gap open penalty score parameters were also default settings. Different parameters were
176 tested during optimization but did not improve mapping. Within-pool (K1) and between-pool
177 (K2) minimum locus depth values selected for the *de novo* assembly did impact the results.
178 dDocent provides graphical outputs to help select these values; however, testing a few different
179 values of each is recommended to fully explore the potential of the data by balancing number of
180 contigs by coverage depth (see ddocent.com/UserGuide for details). Selected values for K1 and
181 K2 were 3 and 3 respectively. Once assembled, sequences were mapped, SNPs were called
182 within the dDocent pipeline using Freebayes, modified for SNP calling in pools (Garrison and
183 Marth 2012, <https://github.com/ekg/freebayes>). Any contigs that aligned to the mitochondrial
184 genome were removed from this nuclear dataset. The contigs that aligned specifically to the
185 mitochondrial control region were saved for SNP validation, to directly compare the results from
186 this pool-seq approach to those previously reported by Clarke et al. (2015).



187

188 SNP calling with FreeBayes was optimized for pooled samples using the ‘pooled-continuous’
189 option, and minor allele frequency was decreased to 0.05 to capture alleles with frequency
190 greater than 5% in the population (See Supplementary Material for code). The dDocent pipeline
191 outputs SNPs in two variant call format files (.vcf), one being all raw SNPs (TotalRawSNPs.vcf)
192 and another with filtered SNPs (Final.recode.vcf) however dDocent does not optimize filtering
193 for pool-seq data. Therefore, the raw SNPs were processed with the pool-seq specific program
194 AssessPool which uses VCFtools and vcflib to filter SNPs (Danecek et al., 2011). SNPs were
195 processed with the following filters: minimum pool number of 2, minimum quality score of 20,
196 minimum depth threshold of 30, maximum amount of missing data of 3, maximum allele length
197 of 10, quality score to depth ratio of 0.25 as well as mean depth per site vs. quality score, and
198 finally a maximum mean depth threshold of 1000 (*Table S1*). AssessPool then sends filtered
199 SNPs to either/or PoPoolation2 (Kofler et al., 2011) and poolfstat (Hivert et al., 2018).
200 Popoolation2 calculates mean pairwise F_{ST} values and significance in the form of p-values
201 obtained using Fisher’s exact test and combined using Fisher’s method (as described in Ryman et
202 al. 2006). Poolfstat (Hivert et al. 2018) takes a different approach to calculating F_{ST} values and in
203 detail in the discussion section. AssessPool then organizes, summarizes, and creates
204 visualizations of the data using RStudio (RStudio Team 2020).

205

206 As a quality control test, sequences from Clarke et al. (2015) were downloaded from GenBank
207 (accession numbers KM267565–KM267626), and SNPs from this data were compared directly
208 to SNPs called within the control region of the mitochondrial pool-seq data generated here.

209 Concordance of this validation set of SNPs was determined by Mantel test in R (Legendre &
210 Legendre, 1998) comparing the matrices of pairwise F_{ST} values among populations.

211 *Cost Analysis*

212 The cost of pool-seq approach compared to Sanger sequencing of individual loci was calculated
213 based on library preparation and sequencing cost at our facility. These costs were translated into
214 functions in RStudio (RStudio Team, 2020) where Sanger sequencing is a fixed rate per
215 individual and pool-seq costs per individual varies based on number of individuals and number
216 of pooled regions per sequencing lane. These functions were then plotted together for
217 comparison.

218

219 **Results**

220 A total of 30.8 million reads were generated for the four geographic regions, which averaged 7.7
221 \pm 3.0 million reads per pooled library. Results from the MutliQC assessment showed fairly
222 homogenous output between libraries in regard to sequence quality scores, GC and per base
223 sequence content, sequence length distributions, duplication levels, overrepresented sequences,
224 and adapter content. Once assembled, aligned, and mapped, 5,792 SNPs were resolved across the
225 mitochondrial and nuclear genomes combined. 4,103 were biallelic SNPs, 168 were multiallelic
226 SNPs and 48 were insertions and deletions (INDELs). INDELs and multiallelic SNPs remain a
227 challenge for quantification software, so we restricted our analysis to biallelic loci. AssessPool
228 creates visualizations of F_{ST} values and allows for visual outlier inspection. None appeared to be
229 present and given these SNPs are randomly chosen across the genome they are assumed to be
230 putatively neutral.

231

232 *Mitochondrial Genome*

233 Analysis of the complete mitochondrial genome (17,774 bp) revealed 804 variable sites (SNPs),
234 681 biallelic and 17 multiallelic. Since coverage in this dataset was fairly low on average, most
235 SNPs did not meet the filter threshold. After further filtering for the highest quality markers, 30
236 SNPs were selected to calculate allele frequencies. Pairwise F_{ST} values were all significant (*Fig*
237 *2, Table S2*). The Red Sea had much higher F_{ST} values (ranging from 0.367 to 0.745) than any
238 inter-Atlantic comparison (ranging from 0.029 to 0.135). However, all comparisons within the
239 Atlantic still showed significant F_{ST} values, the highest being between the North Atlantic and
240 Brazil, the lowest being Brazil and Gulf of Mexico (*Fig 2, Table S2*).

241

242 ***Nuclear loci***

243 Our nuclear data showed 4,988 variants of which 3,422 were biallelic SNPs and 151 were
244 multiallelic SNPs. A total of 346 SNPs remained after the same SNP filtering process was
245 applied as the mitochondrial genome. Nuclear markers showed lower F_{ST} values between
246 locations than the mitochondrial data, yet all P-values were still significant (*Fig 2, Table S2*).
247 The Red Sea showed consistently higher F_{ST} values in comparisons to inter-Atlantic comparisons
248 except for the North Atlantic to Gulf of Mexico comparison, which showed the second highest
249 mean F_{ST} value (*Fig 2, Table S2*). The highest value ($F_{ST} = 0.035$) was observed between Gulf of
250 Mexico and the Red Sea, whereas the lowest ($F_{ST} = 0.014$) between the North Atlantic and
251 Brazil, which had the highest F_{ST} value within the Atlantic for the mitochondrial data.

252

253 ***SNP validation***

254 SNPs called in the mitochondrial control region using the pool-seq protocol were compared with
255 those reported in Clarke et al. (2015). Of the 34 SNPs in their study 14 of them had minor allele

256 count (MAC) of less than or equal to 3 and several were singletons. These SNPs are removed
257 from the pool-seq data due to MAC SNP filter of >3 to remove possible sequencing errors during
258 high-throughput sequencing. Therefore, rare alleles like singletons or any with MAC less than or
259 equal to 3 will inherently be removed from pool-seq data sets. Three SNPs were found in the
260 Clarke study with a MAC of >3 that were not present in the pool-seq data; however, the remaining
261 17 SNPs were all present in our data, plus one that was not found in the Clarke study (*Fig S1*).
262 Despite the loss of these rare alleles from the SNP validation set, pairwise F_{ST} values estimated
263 by both methods remained highly correlated (Mantel test, $r^2 = 0.96$, $p < 0.05$), and comparisons
264 between the Red Sea and all three Atlantic populations showed the same relative magnitude
265 between both methods.

266 **Cost Analysis**

267 The findings for cost analysis indicate that pool-seq reaches a threshold at approximately 300
268 individuals, after which this approach offers cheaper results than individual Sanger sequences.
269 Furthermore, the cost is only twice as expensive at just over 100 individuals (*Fig 3a*). The pool-
270 seq approach provides a far higher ratio of information for the cost, yielding greater population
271 resolution. **The cost does not include analytical time and effort associated with pool-seq;**
272 however, these costs are likely to decrease in the near future as bioinformatic pipelines are
273 improved and become more widely available, for example as applications deployed via cloud
274 based platforms such as Galaxy (<https://usegalaxy.org/>) or CyVerse (<https://cyverse.org/>). The
275 choice of pool-seq methodology has many caveats, which are discussed in greater detail in the
276 ‘considerations on pool-seq’ section of the discussion below.

277 **Discussion**

278 Elasmobranchs are being harvested at unsustainable levels in several commercial fishing
279 industries around the world. A fundamental step in successful management of any species is
280 resolving population boundaries so they can be managed on a genetic stock by stock basis. As
281 genetic sequencing technologies advance, there is greater opportunity to detect even small-scale
282 genetic differences between populations, indicating distinct stocks.



283
284 Here, we validate the utility of pool-seq using the same individuals as a previous study (Clarke et
285 al. 2015) and show that pool-seq recovers additional population structure relative to Sanger
286 sequencing of the mtDNA control region. Pool-seq was able to detect isolated populations within
287 the Gulf of Mexico, in the Western Atlantic, and along the Brazilian coast, where Clarke et al.
288 (2015) found no population structure. As expected, the Red Sea population was highly isolated
289 from Atlantic conspecifics.

290
291 One advantage of this pool-seq approach is that we recover SNPs through-out the entire
292 mitochondrial genome along with thousands of additional nuclear loci that provide higher
293 resolution and detect finer scale population structure. This approach yielded significant genetic
294 structure among inter-Atlantic regions in both mtDNA and nuclear loci, where Sanger
295 sequencing of the mtCR was unable to resolve significant differences among the same
296 populations. The concurrence between the mitochondrial genome and nuclear loci reinforces the
297 conclusion of population structure among all regions sampled in this study.



298
299 In this case, pool-seq lived up to the promise of increased power to detect fine-scale structure,
300 but does it live up to the promise of being cost-effective (Ferretti et al., 2013; Schlötterer et al.,

2014)? Individual extraction costs remain fixed across both approaches and Sanger sequencing generally has a flat rate per individual, including PCR primers and reagents, and sequencing per individual per locus. In contrast, pool-seq has a flat sequencing cost determined by the number of reads generated from the high-throughput sequencing platform, plus a small additional cost per pool and the quantification of DNA. Comparing costs at our institution between a single Sanger sequencing marker and pool-Seq on the Illumina MiSeq platform indicates pool-seq is less expensive if sample size of the study rises above 300 individuals. When higher numbers of individuals per pool are applied, the cost per library preparation is lower and price per individual analyzed is further reduced (*Fig 3b*). This assessment is limited to 12 pools due to the maximum number of reads per lane produced on the MiSeq platform. Therefore, analyzing more than 12 pools on a MiSeq would require additional sequencing lanes and increasing the cost per individual/pool. These methods can be applied to HiSeq or NovaSeq platforms which have higher sequencing capacities per lane and can further increase cost efficiencies. These two platforms can also be run with individual barcodes therefore gaining the individual information lost by pooling specimens. However, read lengths will be half as long as MiSeq which could affect de novo assemblies. This should be considered when designing high-throughput sequencing projects. All of these pool-seq price comparisons are to a single Sanger-sequenced marker, so when considering the information acquired from pool-seq compared to the cost from traditional single mitochondrial marker the price per marker advantage is massively amplified.

320

321 ***Considerations with Pool-seq***

322 As with any sequencing technique, there are still several factors to consider before deciding if
323 pool-seq is appropriate for a particular study. Multiple reviews have been published on high-

324 throughput and pool-seq approaches demonstrating pros, cons, and considerations with these
325 methods, which are beyond the scope of this study. Interested readers should consult Perez-
326 Enciso & Ferretti (2010), Futschik & Schlötterer (2010), Kofler, Betancourt, & Schlötterer
327 (2012), Ferretti et al. (2013), Schlötterer et al., 2014, Andrews & Luikart (2014), and Andrews
328 et al. (2016).

329

330 Pooling assumes individuals are from closely related populations within the same species.

331 Therefore, care needs to be taken to avoid cryptic species, genetic structure over habitats, or

332 other unintentional bias when selecting individuals to pool. For wide ranging pelagic species

333 such as the blue shark or oceanic whitetip it is reasonable to pool individuals from a larger area

334 than it would be for small benthic species such as horn sharks, wobbegongs, or most rays.

335 Population structure may be obscured if the geographic range per pool is too large or if there is

336 complex population structure (*sensu* Bowen et al. 2005), because individuals from multiple sub-

337 populations will be mixed into a single pool from which allele frequencies are calculated.

338 Certainly pool-seq is not appropriate in all cases. It is a cost-saving approach for analyses based

339 on allele frequencies only, because individual information is lost by pooling, including

340 haplotypes/genotypes and linkage disequilibrium information. Also, pooling makes it difficult to

341 distinguish between low frequency alleles in the population and sequencing error. Therefore,

342 careful filtering must be applied to assure valid SNPs are analyzed instead of sequencing noise

343 (Anand et al., 2016; Schlötterer et al., 2014). Finally, the estimation of F_{ST} from pooled data

344 remains a subject of some debate, and new approaches and bias corrections are being actively

345 developed (Kofler et al. 2011; Hivert et al. 2018). The R package poolfstat (Hivert et al., 2018) is

346 based on an analysis-of-variance framework to eliminate biases associated with varying pool



347 sizes. Re-analyzing our data with poolstat yields slightly different F_{ST} values than reported here
348 (see *Table S1*). However, a comparison of F_{ST} matrices produced among the two approaches
349 shows strong correlation (Mantel $r=0.991$ for mitochondrial and $r=0.978$ for nuclear data, $p <$
350 0.05). Therefore, only F_{ST} values calculated by Popoolation were reported here.

351

352 **Conclusions**

353 The finding of population structure on a scale of North Atlantic/Gulf of Mexico/Brazil is nearly
354 unprecedented for a pelagic shark. Population structure in globally distributed sharks is typically
355 detected on a scale of ocean basins (Atlantic versus Indo-Pacific, Castro et al. 2007; Graves &
356 McDowell, 2015) and a few pelagic fishes have no population structure on a global scale (e.g.
357 Basking shark, *Cetorhinus maximus*, Hoelzel et al. 2006; Blue shark *Prionace glauca*, Verissimo
358 et al. 2017; Wahoo, *Acanthocybium solandri*, Theisen et al. 2008). The resolution of isolated
359 populations on the scale of North Atlantic Ocean is more typical of coastal species than pelagic
360 species. The silky shark seems to be a pelagic species with a somewhat coastal population
361 structure. This has strong implications for international management, as smaller stocks imply
362 smaller populations, which are more readily depleted. At a minimum, this data requires
363 rethinking a single population management approach for the Atlantic, and this pattern needs to
364 be investigated for this species across the Indo-Pacific as well.

365

366 Overall this study demonstrates pool-seq is a strong tool for analyzing large portions of the
367 genome which the methods traditionally used for elasmobranchs could not supply. Sharks and
368 rays are an imperiled group of species that could benefit from advanced genomic studies to
369 outline appropriate management units. Finally, the cost of pool-seq is a flat rate with minimal
370 cost additions per individual, where the cost of Sanger increases more per individual and can

371 eventually become more expensive than pool-seq. The cost per genetic marker however makes
372 pool-seq dramatically cheaper, especially when hundreds of individuals are assessed.

373

374 **Acknowledgements**

375 This study was made possible by the generous donation of specimens by Christopher R. Clarke,

376 Mahmood Shivji, Stephen A. Karl, J.D. Filmlalter, and Julia Spaet. We thank members of the

377 ToBo Lab for sharing expertise, advice and discussions that contributed to this manuscript.

378 Special thanks to Darren Lerner, Kim Holland, Carl Meyer, S. Gulak, D. Bethe, D. McCauley, C.

379 Wilson. Guy Harvey Ocean Foundation, and Save Our Seas Foundation. This is contribution

380 #XXXX from the Hawaii Institute of Marine Biology, contribution #XXXX from the Hawaii Sea

381 Grant Program, and contribution #XXXX from the School of Ocean and Earth Science and

382 Technology at the University of Hawaii.

383 **Author Contributions:**

384 Derek W. Kraft – Designed research, performed research, analyzed data, wrote paper, provided
385 funding for research.

386 Emily Conklin – Created bioinformatics pipeline for data analysis and assisted in data analysis

387 Evan Barba – Created bioinformatics pipeline for data analysis and assisted in data analysis

388 Melanie Hutchinson – Helped obtain samples and provided biological insight

389 Robert J. Toonen – Contributed funding, contributed to research design, and contributed to data
390 analysis

391 Zac H. Forsman – Contributed to research design and data analysis

392 Brian W. Bowen – Contributed funding, contributed to research design, and edited paper

393

394 **References**

- 395 Anand, S., Mangano, E., Barizzone, N., Bordoni, R., Sorosina, M., Clarelli, F., Corrado, L.,
396 Boneschi, F. M., D'Alfonso, S., & De Bellis, G. (2016). Next generation sequencing of
397 pooled samples: Guideline for variants' filtering. *Scientific Reports*, 6(August), 1–9.
398 <https://doi.org/10.1038/srep33735>
- 399 Anderson, E. C., Skaug, H. J., & Barshis, D. J. (2014). Next-generation sequencing for molecular
400 ecology: A caveat regarding pooled samples. In *Molecular Ecology* (Vol. 23, Issue 3, pp.
401 502–512). <https://doi.org/10.1111/mec.12609>
- 402 Andrews, K.R., & Luikart, G. (2014). Recent novel approaches for population genomics data
403 analysis. *Molecular Ecology*, 23(7), 1661–1667.
- 404 Andrews, Kimberly R., Good, J. M., Miller, M. R., Luikart, G., & Hohenlohe, P. A. (2016).
405 Harnessing the power of RADseq for ecological and evolutionary genomics. *Nature*
406 *Reviews Genetics*, 17(2), 81–92. <https://doi.org/10.1038/nrg.2015.28>
- 407 Avise, J. C., Bowen, B. W., Bermingham, E., Meylan, A. B., & Lamb, T. (1992). Mitochondrial
408 DNA evolution at a turtle's pace: evidence for low genetic variability and reduced
409 microevolutionary rate in the Testudines. *Molecular Biology & Evolution*, 9, 457–473.
- 410 Barrio, A. M., Lamichhaney, S., Fan, G., Rafati, N., Pettersson, M., Zhang, H., Dainat, J.,
411 Ekman, D., Höppner, M., Jern, P., Martin, M., Nystedt, B., Liu, X., Chen, W., Liang, X.,
412 Shi, C., Fu, Y., Ma, K., Zhan, X., ... Andersson, L. (2016). The genetic basis for ecological
413 adaptation of the Atlantic herring revealed by genome sequencing. *ELife*, 5(MAY2016), 1–
414 32. <https://doi.org/10.7554/eLife.12081>
- 415 Baum, J. K., Myers, R. A., Kehler, D. G., Worm, B., Harley, S. J., & Doherty, P. A. (2003).
416 Collapse and conservation of shark populations in the Northwest Atlantic. *Science*,
417 299(5605), 389–392. <https://doi.org/10.1126/science.1079777>

- 418 Bernard, A. M., Horn, R. L., Chapman, D. D., Feldheim, K. A., Garla, R. C., Brooks, E. J., Gore,
419 M. A., & Shivji, M. S. (2017). Genetic connectivity of a coral reef ecosystem predator: the
420 population genetic structure and evolutionary history of the Caribbean reef shark
421 (*Carcharhinus perezi*). *Journal of Biogeography*, *44*(11), 2488–2500.
422 <https://doi.org/10.1111/jbi.13062>
- 423 Bowen, B. W., Bass, A. L., Soares, L., & Toonen, R. J. (2005). Conservation implications of
424 complex population structure : lessons from the loggerhead turtle (*Caretta caretta*).
425 *Molecular Ecology*, *14*, 2389–2402. <https://doi.org/10.1111/j.1365-294X.2005.02598.x>
- 426 Brown, W. M., George, M., & Wilson, A. C. (1979). Rapid evolution of animal mitochondrial
427 DNA. *Proceedings of the National Academy of Sciences of the United States of America*,
428 *76*(4), 1967–1971. <https://doi.org/10.1146/annurev.es.18.110187.001413>
- 429 Cardeñosa, D., Fields, A. T., Babcock, E. A., Zhang, H., Feldheim, K., Shea, S. K. H., Fischer,
430 G. A., & Chapman, D. D. (2018). CITES-listed sharks remain among the top species in the
431 contemporary fin trade. *Conservation Letters*, *11*(4), 1–7.
432 <https://doi.org/10.1111/conl.12457>
- 433 Carvalho, G. R., & Hauser, L. (1994). Molecular genetics and the stock concept in fisheries.
434 *Reviews in Fish Biology and Fisheries*, *4*(3), 326–350. <https://doi.org/10.1007/BF00042908>
- 435 Castro, A. L. F., Stewart, B. S., Wilson, S. G., Hueter, R. E., Meekan, M. G., Motta, P. J.,
436 Bowen, B. W., & Karl, S. A. (2007). Population genetic structure of Earth’s largest fish, the
437 whale shark (*Rhincodon typus*). *Molecular Ecology*, *16*(24), 5183–5192.
438 <https://doi.org/10.1111/j.1365-294X.2007.03597.x>
- 439 Charlesworth, D., & Wright, S. I. (2001). Breeding systems and genome evolution. *Current*
440 *Opinion in Genetics & Development*, *11*(6), 685–690. [PeerJ reviewing PDF | \(2020:07:51011:0:1:NEW 22 Jul 2020\)](https://doi.org/10.1016/S0959-</p></div><div data-bbox=)

- 441 437X(00)00254-9
- 442 Clarke, C. R., Karl, S. A., Horn, R. L., Bernard, A. M., Lea, J. S., Hazin, F. H., Prodöhl, P. A., &
443 Shivji, M. S. (2015). Global mitochondrial DNA phylogeography and population structure
444 of the silky shark, *Carcharhinus falciformis*. *Marine Biology*, *162*(5), 945–955.
445 <https://doi.org/10.1007/s00227-015-2636-6>
- 446 Clarke, S. C., McAllister, M. K., Milner-Gulland, E. J., Kirkwood, G. P., Michielsens, C. G. J.,
447 Agnew, D. J., Pikitch, E. K., Nakano, H., & Shivji, M. S. (2006). Global estimates of shark
448 catches using trade records from commercial markets. *Ecology Letters*, *9*(10), 1115–1126.
449 <https://doi.org/10.1111/j.1461-0248.2006.00968.x>
- 450 Daly-Engel, T. S., Seraphin, K. D., Holland, K. N., Coffey, J. P., Nance, H. A., Toonen, R. J., &
451 Bowen, B. W. (2012). Global phylogeography with mixed-marker analysis reveals male-
452 mediated dispersal in the endangered scalloped hammerhead shark (*Sphyrna lewini*). *PLoS*
453 *ONE*, *7*(1), e29986. <https://doi.org/10.1371/journal.pone.0029986>
- 454 Danecek, P., Auton, A., Abecasis, G., Albers, C. A., Banks, E., DePristo, M. A., Handsaker, R.
455 E., Lunter, G., Marth, G. T., Sherry, S. T., McVean, G., & Durbin, R. (2011). *The variant*
456 *call format and VCFtools*. *27*(15), 2156–2158.
457 <https://doi.org/10.1093/bioinformatics/btr330>
- 458 DiBattista, J. D., Waldrop, E., Rocha, L. A., Craig, M. T., Beruman, M. L., & Bowen, B. W.
459 (2015). Blinded by the bright: A lack of congruence between color morphs,
460 phylogeography, and taxonomy in a cosmopolitan Indo-Pacific butterflyfish, *Chaetodon*
461 *auriga*. *Journal of Biogeography*, *42*, 1919–1929.
- 462 Dizon, A. E., Lockyer, C., Perrin, W. F., Demaster, D. P., & Sisson, J. (1993). Rethinking the
463 stock concept: a phylogeographic approach. *Biological Conservation*, *64*(2), 176–177.

- 464 [https://doi.org/10.1016/0006-3207\(93\)90670-V](https://doi.org/10.1016/0006-3207(93)90670-V)
- 465 Domingues, Rodrigo R., Bruels, C. C., Gadig, O. B. F., Chapman, D. D., Hilsdorf, A. W. S., &
466 Shivji, M. S. (2018). Genetic connectivity and phylogeography of the night shark
467 (*Carcharhinus signatus*) in the western Atlantic Ocean: Implications for conservation
468 management. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 29(1), 102–114.
469 <https://doi.org/10.1002/aqc.2961>
- 470 Domingues, Rodrigo R., Hilsdorf, A. W. S., Shivji, M. M., Hazin, F. V. H., & Gadig, O. B. F.
471 (2017). Effects of the Pleistocene on the mitochondrial population genetic structure and
472 demographic history of the silky shark (*Carcharhinus falciformis*) in the western Atlantic
473 Ocean. *Reviews in Fish Biology and Fisheries*, 1–15. [https://doi.org/10.1007/s11160-017-](https://doi.org/10.1007/s11160-017-9504-z)
474 9504-z
- 475 Domingues, Rodrigo Rodrigues, Hilsdorf, A. W. S., & Gadig, O. B. F. (2018). The importance
476 of considering genetic diversity in shark and ray conservation policies. *Conservation*
477 *Genetics*, 19, 501–525. <https://doi.org/10.1007/s10592-017-1038-3>
- 478 Dulvy, N. K., Baum, J. K., Clarke, S., Compagno, L. J. V., Cortés, E., Domingo, A., Fordham,
479 S., Fowler, S., Francis, M. P., Gibson, C., Martínez, J., Musick, J. A., Soldo, A., Stevens, J.
480 D., & Valenti, S. (2008). You can swim but you can't hide: The global status and
481 conservation of oceanic pelagic sharks and rays. *Aquatic Conservation: Marine and*
482 *Freshwater Ecosystems*, 18(5), 459–482. <https://doi.org/10.1002/aqc.975>
- 483 Dulvy, N. K., Fowler, S. L., Musick, J. A., Cavanagh, R. D., Kyne, P. M., Harrison, L. R.,
484 Carlson, J. K., Davidson, L. N. K., Fordham, S. V., Francis, M. P., Pollock, C. M.,
485 Simpfendorfer, C. A., Burgess, G. H., Carpenter, K. E., Compagno, L. J. V., Ebert, D. A.,
486 Gibson, C., Heupel, M. R., Livingstone, S. R., ... White, W. T. (2014). Extinction risk and

- 487 conservation of the world's sharks and rays. *ELife*, 2014(3), 1–34.
488 <https://doi.org/10.7554/eLife.00590.001>
- 489 Dulvy, N. K., & Trebilco, R. (2018). Size-Based Insights into the Ecosystem Role of Sharks and
490 Rays. In *Shark Research: Emerging Technologies and Applications for the Field and*
491 *Laboratory*.
- 492 Duncan, K. M., Martin, A. P., Bowen, B. W., & De Couet, H. G. (2006). Global phylogeography
493 of the scalloped hammerhead shark (*Sphyrna lewini*). *Molecular Ecology*, 15(8), 2239–
494 2251. <https://doi.org/10.1111/j.1365-294X.2006.02933.x>
- 495 Ewels, P., Lundin, S., & Max, K. (2016). Data and text mining MultiQC : summarize analysis
496 results for multiple tools and samples in a single report. *Bioinformatics*, 32(19), 3047–3048.
497 <https://doi.org/10.1093/bioinformatics/btw354>
- 498 Ferretti, F., Worm, B., Britten, G. L., Heithaus, M. R., & Lotze, H. K. (2010). Patterns and
499 ecosystem consequences of shark declines in the ocean. In *Ecology Letters* (Vol. 13, Issue
500 8, pp. 1055–1071). <https://doi.org/10.1111/j.1461-0248.2010.01489.x>
- 501 Ferretti, L., Ramos-Onsins, S. E., & Pérez-Enciso, M. (2013). Population genomics from pool
502 sequencing. *Molecular Ecology*, 22(22), 5561–5576. <https://doi.org/10.1111/mec.12522>
- 503 Fischer, M. C., Rellstab, C., Tedder, A., Zoller, S., Gugerli, F., Shimizu, K. K., Holderegger, R.,
504 & Widmer, A. (2013). Population genomic footprints of selection and associations with
505 climate in natural populations of *Arabidopsis halleri* from the Alps. *Molecular Ecology*,
506 22(22), 5594–5607. <https://doi.org/10.1111/mec.12521>
- 507 Futschik, A., & Schlötterer, C. (2010). The next generation of molecular markers from massively
508 parallel sequencing of pooled DNA samples. *Genetics*, 186(1), 207–218.
509 <https://doi.org/10.1534/genetics.110.114397>

- 510 Garrison, E., & Marth, G. (2012). *Haplotype-based variant detection from short-read*
511 *sequencing*. 1–9.
- 512 Gautier, M., Foucaud, J., Gharbi, K., Cézard, T., Galan, M., Loiseau, A., Thomson, M., Pudlo,
513 P., Kerdelhué, C., & Estoup, A. (2013). Estimation of population allele frequencies from
514 next-generation sequencing data: Pool-versus individual-based genotyping. *Molecular*
515 *Ecology*, 22(14), 3766–3779. <https://doi.org/10.1111/mec.12360>
- 516 Green, M. E., Appleyard, S. A., White, W., Tracey, S., Devloo-Delva, F., & Oviden, J. R.
517 (2019). Novel multimarker comparisons address the genetic population structure of silvertip
518 sharks (*Carcharhinus albimarginatus*). *Marine and Freshwater Research*, 70(7), 1007–1019.
519 <https://doi.org/10.1071/MF18296>
- 520 Guo, B., Li, Z., & Merilä, J. (2016). Population genomic evidence for adaptive differentiation in
521 the Baltic Sea herring. *Molecular Ecology*, 25(12), 2833–2852.
522 <https://doi.org/10.1111/mec.13657>
- 523 Heist, E. J. (2004). *Biology of Sharks and Their Relatives* (J. C. Carrier, J. A. Musick, & M. .
524 Heithaus (eds.)). CRC Press, Boca Raton, FL.
- 525 Heithaus, M. R., Frid, A., Wirsing, A. J., & Worm, B. (2008). Predicting ecological
526 consequences of marine top predator declines. *Trends in Ecology and Evolution*, 23(4),
527 202–210. <https://doi.org/10.1016/j.tree.2008.01.003>
- 528 Heupel, M. R., Knip, D. M., Simpfendorfer, C. A., & Dulvy, N. K. (2014). Sizing up the
529 ecological role of sharks as predators. *Marine Ecology Progress Series*, 495, 291–298.
530 <https://doi.org/10.3354/meps10597>
- 531 Hivert, V., Leblois, R., Petit, E. J., Gautier, M., & Vitalis, R. (2018). Measuring genetic
532 differentiation from Pool-seq data. *Genetics*, 210(September), 315–330.

- 533 <https://doi.org/10.1534/genetics.118.300900>
- 534 Hoelzel, A. R., Shivji, M. S., Magnussen, J. E., & Francis, M. P. (2006). Low worldwide genetic
535 diversity in the basking shark (*Cetorhinus maximus*). *Biology Letters*, 2, 639–642.
536 <https://doi.org/10.1098/rsbl.2006.0513>
- 537 Hohenlohe, P. A., Hand, B. K., Andrews, K. R., & Luikart, G. (2018). Population genomics
538 provides key insights in ecology and evolution. *Population Genomics: Concepts,*
539 *Approaches and Applications*, doi:10.1007/13836_2018_20.
540 https://doi.org/10.1007/13836_2018_20
- 541 IUCN, I. U. for the C. of N. (2017). *IUCN Red List of Threatened Species*. Version 2013.1.
542 <http://www.iucnredlist.org/about/overview>
- 543 Jones, F. C., Grabherr, M. G., Chan, Y. F., Russell, P., Mauceli, E., Johnson, J., Swofford, R.,
544 Pirun, M., Zody, M. C., White, S., Birney, E., Searle, S., Schmutz, J., Grimwood, J.,
545 Dickson, M. C., Myers, R. M., Miller, C. T., Summers, B. R., Knecht, A. K., ... Kingsley,
546 D. M. (2012). The genomic basis of adaptive evolution in threespine sticklebacks. *Nature*,
547 484(7392), 55–61. <https://doi.org/10.1038/nature10944>
- 548 Karl, S. A., Castro, A. L. F., Lopez, J. A., Charvet, P., & Burgess, G. H. (2011). Phylogeography
549 and conservation of the bull shark (*Carcharhinus leucas*) inferred from mitochondrial and
550 microsatellite DNA. *Conservation Genetics*, 12(2), 371–382.
551 <https://doi.org/10.1007/s10592-010-0145-1>
- 552 Karl, S. a, Bowen, B. W., & Avise, J. C. (1992). Global population genetic structure and male-
553 mediated gene flow in the green turtle (*Chelonia mydas*): RFLP analysis of anonymous
554 nuclear loci. *Genetics*, 131, 163–173. <https://doi.org/10.1139/z05-185>
- 555 Keeney, D. B., & Heist, E. J. (2006). Worldwide phylogeography of the blacktip shark

556 (Carcharhinus limbatus) inferred from mitochondrial DNA reveals isolation of western
557 Atlantic populations coupled with recent Pacific dispersal. *Molecular Ecology*, 15, 3669–
558 3679.

559 Knapp, I. S. S., Puritz, J. B., Bird, C. E., Whitney, J. L., Sudek, M., Forsman, Z. H., & Toonen,
560 R. J. (2016). *ezRAD- an accessible next-generation RAD sequencing protocol suitable for*
561 *non-model organisms_v3.2*. <https://doi.org/10.17504/protocols.io.e9pbh5n>

562 Kofler, R., Betancourt, A. J., & Schlötterer, C. (2012). Sequencing of pooled DNA samples
563 (Pool-Seq) uncovers complex dynamics of transposable element insertions in *Drosophila*
564 *melanogaster*. *PLoS Genetics*, 8(1). <https://doi.org/10.1371/journal.pgen.1002487>

565 Kofler, R., Pandey, R. V., & Schlötterer, C. (2011). PoPoolation2: Identifying differentiation
566 between populations using sequencing of pooled DNA samples (Pool-Seq). *Bioinformatics*,
567 27(24), 3435–3436. <https://doi.org/10.1093/bioinformatics/btr589>

568 Legendre, P., & Legendre, L. (1998). *Numerical Ecology. 2nd English Edition*.

569 Martin, A. P., Naylor, G., & Palumbi, S. R. (1992). Rates of mitochondrial DNA evolution in
570 sharks are slow compared to mammals. *Nature*, 357, 153–155.
571 <https://doi.org/10.1038/357153a0>

572 Mimee, B., Duceppe, M. O., Véronneau, P. Y., Lafond-Lapalme, J., Jean, M., Belzile, F., &
573 Bélair, G. (2015). A new method for studying population genetics of cyst nematodes based
574 on Pool-Seq and genomewide allele frequency analysis. *Molecular Ecology Resources*,
575 15(6), 1356–1365. <https://doi.org/10.1111/1755-0998.12412>

576 Momigliano, P., Harcourt, R., Robbins, W. D., Jaiteh, V., Mahardika, G. N., Sembiring, A., &
577 Stow, A. (2017). Genetic structure and signatures of selection in grey reef sharks
578 (*Carcharhinus amblyrhynchos*). *Heredity*, 119(3), 142–153.

- 579 <https://doi.org/10.1038/hdy.2017.21>
- 580 Musick, J. A., Burgess, G., Cailliet, G., Camhi, M., & Fordham, S. (2000). Management of
581 Sharks and Their Relatives (Elasmobranchii). *Fisheries*, 25(3), 9–13.
582 [https://doi.org/10.1577/1548-8446\(2000\)025<0009:MOSATR>2.0.CO;2](https://doi.org/10.1577/1548-8446(2000)025<0009:MOSATR>2.0.CO;2)
- 583 Myers, R. A., Baum, J. K., Shepherd, T. D., Powers, S. P., & Peterson, C. H. (2007). Cascading
584 effects of the loss of apex predatory sharks from a coastal ocean. *Science*, 315(5820), 1846–
585 1850. <https://doi.org/10.1126/science.1138657>
- 586 Neiman, M., & Taylor, D. R. (2009). The causes of mutation accumulation in mitochondrial
587 genomes. *Proceedings of the Royal Society B: Biological Sciences*, 276(1660), 1201–1209.
588 <https://doi.org/10.1098/rspb.2008.1758>
- 589 Oliver, S., Braccini, M., Newman, S. J., & Harvey, E. S. (2015). Global patterns in the bycatch
590 of sharks and rays. *Marine Policy*, 54, 86–97. <https://doi.org/10.1016/j.marpol.2014.12.017>
- 591 Ovenden, J. R., Berry, O., Welch, D. J., Buckworth, R. C., & Dichmont, C. M. (2015). Ocean 's
592 eleven : a critical evaluation of the role of population , evolutionary and molecular genetics
593 in the management of wild fi sheries. *Fish Fisheries*, 16(125–159).
594 <https://doi.org/10.1111/faf.12052>
- 595 Pardini, A. T., Jones, C. S., Noble, L. R., Kreiser, B., & Malcolm, H. (2001). Sex-biased
596 dispersal of great white sharks. *Nature*, 412, 139–140.
- 597 Pazmiño, D. A., Maes, G. E., Green, M. E., Simpfendorfer, C. A., Hoyos-Padilla, E. M., Duffy,
598 C. J. A., Meyer, C. G., Kerwath, S. E., Salinas-De-León, P., & Van Herwerden, L. (2018).
599 Strong trans-Pacific break and local conservation units in the Galapagos shark
600 (*Carcharhinus galapagensis*) revealed by genome-wide cytonuclear markers. *Heredity*,
601 120(5), 407–421. <https://doi.org/10.1038/s41437-017-0025-2>

- 602 Perez-Enciso, M., & Ferretti, L. (2010). Massive parallel sequencing in animal genetics :
603 wherefroms and wheretos. *Animal Genetics*, *41*, 561–569. <https://doi.org/10.1111/j.1365-2052.2010.02057.x>
- 604
- 605 Portnoy, D. S., McDowell, J. R., Heist, E. J., Musick, J. A., & Graves, J. E. (2010). World
606 phylogeography and male-mediated gene flow in the sandbar shark, *Carcharhinus*
607 *plumbeus*. *Molecular Ecology*, *19*, 1994–2010.
- 608 Puritz, J. B., Hollenbeck, C. M., & Gold, J. R. (2014). *dDocent* : a RADseq, variant-calling
609 pipeline designed for population genomics of non-model organisms. *PeerJ*, *2*, e431.
610 <https://doi.org/10.7717/peerj.431>
- 611 Rice, J., & Harley, S. (2013). Updated stock assessment of the Silky sharks in the Western and
612 Central Pacific Ocean. *WCPFC-SC9-2013/ SA-WP-03, August*.
- 613 RStudio Team (2020). RStudio: Integrated Development Environment for R. RStudio, PBC,
614 Boston, MA
615 URL <http://www.rstudio.com/>
- 616 Ryman, N., Palm, S., André, C., Carvalho, G. R., Dahlgren, T. G., Jorde, P. E., Laikre, L.,
617 Larsson, L. C., Palmé, A., & Ruzzante, D. E. (2006). Power for detecting genetic
618 divergence: Differences between statistical methods and marker loci. *Molecular Ecology*,
619 *15*(8), 2031–2045. <https://doi.org/10.1111/j.1365-294X.2006.02839.x>
- 620 Schlötterer, C., Tobler, R., Kofler, R., & Nolte, V. (2014). Sequencing pools of individuals-
621 mining genome-wide polymorphism data without big funding. *Nature Reviews Genetics*,
622 *15*(11), 749–763. <https://doi.org/10.1038/nrg3803>
- 623 Schultz, J. K., Feldheim, K. A., Gruber, S. H., Ashley, M. V. H., McGovern, T. M., Bowen, B.
624 W., Bruber, S. H., Ashley, M. V. H., McGovern, T. M., & Bowen, B. W. (2008). Global

625 phylogeography and seascape genetics of the lemon sharks (genus *Negaprion*). *Molecular*
626 *Ecology*, 17(24), 5336–5348. <https://doi.org/10.1111/j.1365-294X.2008.04000.x>

627 Stevens, J. D., Bonfil, R., Dulvy, N. K., & Walker, P. A. (2000). The effects of fishing on sharks,
628 rays, and chimaeras (chondrichthyans), and the implications for marine ecosystems. *ICES*
629 *Journal of Marine Science*, 57(3), 476–494. <https://doi.org/10.1006/jmsc.2000.0724>

630 Tallmon, D. A., Gregovich, D., Waples, R., & Baker, C. S. (2010). When are genetic methods
631 useful for estimating contemporary abundance and detecting population trends? *Molecular*
632 *Ecology Resources*, 10, 684–692.

633 Theisen, T. C., Bowen, B. W., Lanier, W., & Baldwin, J. D. (2008). High connectivity on a
634 global scale in the pelagic wahoo, *Acanthocybium solandri* (tuna family Scombridae).
635 *Molecular Ecology*, 17, 4233–4247.

636 Toonen, R. J., Puritz, J. B., Forsman, Z. H., Whitney, J. L., Fernandez-Silva, I., Andrews, K. R.,
637 & Bird, C. E. (2013). ezRAD: a simplified method for genomic genotyping in non-model
638 organisms. *PeerJ*, 1, e203. <https://doi.org/10.7717/peerj.203>

639 Whitney, N. M., Robbins, W. D., Schultz, J. K., Bowen, B. W., & Holland, K. N. (2012).
640 Phylogeography of the whitetip reef shark (*Triaenodon obesus*): a sedentary species with a
641 broad distribution. *Journal of Biogeography*, 39, 1144–1156.

642 Zink, R.M., Barrowclough, G. (2008). Mitochondrial DNA under siege in avian
643 phylogeography. *Molecular Ecology*, 17, 2107–2121.

644

Figure 1

Sample locations of *Carcharhinus falciformis* followed by sample size

Abbreviations: GM = Gulf of Mexico, NA = North Atlantic, BR = Brazil, RS = Red Sea.

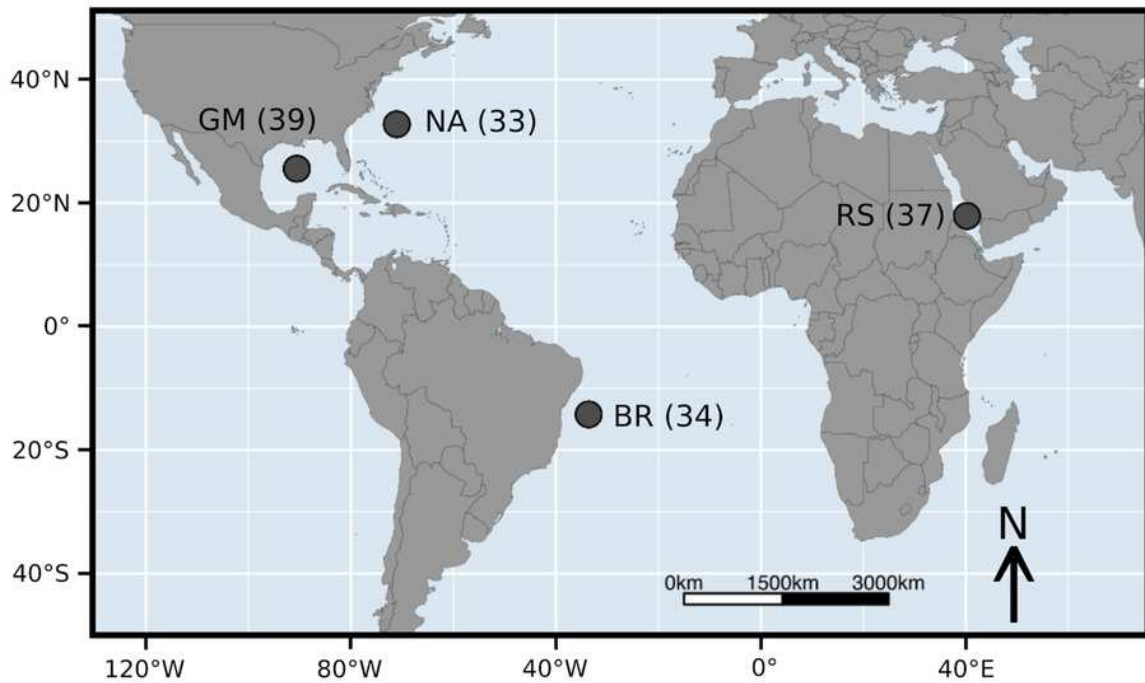


Figure 2

Pairwise F_{ST} values generated by Pool-seq methods

Cool colors (top left) are F_{ST} values calculated from nuclear genome loci, warm colors (bottom right) are F_{ST} values from loci across the entire mitochondrial genome. All pairwise differences are significant ($p < 0.001$). Regional abbreviations are as follows; GM = Gulf of Mexico, BR = Brazil, NA = North Atlantic, RD = Red sea.

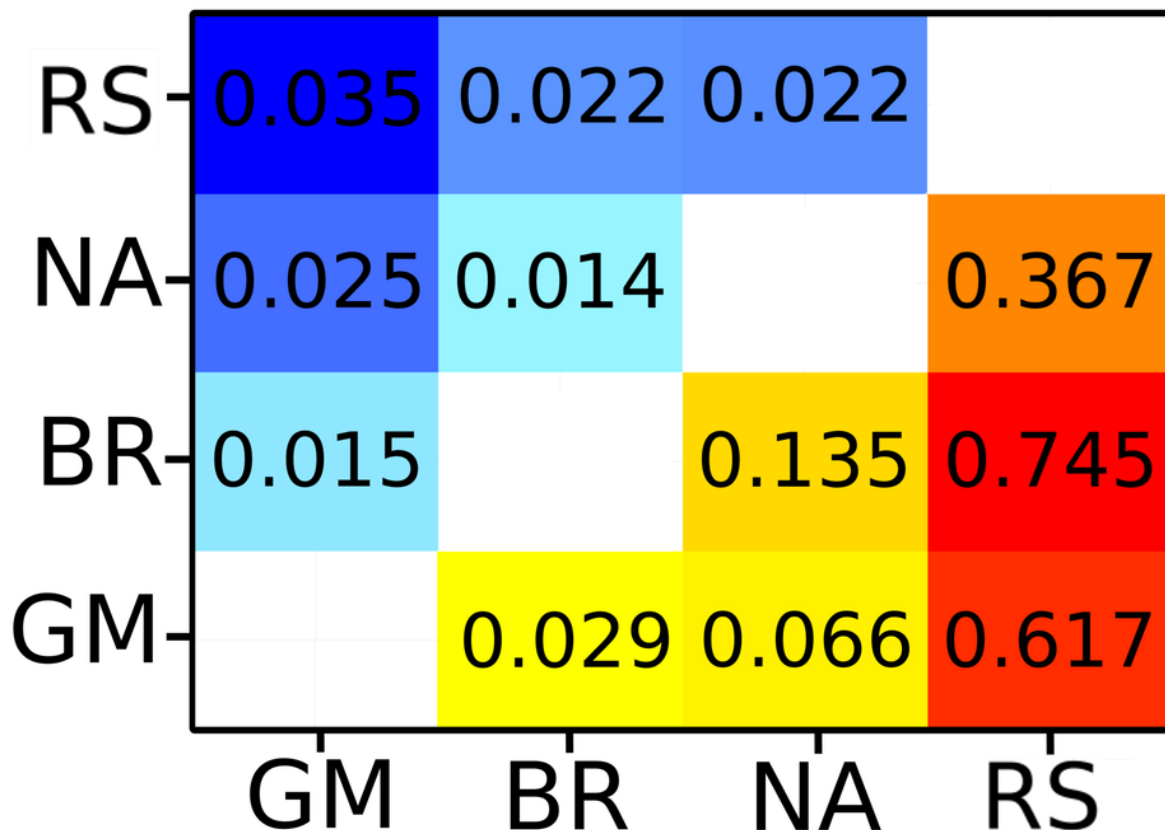


Figure 3

Cost comparisons between sequencing projects using a single Sanger marker to projects using Pool-seq with varying numbers of pools.

(A) Sequencing costs comparing number of individuals to total cost between Sanger at our facility and three Pool-seq projects at our facility containing 4, 8, and 12 pools respectively, where pool sizes change with number of individuals. **(B)** Sequencing cost per individual with fixed pools across different number of Pools.

