PeerJ

Genomics versus mtDNA for resolving stock structure in the silky shark (*Carcharhinus falciformis*)

Derek W. Kraft¹, Emily E. Conklin¹, Evan W. Barba¹, Melanie Hutchinson^{1,2}, Robert J. Toonen¹, Zac H. Forsman¹ and Brian W. Bowen¹

¹ Hawai'i Institute of Marine Biology, University of Hawai'i, Kaneohe, HI, USA

² Joint Institute of Marine and Atmospheric Research, Pacific Islands Fisheries Science Center, NOAA, University of Hawai'i, Honolulu, HI, USA

ABSTRACT

Conservation genetic approaches for elasmobranchs have focused on regions of the mitochondrial genome or a handful of nuclear microsatellites. High-throughput sequencing offers a powerful alternative for examining population structure using many loci distributed across the nuclear and mitochondrial genomes. These single nucleotide polymorphisms are expected to provide finer scale and more accurate population level data; however, there have been few genomic studies applied to elasmobranch species. The desire to apply next-generation sequencing approaches is often tempered by the costs, which can be offset by pooling specimens prior to sequencing (pool-seq). In this study, we assess the utility of pool-seq by applying this method to the same individual silky sharks, Carcharhinus falciformis, previously surveyed with the mtDNA control region in the Atlantic and Indian Oceans. Pool-seq methods were able to recover the entire mitochondrial genome as well as thousands of nuclear markers. This volume of sequence data enabled the detection of population structure between regions of the Atlantic Ocean populations, undetected in the previous study (inter-Atlantic mitochondrial SNPs F_{ST} values comparison ranging from 0.029 to 0.135 and nuclear SNPs from 0.015 to 0.025). Our results reinforce the conclusion that sampling the mitochondrial control region alone may fail to detect fine-scale population structure, and additional sampling across the genome may increase resolution for some species. Additionally, this study shows that the costs of analyzing 4,988 loci using pool-seq methods are equivalent to the standard Sanger-sequenced markers and become less expensive when large numbers of individuals (>300) are analyzed.

Subjects Biogeography, Genetics, Genomics, Molecular Biology, Natural Resource Management **Keywords** Bioinformatics, Cost analysis, Elasmobranch, Genetic markers, Pool-seq, Single nucleotide polymorphisms, Conservation

INTRODUCTION

Many elasmobranchs around the globe have experienced devastating population declines due to overfishing in both target and non-target fisheries (*Musick et al., 2000; Clarke et al., 2006; Ferretti et al., 2010; Heupel et al., 2014; Dulvy et al., 2014; Oliver et al., 2015; Dulvy & Trebilco, 2018*). These species are especially vulnerable to overfishing due to life

Submitted 27 July 2020 Accepted 24 September 2020 Published 21 October 2020

Corresponding author Derek W. Kraft, Kraftd@hawaii.edu

Academic editor Antonio Amorim

Additional Information and Declarations can be found on page 13

DOI 10.7717/peerj.10186

Copyright 2020 Kraft et al.

Distributed under Creative Commons CC-BY 4.0

OPEN ACCESS

history traits such as late maturity, slow growth, low fecundity, and high juvenile mortality, which collectively result in low intrinsic rate of population increase (*Baum et al., 2003*; *Dulvy et al., 2008*). Elasmobranch populations take decades to recover from overfishing, and only if fishing pressure is relieved for an extended period (*Stevens et al., 2000*). Furthermore, many threatened and endangered elasmobranchs have little to no population genetic data that would assist in the resolution of management units (reviewed in *Domingues, Hilsdorf & Gadig (2018)*).

Genetically distinct populations are isolated management units known as stocks; however, stocks can be defined on a smaller scale than genetic populations through other criteria, such as an exclusive economic zone boundry (*Carvalho & Hauser, 1994*; *Ovenden et al., 2015*). Reduced gene flow indicates that if a population is overfished it will not be replenished by immigrants from surrounding populations. This is why managing on a genetic stock-by-stock basis is essential for successful maintenance of exploited species and is sorely needed for over-harvested elasmobranchs (*Dizon et al., 1993*; *Heist, 2004*; *Tallmon et al., 2010*).

For the past two decades the standard for examining population structure in elasmobranchs has been a section of the mitochondrial genome, usually the control region (mtCR) (Duncan et al., 2006; Hoelzel et al., 2006; Keeney & Heist, 2006; Castro et al., 2007; Whitney et al., 2012; Clarke et al., 2015; reviewed in Domingues, Hilsdorf & Gadig (2018)). Though recent studies are moving towards multi-marker approaches (Momigliano et al., 2017; Pazmiño et al., 2018: Green et al., 2019), there is still a large body of literature focusing on mtCR. The mitochondrial genome has a higher rate of mutation than most of the nuclear genome (Brown, George & Wilson, 1979; Charlesworth & Wright, 2001; Neiman & Taylor, 2009) and this rate of mutation is a key advantage in vertebrates with slowly-evolving genomes (Avise et al., 1992; Martin, Naylor & Palumbi, 1992). Elasmobranch mtDNA studies to date have been successful in elucidating population partitions and evolutionary divergences, but the maternal inheritance of mtDNA can limit conclusions about gene flow in cases of sex-biased (usually male) dispersal. Both mtDNA and nuclear markers often have concordant results in sedentary species (Lavery, Moritz & Fielder, 1996; Avise, 2004; Zink & Barrowclough, 2008; DiBattista et al., 2015) but, when examined alone, may miss key components of population structure, particularly in migratory fauna (Pardini et al., 2001; Bowen et al., 2005; Toews & Brelsford, 2012). When highly mobile elasmobranchs are examined with both mtDNA and nuclear markers (usually microsatellites), a different picture often emerges in which females are more resident and males are dispersive (*Pardini et al., 2001*; Schultz et al., 2008; Portnoy et al., 2010; Karl et al., 2011; Daly-Engel et al., 2012; Portnoy et al., 2015: Bernard et al., 2017; Domingues et al., 2018). Identifying outlier SNPs in the nuclear genome can highlight genes possibly under selection, or show functional responses to environmental changes that have important management consequences (Jones et al., 2012; Fischer et al., 2013; Barrio et al., 2016; Guo, Li & Merilä, 2016). Therefore, the combination of mitochondrial and nuclear markers can yield fundamental ecological and evolutionary insights.

High-throughput sequencing is a powerful tool for revealing fine-scale population structure that may be missed by single locus studies (Andrews et al., 2016; Hohenlohe et al., 2018). However, this method can be costly, especially when examining many individuals as is typical of population genetic or phylogeography studies, and the perceived cost may prevent some from considering a high-throughput sequencing approach. For population genetics approaches based on differences in allele frequencies among populations, equimolar pooling of samples before sequencing is an affordable and accurate strategy for large-scale genetic analysis (Schlötterer et al., 2014). Several studies have successfully resolved population structure using a pooled site-associated DNA approach known as pool-seq, including some in commercially valuable marine species (*Gautier et al., 2013*; Mimee et al., 2015). Pool-seq provides estimates of allele frequencies for thousands of loci distributed across the genome simultaneously, which in some cases gives greater statistical power that can actually exceed the accuracy of allele frequency estimates based on individual sequencing (Futschik & Schlötterer, 2010, but also see Anderson, Skaug & Barshis, 2014). Therefore, a comparison of results between the standard mtCR analysis and high-throughput pool-seq is informative in evaluating the relative power and cost of the two approaches for examining population structure.

The silky shark (Carcharhinus falciformis (Müller & Henle, 1839)) is the second most commonly harvested shark on Earth (Rice & Harley, 2013; Oliver et al., 2015). They are one of the top contributors to the shark fin trade and the most common elasmobranch bycatch species in tuna purse-seine fisheries around the world (Clarke et al., 2006; Oliver et al., 2015; Cardeñosa et al., 2018). This pelagic shark, formerly abundant in all tropical oceans, has declined by an estimated 85% in the last 20 years, and is now listed as vulnerable and declining by the International Union for the Conservation of Nature (*Rice & Harley, 2013; IUCN, 2017*). Currently silky shark population assessments are conducted at the scale of regional fishery management organization, and conservation management measures are implemented at this scale in the absence of genetic or movement data to define population boundaries. Clarke et al. (2015) surveyed silky sharks across these regional management regions and found the western Atlantic was strongly differentiated from the Indian Ocean, but the North Atlantic, Gulf of Mexico, and Brazil could not be differentiated and appeared to comprise a single population. In contrast, using the same mtCR marker, Domingues et al. (2017) examined five regions across the Western Atlantic and found the North Western Atlantic was distinct from the South Western Atlantic. The difference between the two studies results from additional sampling in the South West Atlantic from further south than Clarke et al. (2015).

In an era where wildlife management needs far exceed the financial resources to address them, many seek to find the most accessible, robust, and economical means to define management units. In this study, we provide a direct comparison of population genetic analysis methods between Sanger sequencing of the mtCR region and high-throughput sequencing of regional pools of individuals. The same individuals from *Clarke et al. (2015)* were re-sequenced using pool-seq approaches. Regions re-sequenced included Gulf of Mexico, North West Atlantic, and Brazil, as well as one geographically distant location in the Red Sea (Fig. 1). We focused this analysis on SNPs from the mitochondrial DNA as

well as nuclear DNA. We did not analyze any microsatellite loci because they were not a part of *Clarke et al. (2015)*. We then evaluate the economics of conducting pool-seq relative to conventional Sanger sequencing of these same individuals. Ecological and management implications will be addressed in a subsequent companion paper.

MATERIALS AND METHODS

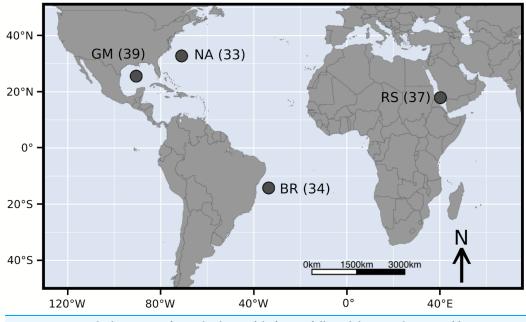
Sampling and sequencing

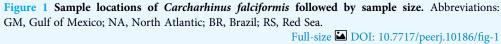
A total of 143 silky shark fin clips or muscle sections were sampled from commercial or artisanal fisheries across four geographic regions and are the same samples examined in *Clarke et al.* (2015). Specifically, we sampled the Gulf of Mexico (GM, n = 39), the North Atlantic (NA, n = 33), Brazil (BR, n = 34), and the Red Sea (RS, n = 37). These sample sizes are slightly lower than *Clarke et al.* (2015). This reduction was due to DNA degradation over time and the need for high-quality genomic DNA for pool-seq. This is contrary to the DNA quality needed for amplifying a single marker from the mitochondrial control region. Additionally only a subset of the Red Sea samples were randomly selected to keep sample sizes relatively similar.

DNA was extracted using Qiagen DNeasy Blood & Tissue kit (Qiagen, Mississauga, ON, Canada), following manufacturer protocols. Extracted DNA quality was assessed visually by gel electrophoresis and imaged using Gel Doc E-Z System (BIO RAD, Hercules, CA, USA). Only DNA aliquots with strong genomic DNA bands were further processed, while degraded or overly digested DNA was discarded. Aliquots of high-quality DNA were quantified using an AccuClear Ultra high sensitivity dsDNA quantitation kit (Biotium, Fremont CA, USA) and a SpectroMax M2 (Molecular Devices, Sunnyvale, CA, USA). Libraries were pooled with an equal amount of DNA (ng/µl) contributed per individual to minimize individual contribution bias, totaling 2,000 ng of DNA per library. Numbers of individuals per pool are displayed in Fig. 1. No PCR was performed to ensure individual DNA contribution was kept equal within and across libraries (Anderson, Skaug & Barshis, 2014). The rest of the library preparation followed the ezRAD library preparation protocol (Toonen et al., 2013; Knapp et al., 2016). This included DNA digested with DpnII restriction enzyme and adapters ligated using a KAPA HyperPrep (KAPA Biosystems, Wilmington, MA, USA). Pooled libraries were sequenced using Illumina MiSeq (v3 2x300bp PE) at the Hawai'i Institute of Marine Biology EPSCoR Core sequencing facility.

Genetic analyses

MultiQC was used to assess sequence quality scores, sequence length distributions, duplication levels, and overrepresented sequences (*Ewels, Lundin & Max, 2016*). To analyze the mitochondrial genome, a previously published mitochondrial genome from *Carcharhinus falciformis* was used as a reference (GeneBank accession number KF801102). Raw paired-end reads were trimmed with TRIMMOMATIC, mapped to the mitochondrial genome reference BWA (mem algorithm), and variants called using the dDocent bioinformatics pipeline, modified for pool-seq (*Puritz, Hollenbeck & Gold, 2014*,





see below for details). Called SNPs were then analyzed with AssessPool (github.com/ ToBoDev/assessPool, see below for details).

The bioinformatics pipeline included dDocent followed by AssessPool. Given that no reference genome was available, a reference was constructed using the dDocent de novo assembly and optimized utilizing the reference optimization steps provided on the dDocent assembly tutorial (http://ddocent.com/assembly/). Before assembly reads were trimmed using default settings and then an overlap (OL) assembly was performed, followed by clustering with CD-HIT with a-c parameter of 90% similarity. For mapping using BWA (mem algorithm) all match, mismatch, and gap open penalty score parameters were also default settings. Different parameters were tested during optimization but did not improve mapping. Within-pool (K1) and between-pool (K2) minimum locus depth values selected for the de novo assembly did impact the results. dDocent provides graphical outputs to help select these values; however, testing a few different values of each is recommended to fully explore the potential of the data by balancing number of contigs by coverage depth (see ddocent.com/UserGuide for details). Selected values for K1 and K2 were 3 and 3 respectively. Once assembled, sequences were mapped, SNPs were called within the dDocent pipeline using FreeBayes, modified for SNP calling in pools (Garrison & Marth, 2012, https://github.com/ekg/freebayes). Any contigs that aligned to the mitochondrial genome were removed from this nuclear dataset. The contigs that aligned specifically to the mitochondrial control region were saved for SNP validation to directly compare the results from this pool-seq approach to those previously reported by Clarke et al. (2015).

SNP calling with FreeBayes was optimized for pooled samples using the 'pooledcontinuous' option, and minor allele frequency was decreased to 0.05 to capture alleles with frequency greater than 5% in the population (See Supplemental Material for code). The dDocent pipeline outputs SNPs in two variant call format files (.vcf), one with all raw SNPs (TotalRawSNPs.vcf) and another with filtered SNPs (Final.recode.vcf) however dDocent does not optimize filtering for pool-seq data. Therefore, the raw SNPs were processed with the pool-seq specific program AssessPool, which uses VCFtools and vcflib to filter SNPs (*Danecek et al., 2011*). SNPs were processed with the following filters: minimum pool number of 2, minimum quality score of 20, minimum depth threshold of 30, maximum amount of missing data of 3, maximum allele length of 10, quality score to depth ratio of 0.25 as well as mean depth per site vs. quality score, and finally a maximum mean depth threshold of 1,000 (Table S1). AssessPool then sends filtered SNPs to either PoPoolation2 (Kofler, Pandey & Schlötterer, 2011) or poolfstat (Hivert et al., 2018). PoPoolation2 calculates mean pairwise F_{st} values and significance in the form of *p*-values obtained using Fisher's Exact Test and combined using Fisher's method (as described in Ryman et al. (2006)). Poolfstat (Hivert et al., 2018) takes a different approach, calculating F_{ST} values based on an analysis-of-variance framework (sensu Weir & Cockerham, 1984) to eliminate biases associated with varying pool sizes. AssessPool then organizes, summarizes, and creates visualizations of the data using RStudio (RStudio Team, 2020).

As a quality control test, sequences from *Clarke et al.* (2015) were downloaded from GenBank (accession numbers KM267565–KM267626), and SNPs from these data were compared directly to SNPs called within the control region of the mitochondrial pool-seq data generated here. Concordance of this validation set of SNPs was determined by Mantel test in R (*Legendre & Legendre, 1998*) comparing the matrices of pairwise F_{ST} values among populations.

Cost analysis

The cost of the pool-seq approach compared to Sanger sequencing of individual loci was calculated based on library preparation and sequencing cost at our facility. We did not include labor but calculated the total cost to generate sequence data from each sample included here from such expenses as the extraction, laboratory consumables, PCR amplification, library preparation, reaction clean-ups, quantification, quality control testing, and sequencing costs. These costs were translated into functions in RStudio (*RStudio Team, 2020*) in which Sanger sequencing is a fixed rate per individual and pool-seq costs are fixed per flow cell on the Illumina MiSeq sequencer MiSeq, but individual cost varies based on number of individuals and number of pooled regions per sequencing run. These functions were then plotted together for comparison.

RESULTS

A total of 30.8 million reads were generated for the four geographic regions, which averaged 7.7 ± 3.0 million reads per pooled library. Results from the MutliQC assessment showed fairly homogenous output between libraries in regard to sequence quality scores,

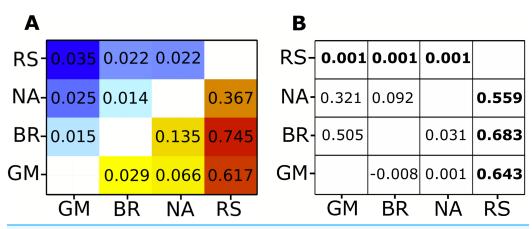


Figure 2 F_{st} heat map for mitochondrial and nuclear data compaired to Clarke et al. ϕ_{ST} and *P*-values. (A) Pairwise F_{st} values generated by pool-seq methods. Cool colors (top left) are F_{st} values calculated from nuclear genome loci, warm colors (bottom right) are F_{st} values from loci across the entire mitochondrial genome. All pairwise differences are significant (P < 0.001). (B) ϕ_{ST} results from *Clarke et al. (2015)* on the lower right triangle and *P*-values on the upper right triangle. Significant *P*-values and corresponding ϕ_{ST} values in bold. Regional abbreviation are as follows; GM, Gulf of Mexico; BR, Brazil; NA, North Atlantic; RS, Red sea. Full-size DOI: 10.7717/peerj.10186/fig-2

GC and per base sequence content, sequence length distributions, duplication levels, overrepresented sequences, and adapter content. Once assembled, aligned, and mapped, 5,792 SNPs were resolved across the mitochondrial and nuclear genomes combined. There were 4,103 biallelic SNPs, 168 multialleleic SNPs and 48 insertions and deletions (INDELs). INDELs and multiallelic SNPs remain a challenge for quantification software, so we restricted our analysis to biallelic loci (*Fracassetti, Griffin & Willi, 2015*). AssessPool creates visualizations of F_{ST} values and allows for visual outlier inspection. No visual outliers were present and given these SNPs are distributed haphazardly across the genome, therefore they are assumed to be putatively neutral.

Mitochondrial genome

Analysis of the complete mitochondrial genome (17,774 bp) revealed 804 variable sites: 681 biallelic and 17 multiallelic SNPs. Because coverage in this dataset was fairly low on average, most of these SNPs did not meet the filter threshold. After further filtering for the highest quality markers, 30 SNPs were selected to calculate allele frequencies. Pairwise F_{st} values were all significant (Fig. 2; Table S2). The Red Sea had much higher F_{st} values (ranging from 0.367 to 0.745) than any inter-Atlantic comparison (ranging from 0.029 to 0.135). However, all comparisons within the Atlantic still showed significant F_{ST} values, the highest being between the North Atlantic and Brazil, and the lowest between Brazil and Gulf of Mexico (Fig. 2; Table S2).

Nuclear loci

Our nuclear data showed 4,988 variants, of which 3,422 were biallelic SNPs and 151 were multialleleic SNPs. A total of 346 SNPs remained after the same filtering process for the highest quality SNPs was applied as for the mitochondrial genome. Nuclear markers showed lower F_{ST} values between locations than the mitochondrial data, yet all

comparisons were still significant (Fig. 2; Table S2). The Red Sea showed consistently higher F_{ST} values in comparison to inter-Atlantic comparisons except for the North Atlantic to Gulf of Mexico comparison, which showed the second highest mean F_{ST} value (Fig. 2; Table S2). The highest value ($F_{ST} = 0.035$) was observed between Gulf of Mexico and the Red Sea, whereas the lowest ($F_{ST} = 0.014$) was between the North Atlantic and Brazil, which had the highest F_{ST} value within the Atlantic for the mitochondrial data.

SNP validation

SNPs called in the mitochondrial control region using the pool-seq protocol were compared with those reported in *Clarke et al. (2015)*. Of the 34 SNPs in their study, 14 had a minor allele count (MAC) of less than or equal to 3 and several were singletons. These SNPs are removed from the pool-seq data due to MAC SNP filter of >3 to remove sequencing errors that might be scored as rare alleles during high-throughput sequencing. Therefore, singletons or any rare allele represented fewer than three times in a population will inherently be removed from pool-seq data sets. Fortunately those rare alleles do not tend to overly impact F_{ST} values and should not bias interpretations of population structure (*Bird et al., 2011; Toonen et al., 2011*). Three SNPs were found in the Clarke study with a MAC of >3 that were not present in the pool-seq data; however, the remining 17 SNPs were all present in our data, plus one that was not found in the Clarke study (Fig. S1). Despite the loss of these rare alleles from the SNP validation set, pairwise F_{st} values estimated by both methods remained highly correlated (Mantel test, $r^2 = 0.96$, p < 0.05), and comparisons between the Red Sea and all three Atlantic populations showed the same relative magnitude between both methods.

Cost analysis

The findings for cost analysis indicate that pool-seq reaches a threshold at approximately 300 individuals, after which this approach offers cheaper results than individual Sanger sequences. Furthermore, the cost is only twice as expensive at just over 100 individuals (Fig. 3A). The pool-seq approach provides a far higher ratio of information for the cost, yielding greater population resolution. This cost assessment does not include analytical time, labor, or effort associated with pool-seq analyses such as access to computer resources and expertise with bioinformatic pipelines. However, these costs are likely to decrease in the near future as bioinformatic pipelines are improved and become more widely available, for example as applications deployed via cloud based platforms such as Galaxy (https://usegalaxy.org/) or CyVerse (https://cyverse.org/). It is also important to note that the choice of pool-seq methodology has many caveats, which are discussed in greater detail in the 'considerations on pool-seq' section of the discussion below.

DISCUSSION

Elasmobranchs are being harvested at unsustainable levels in several commercial fishing industries around the world. A fundamental step in successful management of any species is resolving population boundaries so they can be managed on a genetic stock-by-stock basis. As genetic sequencing technologies advance, there is greater opportunity to

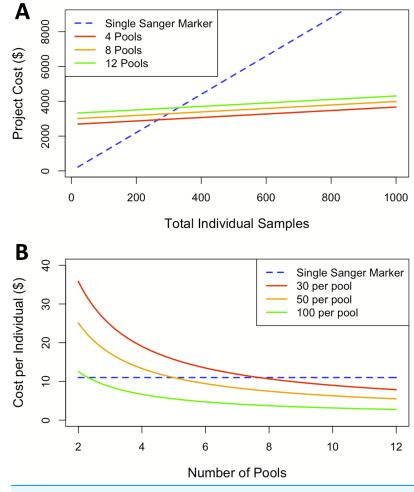


Figure 3 Cost comparisons between sequencing projects using a single Sanger marker to projects using pool-seq with varying numbers of pools. (A) Sequencing costs comparing number of individuals to total cost between Sanger at our facility and three pool-seq projects at our facility containing 4, 8, and 12 pools respectively, where pool sizes change with number of individuals. (B) Sequencing cost per individual with fixed pools across different number of Pools.

Full-size 🖾 DOI: 10.7717/peerj.10186/fig-3

detect even small-scale genetic differences between populations. When these differences amount to statically significant allele frequencies at the population level, this indicates limited exchange among distinct stocks.

Here, we validate the utility of pool-seq using the same individuals as a previous study (*Clarke et al., 2015*) and show that pool-seq recovers additional population structure relative to Sanger sequencing of the mtDNA control region. Pool-seq was able to detect isolated populations between the Gulf of Mexico, Western Atlantic, and along the Brazilian coast, where *Clarke et al. (2015)* found no population structure. As expected, the Red Sea population was highly isolated from Atlantic conspecifics using both approaches.

One advantage of this pool-seq approach is that we recover SNPs throughout the entire mitochondrial genome along with thousands of additional nuclear loci that together provide greater statistical power to detect finer scale population structure (*Ryman & Palm, 2006; Larsson et al., 2009; Kurland et al., 2019*). The pool-seq approach yielded significant genetic structure among inter-Atlantic regions in both mtDNA and nuclear loci, whereas Sanger sequencing of the mtCR lacked power to resolve significant differences among the same populations. The congruence between the mitochondrial genome and nuclear loci reinforces the conclusion of population structure among all regions sampled in this study.

In this case, pool-seq lived up to the promise of increased power to detect fine-scale structure, but does it live up to the promise (Ferretti, Ramos-Onsins & Pérez-Enciso, 2013; Schlötterer et al., 2014) of being cost-effective? Individual extraction costs remain fixed across both approaches and Sanger sequencing generally has a flat rate per individual, including PCR primers and reagents, and sequencing per individual per locus. In contrast, pool-seq has a flat sequencing cost determined by the number of reads generated from the high-throughput sequencing platform, plus a small additional cost per pool for the exact quantification of DNA for equimolar pooling and the library preparation for high-throughput sequencing. Comparing costs at our institution between a single Sanger sequencing marker and pool-seq on the Illumina MiSeq platform indicates pool-seq becomes less expensive when sample size of the study rises above 300 individuals. Although the cost per pool is essentially fixed, when higher numbers of individuals are included per pool, the price per individual analyzed is further reduced (Fig. 3B). Our comparison here is limited to 12 pools due to the maximum number of reads per lane produced on the MiSeq platform. Therefore, analyzing more than 12 pools would require additional sequencing runs and result in a step increase in the cost per individual/pool, although this would differ among other Illumina machines (such as the HiSeq, NextSeq or NovaSeq) or other high-throughput sequencing platforms (such as the PacBio Sequel II). Larger numbers of pools could be run on some of these machines, but with differing individual read lengths and sequencing depths, which also introduce other trade-offs. Likewise, samples can also be run with individual barcodes, therefore gaining the individual information lost by pooling specimens, but with increased initial setup and sequencing costs. There are so many options by which to apply these methods that we cannot possibly consider them all here, and the availability, cost, and trade-offs associated with each should be ideally considered by individuals when designing high-throughput sequencing projects. In our case, we considered only the options currently available to us through our campus sequencing core, and all these pool-seq price comparisons are to a single Sanger-sequenced marker. Thus, when considering the information acquired from pool-seq compared to the cost from traditional single mitochondrial marker the price per individual advantage is massively amplified.

Considerations with pool-seq

As with any sequencing technique, there are still several factors to consider before deciding if pool-seq is appropriate for a particular study. Multiple reviews have been published on high-throughput and pool-seq approaches demonstrating pros, cons, and considerations with these methods, which are beyond the scope of this study. Interested readers should consult *Perez-Enciso & Ferretti* (2010), *Futschik & Schlötterer* (2010), *Kofler, Betancourt & Schlötterer* (2012), *Ferretti, Ramos-Onsins & Pérez-Enciso* (2013), *Schlötterer et al.* (2014), *Andrews & Luikart* (2014), *Andrews et al.* (2016) and *Kurland et al.* (2019).

Pooling assumes individuals are from the interbreeding individuals within a single population of the same species. Therefore, care needs to be taken to avoid cryptic species, combining multiple populations (Wahlund effect), or other unintentional bias when selecting individuals to pool (Garnier-Géré & Chikhi, 2013). For wide ranging pelagic species such as the blue shark or oceanic whitetip it seems reasonable to pool individuals from a larger area than it would be for small benthic species such as horn sharks, wobbegongs, or most rays. Population structure may be obscured if the geographic range per pool is too large or if there is complex population structure (sensu *Bowen et al., 2005*), because individuals from multiple sub-populations will be mixed into a single pool from which allele frequencies are calculated. Certainly pool-seq is not appropriate in all cases. It is a cost-saving approach for analyses based on allele frequencies only, because individual information is lost by pooling, including haplotypes/genotypes and linkage disequilibrium information. Also, pooling makes it difficult to distinguish between low frequency alleles in the population and sequencing error. Therefore, careful filtering must be applied to ensure only valid SNPs are analyzed instead of analyzing sequencing noise (Anand et al., 2016; Schlötterer et al., 2014). Finally, the estimation of F_{ST} from pooled data remains a subject of some debate, and new approaches and bias corrections are being actively developed (Kofler, Pandey & Schlötterer, 2011; Hivert et al., 2018). To account for this uncertainty, we include analyses based on both the original PoPoolation2 (Kofler, Pandey & Schlötterer, 2011) package and the newer poolfstat (Hivert et al., 2018) that explicitly considers potential biases associated with varying pool sizes. The two approaches yield slightly different F_{st} values (see Table S2), however a comparison of the two F_{st} matrices shows strong correlation (Mantel r = 0.991 for mitochondrial and r = 0.978 for nuclear data, p < 0.05). Therefore, only those F_{st} values calculated by PoPoolation2 are reported in the main text for ease of presentation.

Though pool-seq has been shown to be an affordable and reliable tool for population genomics (*Futschik & Schlötterer*, 2010; *Gautier et al.*, 2013; *Rellstab et al.*, 2013; *Konczal et al.*, 2014; *Schlötterer et al.*, 2014; *Kurland et al.*, 2019), projects with larger budgets could allocate funds for any of a variety of other genomic sequencing techniques such as individual RADseq libraries (*Hohenlohe et al.*, 2010), GBS (*Narum et al.*, 2013), SNP arrays (*Qi et al.*, 2017), bait capture (*Feutry et al.*, 2020), or low coverage genomewide sequencing (*Therkildsen & Palumbi*, 2017). These approaches allow for individual genotyping to examine questions that require individual-level information and could provide a deeper assessment of populations. However it is also important to consider not all labs can afford to generate genomic level data, especially in developing countries, and having a cost-effective alternative to single marker studies will continue to be invaluable to many.

CONCLUSIONS

The finding of population structure on the scale of North Atlantic/Gulf of Mexico/Brazil is nearly unprecedented for a pelagic shark. Population structure in globally distributed sharks is typically detected on a scale of ocean basins (Atlantic versus Indo-Pacific, *Castro et al., 2007; Graves & McDowell, 2015*) and a few pelagic fishes have no population structure on a global scale (e.g., Basking shark, *Cetorhinus maximus, Hoelzel et al., 2006;* Blue shark *Prionace glauca, Veríssimo et al., 2017;* Wahoo, *Acanthocybium solandri, Theisen et al., 2008*). The resolution of isolated populations on the scale of North Atlantic Ocean is more typical of coastal species than pelagic species. The silky shark seems to be a pelagic species with a somewhat coastal population structure. This has strong implications for international management because smaller stocks imply smaller populations which are more readily depleted. At a minimum, these data require rethinking a single population management approach for the Atlantic, and this pattern needs to be investigated for this species across the Indo-Pacific as well.

Overall this study demonstrates pool-seq is a powerful and cost-effective tool for analyzing large portions of the genome which the methods traditionally used for elasmobranchs could not supply. Sharks and rays are an imperiled group of species that could benefit from advanced genomic studies to outline appropriate management units. Finally, although the technology is becoming cheaper and easier to apply, it is a common pitfall to assume everyone in the field can afford, or must use, these approaches to produce defensible science. Bowen et al. (2014) advocate judicious rather than wholesale application of genomic approaches as the most robust course of study, particularly when considering the global inequities in available research budgets. Sanger sequencing is still more cost effective for small numbers of individuals, but as the number of individuals included in a study rise, the cost per individual reaches the point where high throughput sequencing studies can be cheaper than sequencing a single mitochondrial marker from each individual. We provide an example of just such a case here, and highlight the potential advantage of cost savings together with increased power for resolution of fine scale population structure. Though there is still additional cost of using cluster computer servers and bioinformatics programs, these cost are dropping as technology advances. When study organism and sampling strategies are assessed and implemented into the study design, pool-seq has great promise for augmenting the scientific foundations for management of marine recourses.

ACKNOWLEDGEMENTS

This study was made possible by the generous donation of specimens by Christopher R. Clarke, Mahmood Shivji, Stephen A. Karl, J.D. Filmalter, and Julia Spaet. We thank members of the ToBo Lab for sharing expertise, advice and discussions that contributed to this manuscript. Special thanks to Darren Lerner, Kim Holland, Carl Meyer, S. Gulak, D. Bethe, D. McCauley, C. Wilson, Guy Harvey Ocean Foundation, and Save Our Seas Foundation. The views expressed herein are those of the authors and do not necessarily reflect the views of NOAA or any of its subagencies. This is contribution #1821 from the Hawaii Institute of Marine Biology, contribution #JC-15-32 from the Hawaii Sea Grant Program, and contribution #11128 from the School of Ocean and Earth Science and Technology at the University of Hawaii.

ADDITIONAL INFORMATION AND DECLARATIONS

Funding

This paper is funded by a grant from the National Oceanic and Atmospheric Administration, Project R/FM-18, which is sponsored by the University of Hawaii Sea Grant College Program, SOEST, under Institutional Grant No. NA05OAR4171048 (to Brian W. Bowen) from NOAA Office of Sea Grant, Department of Commerce. Additional funding was provided by the National Science Foundation (OCE-15-58852 to Brian W. Bowen). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Grant Disclosures

The following grant information was disclosed by the authors: National Oceanic and Atmospheric Administration: R/FM-18. NOAA Office of Sea Grant, Department of Commerce: NA05OAR4171048. National Science Foundation: OCE-15-58852.

Competing Interests

Robert J. Toonen is an Academic Editor for PeerJ.

Author Contributions

- Derek W. Kraft conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Emily E. Conklin analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Evan W. Barba analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Melanie Hutchinson performed the experiments, authored or reviewed drafts of the paper, and approved the final draft.
- Robert J. Toonen conceived and designed the experiments, analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.
- Zac H. Forsman conceived and designed the experiments, analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.
- Brian W. Bowen conceived and designed the experiments, authored or reviewed drafts of the paper, and approved the final draft.

Animal Ethics

The following information was supplied relating to ethical approvals (i.e., approving body and any reference numbers):

Samples were collected aboard commercial fishing vessels therefore no IACUC approvals were needed.

Data Availability

The following information was supplied regarding data availability: Raw data are available at the NCBI Sequence Read SRA: PRJNA647728

Supplemental Information

Supplemental information for this article can be found online at http://dx.doi.org/10.7717/ peerj.10186#supplemental-information.

REFERENCES

- Anand S, Mangano E, Barizzone N, Bordoni R, Sorosina M, Clarelli F, Corrado L, Boneschi FM, D'Alfonso S, De Bellis G. 2016. Next generation sequencing of pooled samples: guideline for variants' filtering. *Scientific Reports* 6(1):1–9 DOI 10.1038/s41598-016-0001-8.
- Anderson EC, Skaug HJ, Barshis DJ. 2014. Next-generation sequencing for molecular ecology: a caveat regarding pooled samples. *Molecular Ecology* 23(3):502–512 DOI 10.1111/mec.12609.
- Andrews KR, Good JM, Miller MR, Luikart G, Hohenlohe PA. 2016. Harnessing the power of RADseq for ecological and evolutionary genomics. *Nature Reviews Genetics* 17(2):81–92 DOI 10.1038/nrg.2015.28.
- Andrews KR, Luikart G. 2014. Recent novel approaches for population genomics data analysis. Molecular Ecology 23(7):1661–1667 DOI 10.1111/mec.12686.
- Avise JC. 2004. Molecular markers, natural history and evolution. In: *Sinauer Associates*. 2nd Edition. New York: Chapman & Hall.
- Avise JC, Bowen BW, Bermingham E, Meylan AB, Lamb T. 1992. Mitochondrial DNA evolution at a turtle's pace: evidence for low genetic variability and reduced microevolutionary rate in the Testudines. *Molecular Biology & Evolution* 9:457–473.
- Barrio AM, Lamichhaney S, Fan G, Rafati N, Pettersson M, Zhang H, Dainat J, Ekman D, Höppner M, Jern P, Martin M, Börn N, Liu X, Chen W, Liang X, Shi C, Fu Y, Ma K, Zhan X, Feng C, Gustafson U, Rubin C-J, Sällman Almén M, Blass M, Casini M, Folkvord A, Laikre L, Ryman N, Ming-Yuen Lee S, Xu X, Andersson L. 2016. The genetic basis for ecological adaptation of the Atlantic herring revealed by genome sequencing. *eLife* 5:1–32 DOI 10.7554/eLife.12081.
- Baum JK, Myers RA, Kehler DG, Worm B, Harley SJ, Doherty PA. 2003. Collapse and conservation of shark populations in the Northwest Atlantic. *Science* 299(5605):389–392 DOI 10.1126/science.1079777.
- Bernard AM, Horn RL, Chapman DD, Feldheim KA, Garla RC, Brooks EJ, Gore MA, Shivji MS. 2017. Genetic connectivity of a coral reef ecosystem predator: the population genetic structure and evolutionary history of the Caribbean reef shark (*Carcharhinus perezi*). *Journal of Biogeography* 44(11):2488–2500 DOI 10.1111/jbi.13062.
- **Bird CE, Karl SA, Smouse PE, Toonen RJ. 2011.** Detecting and measuring genetic differentiation. *Phylogeography and Population Genetics in Crustacea* **19(3)**:31–55.

- Bowen BW, Bass AL, Soares L, Toonen RJ. 2005. Conservation implications of complex population structure: lessons from the loggerhead turtle (*Caretta caretta*). *Molecular Ecology* **14(8)**:2389–2402 DOI 10.1111/j.1365-294X.2005.02598.x.
- Bowen BW, Shanker K, Yasuda N, Celia M, Malay MC(Machel) D, Von der Heyden S, Paulay G, Rocha LA, Selkoe KA, Barber PH, Williams ST, Lessios HA, Crandall ED, Bernardi G, Meyer CP, Carpenter KE, Toonen RJ. 2014. Phylogeography unplugged: comparative surveys in the genomic era. *Bulletin of Marine Science* 90(1):13–46.
- Brown WM, George M, Wilson AC. 1979. Rapid evolution of animal mitochondrial DNA. Proceedings of the National Academy of Sciences of the United States of America 76(4):1967–1971 DOI 10.1073/pnas.76.4.1967.
- Cardeñosa D, Fields AT, Babcock EA, Zhang H, Feldheim K, Shea SKH, Fischer GA, Chapman DD. 2018. CITES-listed sharks remain among the top species in the contemporary fin trade. *Conservation Letters* 11(4):e12457 DOI 10.1111/conl.12457.
- Carvalho GR, Hauser L. 1994. Molecular genetics and the stock concept in fisheries. *Reviews in Fish Biology and Fisheries* 4(3):326–350 DOI 10.1007/BF00042908.
- Castro ALF, Stewart BS, Wilson SG, Hueter RE, Meekan MG, Motta PJ, Bowen BW, Karl SA. 2007. Population genetic structure of Earth's largest fish, the whale shark (*Rhincodon typus*). *Molecular Ecology* 16(24):5183–5192 DOI 10.1111/j.1365-294X.2007.03597.x.
- Charlesworth D, Wright SI. 2001. Breeding systems and genome evolution. *Current Opinion in Genetics & Development* 11(6):685–690 DOI 10.1016/S0959-437X(00)00254-9.
- Clarke CR, Karl SA, Horn RL, Bernard AM, Lea JS, Hazin FH, Prodöhl PA, Shivji MS. 2015. Global mitochondrial DNA phylogeography and population structure of the silky shark, Carcharhinus falciformis. *Marine Biology* **162(5)**:945–955 DOI 10.1007/s00227-015-2636-6.
- Clarke SC, McAllister MK, Milner-Gulland EJ, Kirkwood GP, Michielsens CGJ, Agnew DJ, Pikitch EK, Nakano H, Shivji MS. 2006. Global estimates of shark catches using trade records from commercial markets. *Ecology Letters* **9(10)**:1115–1126 DOI 10.1111/j.1461-0248.2006.00968.x.
- Daly-Engel TS, Seraphin KD, Holland KN, Coffey JP, Nance HA, Toonen RJ, Bowen BW. 2012. Global phylogeography with mixed-marker analysis reveals male-mediated dispersal in the endangered scalloped hammerhead shark (sphyrna lewini). *PLOS ONE* 7(1):e29986 DOI 10.1371/journal.pone.0029986.
- Danecek P, Auton A, Abecasis G, Albers CA, Banks E, Depristo MA, Handsaker RE, Lunter G, Marth GT, Sherry ST, Mcvean G, Durbin R, 1000 Genomes Project Analysis Group. 2011. The variant call format and VCFtools. *Bioinformatics* 27(15):2156–2158
 DOI 10.1093/bioinformatics/btr330.
- DiBattista JD, Waldrop E, Rocha LA, Craig MT, Beruman ML, Bowen BW. 2015. Blinded by the bright: a lack of congruence between color morphs, phylogeography, and taxonomy in a cosmopolitan Indo-Pacific butterflyfish, Chaetodon auriga. *Journal of Biogeography* 42(10):1919–1929 DOI 10.1111/jbi.12572.
- Dizon AE, Lockyer C, Perrin WF, Demaster DP, Sisson J. 1993. Rethinking the stock concept: a phylogeographic approach. *Biological Conservation* **64(2)**:176–177 DOI 10.1016/0006-3207(93)90670-V.
- **Domingues RR, Bruels CC, Gadig OBF, Chapman DD, Hilsdorf AWS, Shivji MS. 2018.** Genetic connectivity and phylogeography of the night shark (*Carcharhinus signatus*) in the western Atlantic Ocean: Implications for conservation management. *Aquatic Conservation: Marine and Freshwater Ecosystems* **29(1)**:102–114 DOI 10.1002/aqc.2961.

- **Domingues RR, Hilsdorf AWS, Gadig OBF. 2018.** The importance of considering genetic diversity in shark and ray conservation policies. *Conservation Genetics* **19(3)**:501–525 DOI 10.1007/s10592-017-1038-3.
- **Domingues RR, Hilsdorf AWS, Shivji MM, Hazin FVH, Gadig OBF. 2017.** Effects of the Pleistocene on the mitochondrial population genetic structure and demographic history of the silky shark (*Carcharhinus falciformis*) in the western Atlantic Ocean. *Reviews in Fish Biology and Fisheries* **28(1)**:213–227 DOI 10.1007/s11160-017-9504-z.
- Dulvy NK, Baum JK, Clarke S, Compagno LJV, Cortés E, Domingo A, Fordham S, Fowler S, Francis MP, Gibson C, Martínez J, Musick JA, Soldo A, Stevens JD, Valenti S. 2008. You can swim but you can't hide: the global status and conservation of oceanic pelagic sharks and rays. *Aquatic Conservation: Marine and Freshwater Ecosystems* 18(5):459–482 DOI 10.1002/aqc.975.
- Dulvy NK, Fowler SL, Musick JA, Cavanagh RD, Kyne PM, Harrison LR, Carlson JK, Nk Davidson L, Fordham SV, Francis MP, Pollock CM, Simpfendorfer CA, Burgess GH, Carpenter KE, Jv Compagno L, Ebert DA, Gibson C, Heupel MR, Livingstone SR, Sanciangco JC, Stevens JD, Valenti S, White WT. 2014. Extinction risk and conservation of the world's sharks and rays. *eLife* 2014(3):1–34 DOI 10.7554/eLife.00590.001.
- **Dulvy NK, Trebilco R. 2018.** Size-based insights into the ecosystem role of sharks and rays. In: Carrier J, Heithaus MR, Simpfendorfer CA, eds. *Shark Research: Emerging Technologies and Applications for the Field and Laboratory.* Vol. 1. Boca Ranton: CRC Press, 25–44.
- Duncan KM, Martin AP, Bowen BW, De Couet HG. 2006. Global phylogeography of the scalloped hammerhead shark (*Sphyrna lewini*). *Molecular Ecology* 15(8):2239–2251 DOI 10.1111/j.1365-294X.2006.02933.x.
- Ewels P, Lundin S, Max K. 2016. Data and text mining MultiQC : summarize analysis results for multiple tools and samples in a single report. *Bioinformatics* 32(19):3047–3048 DOI 10.1093/bioinformatics/btw354.
- Ferretti L, Ramos-Onsins SE, Pérez-Enciso M. 2013. Population genomics from pool sequencing. Molecular Ecology 22(22):5561–5576 DOI 10.1111/mec.12522.
- Ferretti F, Worm B, Britten GL, Heithaus MR, Lotze HK. 2010. Patterns and ecosystem consequences of shark declines in the ocean. *Ecology Lettersol.* 13(ue 8):1055–1071 DOI 10.1111/j.1461-0248.2010.01489.x.
- Feutry P, Devloo-Delva F, Tran LYA, Mona S, Gunasekera RM, Johnson G, Pillans RD, Jaccoud D, Kilian A, Morgan DL, Saunders T, Bax NJ, Kyne PM. 2020. One panel to rule them all: DArTcap genotyping for population structure, historical demography, and kinship analyses, and its application to a threatened shark. *Molecular Ecology Resources* 162(4):13204 DOI 10.1111/1755-0998.13204.
- Fischer MC, Rellstab C, Tedder A, Zoller S, Gugerli F, Shimizu KK, Holderegger R, Widmer A. 2013. Population genomic footprints of selection and associations with climate in natural populations of *Arabidopsis halleri* from the Alps. *Molecular Ecology* 22(22):5594–5607 DOI 10.1111/mec.12521.
- **Fracassetti M, Griffin PC, Willi Y. 2015.** Validation of pooled whole-genome re-sequencing in Arabidopsis lyrata. *PLOS ONE* **10**:e0140462.
- Futschik A, Schlötterer C. 2010. The next generation of molecular markers from massively parallel sequencing of pooled DNA samples. *Genetics* 186(1):207–218 DOI 10.1534/genetics.110.114397.

- Garnier-Géré P, Chikhi L. 2013. Population subdivision, Hardy-Weinberg equilibrium, and the Wahlund effect. eLS Books, Wiley Online Library Pub. Available at https://onlinelibrary.wiley. com/doi/abs/10.1002/9780470015902.a0005446.pub3.
- Garrison E, Marth G. 2012. Haplotype-based variant detection from short-read sequencing. *Available at http://arxiv.org/abs/1207.3907*.
- Gautier M, Foucaud J, Gharbi K, Cézard T, Galan M, Loiseau A, Thomson M, Pudlo P, Kerdelhué C, Estoup A. 2013. Estimation of population allele frequencies from next-generation sequencing data: pool-versus individual-based genotyping. *Molecular Ecology* 22(14):3766–3779 DOI 10.1111/mec.12360.
- Graves JE, McDowell JR. 2015. Population structure of istio- phorid billfishes. *Fisheries Research* 166:21–28.
- Green ME, Appleyard SA, White W, Tracey S, Devloo-Delva F, Ovenden JR. 2019. Novel multimarker comparisons address the genetic population structure of silvertip sharks (Carcharhinus albimarginatus). *Marine and Freshwater Research* **70**(7):1007–1019 DOI 10.1071/MF18296.
- **Guo B, Li Z, Merilä J. 2016.** Population genomic evidence for adaptive differentiation in the Baltic Sea herring. *Molecular Ecology* **25(12)**:2833–2852 DOI 10.1111/mec.13657.
- Heist EJ. 2004. Biology of sahrks and their relatives. Boca Raton: CRC Press.
- Heupel MR, Knip DM, Simpfendorfer CA, Dulvy NK. 2014. Sizing up the ecological role of sharks as predators. *Marine Ecology Progress Series* 495:291–298 DOI 10.3354/meps10597.
- Hivert V, Leblois R, Petit EJ, Gautier M, Vitalis R. 2018. Measuring genetic differentiation from pool-seq data. *Genetics* 210(1):315–330 DOI 10.1534/genetics.118.300900.
- Hoelzel AR, Shivji MS, Magnussen JE, Francis MP. 2006. Low worldwide genetic diversity in the basking shark (*Cetorhinus maximus*). *Biology Letters* 2(4):639–642 DOI 10.1098/rsbl.2006.0513.
- Hohenlohe PA, Bassham S, Etter PD, Stiffler N, Johnson EA, Cresko WA. 2010. Population genomics of parallel adaptation in threespine stickleback using sequenced RAD tags. *PLOS Genetics* 6(2):e1000862 DOI 10.1371/journal.pgen.1000862.
- Hohenlohe PA, Hand BK, Andrews KR, Luikart G. 2018. Population genomics provides key insights in ecology and evolution. In: Rajora O, ed. *Population Genomics*. Cham: Springer.
- **IUCN. 2017.** *IUCN red list of threatened species.* Version 2013.1. *Available at http://www.iucnredlist.org/about/overview.*
- Jones FC, Grabherr MG, Chan YF, Russell P, Mauceli E, Johnson J, Swofford R, Pirun M, Zody MC, White S, Birney E, Searle S, Schmutz J, Grimwood J, Dickson MC, Myers RM, Miller CT, Summers BR, Knecht AK, Brady SD, Zhang H, Pollen AA, Howes T, Amemiya C, Lander ES, Di Palma F, Lindblad-Toh K, Kingsley DM. 2012. The genomic basis of adaptive evolution in threespine sticklebacks. *Nature* 484(7392):55–61 DOI 10.1038/nature10944.
- Karl SA, Castro ALF, Lopez JA, Charvet P, Burgess GH. 2011. Phylogeography and conservation of the bull shark (*Carcharhinus leucas*) inferred from mitochondrial and microsatellite DNA. *Conservation Genetics* 12(2):371–382 DOI 10.1007/s10592-010-0145-1.
- Keeney DB, Heist EJ. 2006. Worldwide phylogeography of the blacktip shark (*Carcharhinus limbatus*) inferred from mitochondrial DNA reveals isolation of western Atlantic populations coupled with recent Pacific dispersal. *Molecular Ecology* 15(12):3669–3679 DOI 10.1111/j.1365-294X.2006.03036.x.
- Knapp ISS, Puritz JB, Bird CE, Whitney JL, Sudek M, Forsman ZH, Toonen RJ. 2016. ezRADan accessible next-generation RAD sequencing protocol suitable for non-model organisms_v3.2. *protocols.io* DOI 10.17504/protocols.io.e9pbh5n.

- Kofler R, Betancourt AJ, Schlötterer C. 2012. Sequencing of pooled DNA samples (Pool-Seq) uncovers complex dynamics of transposable element insertions in *Drosophila melanogaster*. *PLoS Genetics* 8(1):e1002487 DOI 10.1371/journal.pgen.1002487.
- Kofler R, Pandey RV, Schlötterer C. 2011. PoPoolation2: identifying differentiation between populations using sequencing of pooled DNA samples (Pool-Seq). *Bioinformatics* 27(24):3435–3436 DOI 10.1093/bioinformatics/btr589.
- Konczal M, Koteja P, Stuglik MT, Radwan J, Babik W. 2014. Accuracy of allele frequency estimation using pooled RNA-Seq. *Molecular Ecology Resources* 14(2):381–392 DOI 10.1111/1755-0998.12186.
- Kurland S, Wheat CW, De la Paz Celorio Mancera M, Kutschera VE, Hill J, Andersson A, Rubin CJ, Andersson L, Ryman N, Laikre L. 2019. Exploring a pool-seq-only approach for gaining population genomic insights in nonmodel species. *Ecology and Evolution* 9(19):11448–11463 DOI 10.1002/ece3.5646.
- Larsson LC, Charlier J, Laikre L, Ryman N. 2009. Statistical power for detecting genetic divergence—organelle versus nuclear markers. *Conservation Genetics* 10(5):1255.
- Lavery S, Moritz C, Fielder D. 1996. Genetic patterns suggest exponential population growth in a declining species. *Molecular Biology and Evolution* 13:1106–1113.
- Legendre P, Legendre L. 1998. Numerical Ecology. Second Edition. Amstrdam: Elsevier.
- Martin AP, Naylor G, Palumbi SR. 1992. Rates of mitochondrial DNA evolution in sharks are slow compared with mammals. *Nature* 357(6374):153–155 DOI 10.1038/357153a0.
- Mimee B, Duceppe MO, Véronneau PY, Lafond-Lapalme J, Jean M, Belzile F, Bélair G. 2015. A new method for studying population genetics of cyst nematodes based on pool-seq and genomewide allele frequency analysis. *Molecular Ecology Resources* **15(6)**:1356–1365 DOI 10.1111/1755-0998.12412.
- Momigliano P, Harcourt R, Robbins WD, Jaiteh V, Mahardika GN, Sembiring A, Stow A. 2017. Genetic structure and signatures of selection in grey reef sharks (*Carcharhinus amblyrhynchos*). *Heredity* **119**(3):142–153 DOI 10.1038/hdy.2017.21.
- Musick JA, Burgess G, Cailliet G, Camhi M, Fordham S. 2000. Management of sharks and their relatives (*Elasmobranchii*). *Fisheries* 25(3):9–13 DOI 10.1577/1548-8446(2000)025<0009:MOSATR>2.0.CO;2.
- Narum SR, Buerkle CA, Davey JW, Miller MR, Hohenlohe PA. 2013. Genotyping-by-sequencing in ecological and conservation genomics. *Molecular Ecology* 22(11):2841–2847.
- Neiman M, Taylor DR. 2009. The causes of mutation accumulation in mitochondrial genomes. *Proceedings of the Royal Society B: Biological Sciences* 276(1660):1201–1209 DOI 10.1098/rspb.2008.1758.
- Oliver S, Braccini M, Newman SJ, Harvey ES. 2015. Global patterns in the bycatch of sharks and rays. *Marine Policy* 54:86–97 DOI 10.1016/j.marpol.2014.12.017.
- **Ovenden JR, Berry O, Welch DJ, Buckworth RC, Dichmont CM. 2015.** Ocean's eleven: a critical evaluation of the role of population, evolutionary and molecular genetics in the management of wild fi sheries. *Fish Fisheries* **16(1)**:125–159 DOI 10.1111/faf.12052.
- Pardini AT, Jones CS, Noble LR, Kreiser B, Malcolm H. 2001. Sex-biased dispersal of great white sharks. *Nature* 412:139–140.
- Pazmiño DA, Maes GE, Green ME, Simpfendorfer CA, Hoyos-Padilla EM, Duffy CJA, Meyer CG, Kerwath SE, De-León PS, Van Herwerden L. 2018. Strong trans-Pacific break and local conservation units in the Galapagos shark (*Carcharhinus galapagensis*) revealed by genome-wide cytonuclear markers. *Heredity* 120(5):407–421 DOI 10.1038/s41437-017-0025-2.

- Perez-Enciso M, Ferretti L. 2010. Massive parallel sequencing in animal genetics: wherefroms and wheretos. *Animal Genetics* 41(6):561–569 DOI 10.1111/j.1365-2052.2010.02057.x.
- Portnoy DS, McDowell JR, Heist EJ, Musick JA, Graves JE. 2010. World phylogeography and male-mediated gene flow in the sandbar shark, Carcharhinus plumbeus. *Molecular Ecology* 19:1994–2010.
- Portnoy DS, Puritz JB, Hollenbeck CM, Gelsleichter J, Chapman D, Gold JR. 2015. Selection and sex-biased dispersal in a coastal shark: the influence of philopatry on adaptive variation. *Molecular Ecology* 24(23):5877–5885.
- **Puritz JB, Hollenbeck CM, Gold JR. 2014.** *dDocent*: a RADseq, variant-calling pipeline designed for population genomics of non-model organisms. *PeerJ* **2**:e431 DOI 10.7717/peerj.431.
- Qi H, Song K, Li C, Wang W, Li B, Li L, Zhang G. 2017. Construction and evaluation of a high-density SNP array for the Pacific oyster (*Crassostrea gigas*). *PLOS ONE* 12(3):e0174007.
- **Rice J, Harley S. 2013.** Updated stock assessment of the Silky sharks in the Western and Central Pacific Ocean. In: *SCientific Committee Ninth Regular Session*, 6–14 August 2013, Pohnpei, Federated States of Micronesia.
- **Rellstab C, Zoller S, Tedder A, Gugerli F, Fischer MC. 2013.** Validation of SNP allele frequencies determined by pooled next-generation sequencing in natural populations of a non-model plant species. *PLOS ONE* DOI 10.1371/journal.pone.0080422.
- RStudio Team. 2020. RStudio: integrated development environment for R. Boston: RStudio, PBC.
- Ryman N, Palm S. 2006. POWSIM: a computer program for assessing statistical power when testing for genetic differentiation. *Molecular Ecology Notes* 6(3):600–602 DOI 10.1111/j.1471-8286.2006.01378.x.
- Ryman N, Palm S, André C, Carvalho GR, Dahlgren TG, Jorde PE, Laikre L, Larsson LC, Palmé A, Ruzzante DE. 2006. Power for detecting genetic divergence: differences between statistical methods and marker loci. *Molecular Ecology* 15:2031–2045 DOI 10.1111/j.1365-294X.2006.02839.x.
- Schlötterer C, Tobler R, Kofler R, Nolte V. 2014. Sequencing pools of individuals—mining genome-wide polymorphism data without big funding. *Nature Reviews Genetics* 15(11):749–763 DOI 10.1038/nrg3803.
- Schultz JK, Feldheim KA, Gruber SH, Ashley MVH, McGovern TM, Bowen BW, Bruber SH, Ashley MVH, McGovern TM, Bowen BW. 2008. Global phylogeography and seascape genetics of the lemon sharks (genus Negaprion). Molecular Ecology 17(24):5336–5348 DOI 10.1111/j.1365-294X.2008.04000.x.
- Stevens JD, Bonfil R, Dulvy NK, Walker PA. 2000. The effects of fishing on sharks, rays, and chimaeras (chondrichthyans), and the implications for marine ecosystems. *ICES Journal of Marine Science* 57(3):476–494 DOI 10.1006/jmsc.2000.0724.
- Tallmon DA, Gregovich D, Waples R, Baker CS. 2010. When are genetic methods useful for estimating contemporary abundance and detecting population trends? *Molecular Ecology Resources* 10(4):684–692 DOI 10.1111/j.1755-0998.2010.02831.x.
- Theisen TC, Bowen BW, Lanier W, Baldwin JD. 2008. High connectivity on a global scale in the pelagic wahoo, *Acanthocybium solandri* (tuna family Scombridae). *Molecular Ecology* 17(19):4233–4247 DOI 10.1111/j.1365-294X.2008.03913.x.
- Therkildsen NO, Palumbi SR. 2017. Practical low-coverage genomewide sequencing of hundreds of individually barcoded samples for population and evolutionary genomics in nonmodel species. *Molecular Ecology Resources* 17(2):194–208 DOI 10.1111/1755-0998.12593.
- Toonen RJ, Andrews KR, Baums IB, Bird CE, Concepcion GT, Daly-Engel TS, Eble JA, Faucci A, Gaither MR, Iacchei M, Puritz JB, Schultz JK, Skillings DJ, Timmers MA,

Bowen BW. 2011. Defining boundaries for ecosystem-based management: a multispecies case study of marine connectivity across the hawaiian archipelago. *Journal of Marine Biology* **2011**:1–13 DOI 10.1155/2011/460173.

- Toonen RJ, Puritz JB, Forsman ZH, Whitney JL, Fernandez-Silva I, Andrews KR, Bird CE. 2013. ezRAD: a simplified method for genomic genotyping in non-model organisms. *PeerJ* 1(14):e203 DOI 10.7717/peerj.203.
- **Toews DPL, Brelsford A. 2012.** The biogeography of mitochondrial and nuclear discordance in animals. *Molecular Ecology* **21**:3907–3930.
- Veríssimo A, Sampaio Í, McDowell JR, Alexandrino P, Mucientes G, Queiroz N, da Silva C, Jones CS, Noble LR. 2017. World without borders—genetic population structure of a highly migratory marine predator, the blue shark (Prionace glauca). *Ecology and Evolution* 7:4768–4781 DOI 10.1002/ece3.2987.
- Weir BS, Cockerham CC. 1984. Estimating F-statistics for the analysis of population structure. *Evolution* 38(6):1358–1370.
- Whitney NM, Robbins WD, Schultz JK, Bowen BW, Holland KN. 2012. Phylogeography of the whitetip reef shark (*Triaenodon obesus*): a sedentary species with a broad distribution. *Journal of Biogeography* **39(6)**:1144–1156 DOI 10.1111/j.1365-2699.2011.02660.x.
- Zink RM, Barrowclough G. 2008. Mitochondrial DNA under siege in avian phylogeography. *Molecular Ecology* 17(9):2107–2121 DOI 10.1111/j.1365-294X.2008.03737.x.