Are giant clams (*Tridacna maxima*) distractible? A multi-modal study

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To properly assess risk, an animal must focus its attention on relevant external stimuli; however, attention can be reallocated when distracting stimuli are present. This reallocation of attention may interfere with an individual's ability to effectively assess risk and may impede its response. Multiple stimuli presented together can have additive effects as distractors, and these include stimuli in different modalities. Although changes in noise and water flow are detectable by some bivalves, this has not been studied in the context of risk assessment or distraction. We experimentally exposed giant clams (*Tridacna maxima*) to changes in water particle movement through underwater sound (motorboat noise) and increased water flow to determine whether these stimuli, individually or together, modified risk assessment or caused distraction. We found that clams responded to sound, flow, and their combination by increasing frequency of mantle retractions (a potential anti-predator response) when exposed to a stimulus. Sound alone did not change risk assessment in either the latency to close or to reemerge following closure. However, when exposed to both stimuli simultaneously, clams increased their latency to close. We suggest that clams perceive sound and flow in an additive way, and are thus distracted. Interestingly, and uniquely, clams discriminate these multimodal stimuli through a single sensory modality. For sessile clams, anthropogenic noise is detectable, yet unavoidable, suggesting that they be especially vulnerable to marine noise pollution.

Are giant clams (*Tridacna maxima*) distractible? A multi-modal study.

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Abstract

To properly assess risk, an animal must focus its attention on relevant external stimuli; however, attention can be reallocated when distracting stimuli are present. This reallocation of attention may interfere with an individual's ability to effectively assess risk and may impede its response. Multiple stimuli presented together can have additive effects as distractors, and these include stimuli in different modalities. Although changes in noise and water flow are detectable by some bivalves, this has not been studied in the context of risk assessment or distraction. We experimentally exposed giant clams (*Tridacna maxima*) to changes in water particle movement through underwater sound (motorboat noise) and increased water flow to determine whether these stimuli, individually or together, modified risk assessment or caused distraction. We found that clams responded to sound, flow, and their combination by increasing frequency of mantle retractions (a potential anti-predator response) when exposed to a stimulus. Sound alone did not change risk assessment in either the latency to close or to reemerge following closure. However, when exposed to both stimuli simultaneously, clams increased their latency to close. We suggest that clams perceive sound and flow in an additive way, and are thus distracted. Interestingly, and uniquely, clams discriminate these multimodal stimuli through a single sensory modality. For sessile clams, anthropogenic noise is detectable, yet unavoidable, suggesting that they be especially vulnerable to marine noise pollution.

Keywords Multimodal • Giant Clam • Multisensory • Anthropogenic noise

Introduction

41   Most animals experience some form of predation risk in their lifetimes (Lima & Dill 1990).

42   Individuals must be able to properly assess risk of predation, weighing the costs and benefits of

43   either remaining or fleeing (Caro 2005). Flight or other forms of protection are often associated

44   with large energetic costs; individuals must expend energy to flee and lose out on potential

45   energy that could have been obtained through feeding (Ydenberg & Dill 1986; Caro 2005).

46   Attention is the process of filtering out irrelevant stimuli from the environment in order

47   for the individual to focus on biologically relevant stimuli (Dukas 2004). Attention to tasks and

48   external stimuli is limited, divisible and is sensitive to different modalities (Dukas 2004; Chan et

49   al. 2010; Blumstein 2013). External stimuli can interfere with risk assessment by both masking

50   biologically important signals (Zhou, Radford, & Magrath 2019) and distracting individuals, thus

51   preventing them from focusing on avoiding predation (Rabin, Coss & Owings 2006; Karp &

52   Root 2009; Chan et al. 2010). Distraction can also impair performance on other relevant tasks

53   (Maes & de Groot 2003); potentially reducing feeding and reproductive opportunities. External

54   acoustic and visual stimuli are commonly studied as distractions that impede anti-predatory

55   responses in vertebrates (Grueninger & Pribram 1969; Dukas 2004; Wale, Simpson, & Radford

56   2013). However, there is less literature on distraction in invertebrates. A growing literature on

57   giant clams' (*Tridacna maxima*) risk assessment shows that they respond to both visual and

58   tactile threats (Stasek 1965; Wilkens 1986; Land 2002; Soo & Todd 2014; Johnson et al. 2016;

59   Dehaudt et al. 2019).

60   However, bivalves assess risk in other modalities as well. They alter their anti-predator

61   behaviors based on different chemical and auditory stimuli (Neo & Todd 2011; Castorani &

62   Hovel 2015; Peng et al. 2016). They possess labial palps on their mantle, gills and siphon that

63   can detect and respond to local water particle accelerations (Peng et al. 2016; Nedelec et al.

64   2016). Their palps also allow them to sense pressure changes from underwater sound waves and

65   bivalves have been documented to significantly react to changes in underwater acoustic stimuli

66   (e.g., Mosher 1972; Peng et al. 2016; Shi et al. 2019; Charifi et al. 2018; Roberts et al. 2015).

67   Anthropogenic noise is one such external stimulus that could distract bivalves. Marine

68   anthropogenic noise has been steadily rising over recent years due to shipping, motorboat

69   activity, and pile-driving (Hildebrand 2009). There remains an overall lack of study regarding

70   the detrimental effects of noise with respect to invertebrates (Morley, Jones, & Radford 2014)

71   and bivalves especially (Peng et al. 2016; Shei et al. 2019; Charifi et al. 2018; Roberts et al.

72   2015, Solan et al. 2016). Bivalves exposed to anthropogenic noise increase bioaccumulation of

73   toxic metals (Peng et al. 2016), alter their digging, sediment uptake and valve closure patterns

74   (Shei et al. 2019; Solan et al. 2016; Roberts et al. 2015) and experience reduced growth (Charifi

75   et al. 2018). However, these studies did not look specifically at how noise potentially distracts

76   bivalves from performing biologically important tasks. This research is critical because while

77   mobile animals can respond to noise pollution through movement and migration away from areas

78   of high intensity (Hirst & Rodhouse 2000; Slabbekoorn et al. 2010), sessile marine invertebrates

79   lack escape as an option. Since these species may have important roles in their community (e.g.,

80   Jordan & Valiela 1982; Carballo & Naranjo 2002; Neo et al. 2015; Solan et al 2016), we need to

81   know more about how they may be negatively impacted by anthropogenic stimuli. The little

82   research that has been done calls for *in situ* testing and further understanding of the distracting

83   effects of anthropogenic noise (Aguilar de Soto et al. 2013; Buscaino et al. 2016).

84   Acoustic stimuli propagate waves that oscillate water particles without creating a clear

85   directional flow (Nedelec, Campbell & Radford 2016); however these directional changes in

86   underwater flow could also distract bivalves. The effect of changes in water flow has been

87   investigated in bivalves in terms of how it influences their feeding (Levinton 1991; Pilditch &

88   Grant 1999). However, we are aware of no studies on changes in flow as a potential bivalve

89   distractor. Changes in water flow could indicate an increased potential for feeding; any increase

90   in underwater velocity could bring more edible plankton in the way of filter-feeding bivalves

91   (Levington 1991) thereby altering bivalves' risk versus reward perception. Flow changes could

92   also indicate changes in threat level because it is one of many factors naturally associated with

93   the presence of a moving predator (McHenry et al. 2009).

94   Giant clams may commonly experience both stimuli in a multimodal way. Multimodal

95   assessment may provide more information about threats and hence reduce uncertainty about risk

96   leading to an increase or decrease in anti-predator responses (Munoz & Blumstein 2012). Most

97   multimodal research has explored the additive effect of multiple stimuli as a multisensory

98   experience (Munoz & Blumstein 2012, 2019). Whereas a combination of sound and flow will

99   indeed be considered multimodal, giant clams experience changes in water particle motion

100 through a single sensory channel: their palps. If clams respond differently to multimodal

101 stimulation compared to unimodal stimulation, then this will indicate that giant clams could

102 modify their anti-predator behavior in a multimodal way and that this was achieved by a single

103 sensory pathway binding multimodal stimuli. Specifically, the presence of additional stimuli in

104 the form of sound or flow could potentially distract giant clams from the presence of a predator,

105 causing a delay in how fast they retract their mantle.

106   We asked whether changes in water particle motion due to unimodal sound, unimodal

107 water flow, or a multimodal combination of the two would distract giant clams or modify their

108 risk assessment. When perturbed, giant clams will partially or fully retract their mantle as a

109 potential anti-predatory response (Johnson et al. 2016; Dehaudt et al. 2019; Land 2002; Wilkens

110   1986). Thus, we used the latency to fully retract their mantle as an indicator of distraction levels

111   and the latency for the mantle to fully re-emerge from its shell as a measure of perceived risk.

112

113   **Materials & Methods**

114

115   **Study subjects.**

116   We marked giant clams ($n = 32$) for subsequent study within Gump Reef in Cook's Bay,

117   Mo'orea, French Polynesia (-17.482215, -149.827079), a marine protected area that has been the

118   site of previous studies of clam anti-predator behavior (Johnson et al. 2016; Dehaudt et al. 2019).

119   Each clam was >5m apart from each other. We measured each clam's depth and the length of the

120   each clam's shell. Clams received four total treatments over four separate days (with a planned

121   day off between subsequent treatments) in a Latin square design: control, flow only, sound only,

122   and flow and sound together. We initially planned to conduct all experiments over 8 days, but a

123   large storm generated considerable runoff which filled the bay with sediment and created poor

124   visibility making it impossible to study the clams. Thus, we waited until the water cleared and

125   thus conducted two treatments 3 days later than planned. We refer to the first two treatments as

126   pre-plume and the second two treatments as post-plume. Data were collected in the mornings,

127   except for one day in which data had to collected in the afternoon due to water clarity issues

128   from storm water runoff. Clams were studied under permits issued by the Government of French

129   Polynesia (permit approved on 21 November 2019).

130

131   **Flow Calibration.**

132 We used an underwater aquarium pump (1/55 HP, ECO-FLO, Ashland OH) to manipulate water

133 velocity over the clam. Before experimentation, we calibrated flow by video recording

134 (Crosstour CT7000 underwater video camera, Shenzhen, China) the time it took red dye to travel

135 10 cm through a clear, plastic tube. We measured the velocity to be 0.43 (± 0.06 SD) m/s (N = 10

136 measurements) where the jet exiting the nozzle of the pump was 15 cm away from the opening

137 of the tube.

138

139 **Sound Calibration.**

140 We used an Oceanears DRS-8 MOD 2 underwater speaker (North Canton, OH) to broadcast

141 recordings of outboard motorboats collected underwater. Five recordings of motorboat sounds of

142 similar amplitude were obtained from a previous study (Simpson et al. 2016). The five motorboat

143 recordings consist of a research vessel (5-m-long aluminum hull boat with 30 hp Suzuki

144 outboard motors) motoring 10-200 m away from the hydrophone at various speeds, mimicking

145 boat operations that are common in coral reef environments (Simpson et al. 2016).

146 To determine how the broadcast sounds compared to that of the reef environment, we

147 broadcast and re-recorded our exemplars underwater using a Wildlife Acoustics Song Meter

148 SM2+ (Concord, MA) with a Wildlife Acoustics hydrophone. We first calibrated the hydrophone

149 by playing 45 s of white noise of a known amplitude (90 dB re 20 µPa measured at 1 m) through

150 an Oceanears DRS-8 MOD 2 underwater speaker (North Canton, OH) above water. The white

151 noise was generated in Audacity (version 2.3.3, 2019) and was measured with a Radioshack (33-

152 2055) Digital Sound Level Meter set to weighting A (Radioshack, Fort Worth, Texas) placed 1

153 m away from the speaker. The white noise was then broadcast underwater 1 m from the

154 hydrophone at 0.5 m depth. We also similarly broadcasted the five motorboat exemplars (89-91

155   dB re 20 µPa measured at 1 m) against the calibrated hydrophone. In Praat (version 6.1.109,

156   2020), we calculated the average intensity (minimum pitch = 100 Hz) of the white noise and then

157   adjusted the scale to reflect the intensity measured above water to which we added 61.5 dB to

158   roughly estimate intensity re 1 µPa underwater (National Research Council 1994, Gausland

159   2000). This created a scaling factor that we used to adjust the amplitude of each motorboat

160   exemplar. Using Praat, we calculated the average intensity of our stimuli across exemplars,

161   which was 150.6 dB (± 1.03 SD) re 1 µPa.

162

163   **Stimulus presentation.**

164   Clams received four different treatments over four days: control (pump and speaker off), flow

165   only (pump on, speaker off), sound only (pump off, speaker on) and flow and sound (pump and

166   speaker on) in a Latin Square design. For all four treatments, the clam was subjected to the

167   presence of both the pump and speaker, whether on or off, for 60 s before it was perturbed. The

168   pump and speaker were attached to a 1.5 m PVC pipe. The hose from the pump was positioned

169   15 cm away from the clam while the speaker was 1 m away pointed at the center of the clam's

170   mantle. This ensured that the clam was subjected to 0.4 (± 0.07 SD) m/s of flow and 149.13 dB

171   (± 1.03 SD) re 1 µPa. Clams receiving sound or flow and sound treatment were randomly

172   exposed to one of five motor boat noises.

173   During this 60 s period before perturbation, we counted the number of mantle twitches

174   (partial retractions) as an indicator of perceived threat. The entire length of the clam's mantle

175   was then rubbed back and forth twice using a pencil's eraser attached to a different 2 m long

176   PVC pipe that permitted the person to be ≥ 1 m away. The latency to close was recorded using

177   either Crosstour CT7000 or a GoPro Hero 4 camera. The latency to re-open was measured both

178   in the field using a digital stopwatch, and using the camera. We defined latency to close as the

179   first sign of retraction of the mantle to the point where the clam is fully or almost closed. We

180   defined latency to reemerge as the time from the first contact of the pencil to the clam returning

181   to its previous undisturbed state that the observer has determined prior to the clam being rubbed.

182   If the clam did not fully reemerge within 3 min, we determined reemergence as the point where

183   its mantle was most fully extended. The camera recorded the entire process from the 60 s before

184   perturbation until the observer was confident the clam had fully reemerged. In 16 cases we were

185   unable to confidently measure the latency to re-open in the field. Due to the extremely high

186   correlation between field and camera playback latency times, we estimated these recorded times

187   from field observations using a prediction equation based on the correlation between

188   measurements (estimated time recorded from video = 1.0263 * stopwatch time + 1.9489; R^2=

189   0.98; $p < 0.001$, $N = 103$ measurements to estimate latency to re-emerge).

190   To reduce experimental interference, we did not record the precise velocity of ambient

191   water flow during the experiments. However, we noted flow as either negligible or noticeable by

192   observing pieces of macroalgae moving underwater after conducting each experimental

193   treatment. If pieces of macroalgae in the water column at the same depth of the clam moved in a

194   single direction, flow was marked as noticeable. We estimated an upper limit range of ambient

195   velocity during testing by recording the velocity at which a piece of macroalgae moved 10 cm

196   through a clear, plastic tube with a GoPro nine separate times during a day with particularly high

197   underwater flow. We estimated this velocity as 0.09 (± 0.064 SD) m/s, which was substantially

198   lower than our experimental velocity. Thus, we infer that our experimental velocity was

199   substantially greater than background changes in flow.

200

Statistical analysis.

Prior to analysis, we plotted histograms of variables to check for normality and $\log_{10}$ transformed retractions and latency to re-emerge.

To determine if stimuli were detected by giant clams, we fitted a linear mixed model with the number of retractions as the dependent variable using the lme4 version 1.1-21, (Bates et al. 2019) and lmerTest version 3.1-1 (Kuznetsova, Brokhoff, & Christensen 2017) packages in R version 3.6.2 (R Core Team 2019). Treatment, trial number, clam size, and plume presence (defined as before or after the disruptive plume) were added as fixed effects. To account for repeated measures, subject ID was added as a random effect. We tested the assumptions of our mixed model by plotting a histogram of residuals (they were approximately normal), plotting a q-q plot (they were approximately linear), and plotting fitted versus predicted values (there was no obvious pattern). We also compared our model to a linear model without subject ID as a random effect to determine if there was a significant difference in individuals' number of retractions. We estimated marginal means using the MuMIn version 1.43.15 (Barton 2019) and emmeans version 1.4.4 (Lenth et al. 2020) packages in R and estimated the d-score as a measure of treatment effect size. We plotted estimated marginal means using plotly version 4.9.2 (Sievert et al. 2020).

To determine if multimodal stimuli distracted giant clams, we fitted a linear mixed model as above with the latency to close, and another model with the latency to re-emerge as dependent variables. Again, we added treatment, trial number, clam size, and plume presence as fixed effects and added subject ID as a random effect. Model fit was determined by examining the normality of residuals in a histogram and quantile plot. Again, both models were compared to a

223   linear model without the random effect of subject ID. Estimated marginal means were plotted

224   with standard errors.

225   We tested for exemplar effects by creating a dataset containing only those where

226   motorboat noise was used and then fitting each of our models with sound exemplar as a fixed

227   factor to determine if any variation in response was explained by the specific exemplar used.

228   All data and code are contained in Supplementary Tables 1 and 2. Throughout, we set our

229   alpha to 0.05, and thus interpret p < 0.05 as significant.

230

231   **Results**

232   After controlling for significant random effects explained by individual (Chi-square = 2.83, 1 df,

233   $p = 0.007$), and non-significant fixed effects of trial number, plume presence and clam size, there

234   were more twitches recorded during a 60 s period when clams were exposed to sound, flow and

235   the combination of the two stimuli compared to the no stimuli treatment (Table 1a, Linear mixed

236   effects model, $p < 0.05$ for all). There were large effect sizes between the control and each

237   treatment (Fig. 1a, Supplementary Table 3; $d > 0.75$). Additionally, there were moderate effect

238   sizes $(0.75 > d > 0.4)$ between unimodal sound and unimodal flow as well as between unimodal

239   sound and multimodal flow and sound (Fig. 1a, Supplementary Table 3).

240   After controlling for significant random effects explained by individual (Chi-square =

241   10.848, 1 df, $p < 0.001$), significant fixed effects of clam size, and non-significant fixed effects

242   of trial number and plume presence, we found that the multimodal treatment (combination of

243   sound and flow) significantly increased clams' latency to close (Table 1b , Linear mixed effects

244   model, $p < 0.05$) while the other two unimodal treatments (flow only and sound only) had no

245   effect on latency to close when compared to the control (Table 1b, Linear mixed effects model, p

246   > 0.05). There was a small to moderate effect size between multimodal flow and sound and the

247   control; all other comparisons had small effect sizes (Fig. 1b, Table 1b; d = 0.35 and d < 0.3,

248   respectively).

249   After controlling for significant random effects explained by individual (Chi-square =

250   22.442, 1 df, p < 0.001), significant fixed effects of plume presence and non-significant fixed

251   effects of trial number and clam size, we found that clams did not significantly change their

252   latency to reemerge after treatments when compared to the control (Table 1c, Linear mixed

253   effects model, p > 0.05 for all). We found moderate effect sizes (0.75 > d > 0.4) between

254   unimodal flow treatment compared to all other treatments as well as the control.

255   When we tested for exemplar effects, variation in sound exemplars did not explain any

256   variation in clam response.

257

258   **Discussion**

259   Animals often use cues from a variety of modalities to reduce uncertainty about their

260   environment (Dall & Johnstone 2002), which potentially permits more precise assessments of

261   risk. Here we used two stimuli, from different modalities, either alone or together, to study anti-

262   predator responses in clams. We found that clams increased their frequency of partial retractions

263   during the 60 s acclimation period after exposure to any treatment. Clam mantle retraction is

264   presumably a discrete anti-predator response used for protection and to scare potential predators

265   with the water expelled from its siphon during retraction (Wilkens 1986; Land 2002).

266   While twitches were not complete valve closures, the twitches were still a physical response that

267   is closely linked to the species' highest level antipredator response—full mantle closure.

268   Increased anti-predator behavior indicates that motorboat noise may act as a giant clam stressor

269   (Wright et al. 2007). Previous studies have shown the adverse effects of anthropogenic noise on

270   bivalves' physiological processes such as metabolism and growth rate (Roberts et al. 2015; Peng

271   et al. 2016; Charifi et al. 2018; Shi et al. 2019). Thus, our results add to previous studies in

272   assessing whether the affected physiological processes in bivalves are the result of anthropogenic

273   noise-based distraction.

274   Sound and flow did not significantly alter clams' latency to close when tested as separate

275   unimodal treatments. However, when sound and flow were presented together as a multimodal

276   stimulus, clams closed more slowly, indicating that the presentation of multiple stimuli can be

277   distracting. Given that there was no significant effect of each stimulus when presented on its own

278   and that giant clams perceive both stimuli through their palps, a unisensory receptor, we

279   conclude that these stimuli have an additive effect on compromising clams' attention. These

280   results are consistent with the distracted prey hypothesis in which peripheral stimuli may

281   compromise an individual's attention from biologically necessary tasks (Chan et al. 2010). Less

282   vigilant prey become more likely targets for predators (Krause & Godin 1996), and a slower

283   closing time in clams encountering a multimodal stimulus indicates that distracted clams may be

284   more vulnerable to predatory damage.

285   Clams did not differ in their reemergence time between exposure to treatment and

286   control. However, the flow as a unimodal treatment had a modest, but not significant, decrease in

287   reemergence time from the control (Table 1c) and had a moderate to large effect size when

288   compared to unimodal sound and multimodal sound and flow. This modest decrease in

289   reemergence time in response to increased flow around the clam may indicate an increased

290   opportunity to filter feed because there are more sediments and nutrients moving through its

291   environment (Taghon, Nowell, & Jumars 1980; Trager, Hwang, & Strickler 1990; Skilleter &

292 Peterson 1994, Hardy & Hardy 1969, Hawkins & Klumpp 1995). Thus, clams may reemerge

293 faster to take advantage of increased nutrient availability. As a result, increased water flow

294 cannot be considered a neutral distractor. It is rather unique in that it is a biologically significant

295 stimulus, forcing clams to factor foraging opportunity into their decision-making process. Clams

296 must simultaneously manage predation risk and obtain energy through feeding behavior: two

297 potentially exclusive behaviors with their own set of benefits and consequences. If clams

298 reemerge sooner, they could potentially make the mistake of exposing themselves to predators;

299 however, if they stay closed for too long, they are missing out on feeding opportunities. Further

300 studies are needed that explore reemergence time of clams in areas with naturally high levels of

301 flow to determine if they behave differently from clams in areas, like our study site on Gump

302 Reef, with relatively low flow.

303   These findings further illustrate the negative effects that anthropogenic noise can have on

304 bivalves (Peng et al. 2016; Shi et al. 2019; Charifi et al. 2018; Roberts et al. 2015). These prior

305 studies identified alterations in essential processes such as digging behavior, metabolism, and

306 growth in bivalves (Peng et al. 2016; Shi et al. 2019; Charifi et al. 2018). Of the prior studies that

307 identified negative effects, none, until this one, specifically focused on anthropogenic noise as a

308 distraction. A recent meta-analysis reported anthropogenic noise negatively affects families

309 across the animal kingdom, potentially altering distribution, communication, foraging, and

310 homeostasis in these families (Kunc & Schmidt 2020). However, species' responses differ

311 widely based on ecological context and physical constraints (Wright et al. 2007). While mobile

312 species can avoid anthropogenic noise stressors (e.g., Francis, Ortega, & Cruz 2009; Francis et

313 al. 2011; Goodwin & Schriver 2011; Nowacek et al. 2007; Weilgart 2007, 2018), giant clams

314 are, for the most part, sessile and must endure any anthropogenic disturbance. Giant clams play

315   important ecological roles on coral reefs, including water filtration and nutrient sequestration

316   (Neo et al. 2015). Thus, anything that makes more clams more vulnerable to predators may

317   contribute to the loss of these important services that giant clams provide.

318   Multimodal stimuli are traditionally defined as separate stimuli that are perceived through

319   different sensory channels (Partan & Marler 2005). Many animals are able to detect multiple

320   stimuli across different modalities and create a unified percept by integrating them (Brown &

321   Magnavacca 2003; Partan et al. 2010). In contrast to all other studies we are aware of, giant

322   clams perceive multimodal stimuli through a single sensory modality, their palps, but

323   nevertheless integrated these stimuli. It is important to note that giant clams depend heavily on

324   visual cues for risk assessment (Wilkens 1986, Stasek 1965). Our manipulations did not create

325   any shading events or cause any visible turbulence, thus we infer that clams were not relying on

326   visual cues to sense the particle motion caused by noise or flow. We therefore suggest that local

327   changes in water flow and changes in underwater noise are indeed multimodal although

328   perceived through a single sensory channel by clams. The clams' ability to react differently to

329   either unimodal or multimodal treatments of sound and flow is a novel example of how animals

330   can bind multimodal stimulus together through a single sensory pathway to modify their risk

331   assessment.

332   Sound and flow changes sensed by clams could also be viewed as multi-attribute stimuli.

333   Unlike multimodal stimuli, which are typically perceived through different sensory channels,

334   multi-attribute stimuli are perceived through a single sensory channel. Multi-attribute stimuli are

335   often studied in bioacoustics (e.g., Gerhardt 1992; Castellano & Rosso 2007; Naguib & Wiley

336   2001; Gerhardt & Schul 1999; Kershenbaum et al. 2016). In these cases, different components of

337   an acoustic signal, including rate and frequency rises, can send multiple messages or further

338   inform an individual. Previous studies on crayfish (*Orconectes virilis*) demonstrated that animals

339   can use multiple separate chemical cues perceived through a single olfactory channel (Hazlett

340   1999; Bouwma & Hazlett 2001). Female Italian treefrogs (*Hyla intermedia*) weight separate

341   components of male mating calls differently when selecting a mate (Castellano & Rosso 2007).

342   Similarly, great tits (*Parus major*) assess multiple acoustic attributes of song (Weary 1990).

343   Female bluehead wrasse (*Thalassoma bifasciatum)* use multiple visual cues, such as coloration

344   and courtship display, to assess risk (Warner & Dill 2000). In honeybees, different olfactory

345   stimuli are processed through different clusters of glomeruli, enabling them to process different

346   components of odor through different pathways (Abel, Rybak, & Menzel 2001; Kirschner et al.

347   2006). The majority of these previous studies have focused on multi-attribute signaling and not

348   strictly multi-attribute risk assessment (Munoz & Blumstein 2012, 2019). In addition, these

349   previous studies focused mainly on terrestrial organisms; more studies are needed on multi-

350   attribute stimulus assessment in marine organisms. We have shown that giant clams can bind

351   together multiple stimuli into one sensory channel, an ability that we suspect is shared with other

352   species.

353

Conclusions

355   Giant clams increased their frequency of mantle retractions, a potential anti-predator behavior, in

356   response to motorboat playback, increased flow and the combination of the two. Clams also

357   slowed their closing times when exposed to a combination of sound and flow, indicating

358   distraction through multimodal means. These findings add evidence to the deleterious nature of

359   anthropogenic noise on sessile marine invertebrates and present a novel case of multimodal

360 stimuli processed together through a single sensory channel. We recommend further study of

361 unisensory systems and their relationships to traditionally multimodal stimuli.

362

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370

References

372 Abel R, Rybak J, Menzel R. 2001. Structure and response patterns of olfactory interneurons in
373   the honeybee, *Apis mellifera. Journal of Comparativve Neurology* **437**: 363-383.
374 Aguilar de Soto NA, Delorme N, Atkins J, Howard S, Williams J, Johnson M. 2013.
375   Anthropogenic noise causes body malformations and delays development in marine larvae.
376   *Science Reports* **3**:2831.
377 Barton, Kamil. 2019. MuMIn: multi-model inference. R package version, 1.43.6
378 Bates D, Maechler M, Bolker B, Walker S. 2014. Linear mixed-effects models using 'eigen' and
379   S4. *Journal of Statistical Software* **67**:1-48.
380 Blumstein DT. 2013. Attention, habituation, and antipredator behaviour: implications for urban
381   birds. In: Gil D, Brumm H (ed) *Avian urban ecology: behavioural and physiological*
382   *adaptations*. Oxford. 41-53
383 Bouwma P, Hazlett BA. 2001. Integration of multiple predator cues by the crayfish Orconectes
384   propinquus. *Animal Behaviour* **61**:771–776
385 Brown GE, Magnavacca G. 2003. Predator inspection behaviour in a characin fish: an interaction
386   between chemical and visual information? *Ethology* **109**:739–750
387 Buscaino G, Ceraulo M, Pieretti N, Corrias V, Farina A, Filiciotto F, Maccarrone V, Grammauta
388   R, Caruso F, Giuseppe A, Mazzola S. 2016. Temporal patterns in the soundscape of the
389   shallow waters of a Mediterranean marine protected area. *Science Report* **6**:34230.
390 Carballo JL, Naranjo S. 2002. Environmental assessment of a large industrial marine complex
391   based on a community of benthic filter-feeders. *Marine Pollution Bulletin* **44**:605–610
392 Caro T. 2005. *Antipredator defenses in birds and mammals*. University of Chicago Press;
393   Chicago, IL.

394 Castellano S, Rosso A. 2007. Female preferences for multiple attributes in the acoustic signals of
395   the Italian treefrog, *Hyla intermedia*. Behavioral *Ecology and Sociobiology* **61**:1293–1302
396 Castorani MCN, Hovel KA. 2015. Native predator chemical cues induce anti-predation
397   behaviors in an invasive marine bivalve. *Biological Invasions* **18**:169–181.
398 Chan AAYH, Giraldo-Perez P, Smith S, Blumstein DT. 2010. Anthropogenic noise affects risk
399   assessment and attention: the distracted prey hypothesis. *Biology Letters* **6**:458–461
400 Charifi M, Miserazzi A, Sow M, Perrigault M, Gonzalez P, Ciret P, Benomar, S, Massabuau J.
401   2018. Noise pollution limits metal bioaccumulation and growth rate in a filter feeder, the
402   Pacific oyster *Magallana gigas. PLOS One* **13**:e0194174.
403 Dall SRX, Johnstone RA. 2002. Managing uncertainty: Information and insurance under the risk
404   of starvation. *Philosophical Transactions of the Royal Society B Biological Sciences*
405   **357**:1519–1526.
406 Dehaudt B, Nguyen M, Vadlamudi A, Blumstein DT. 2019. Giant clams discriminate threats
407   along a risk gradient and display varying habituation rates to different stimuli. *Ethology*
408   **125**:392-398.
409 Dukas R. 2004. Causes and consequences of limited attention. *Brain, Behavior and Evolution*
410   **63**:197–210.
411 Francis CD, Ortega CP, Cruz A. 2009. Noise pollution changes avian communities and species
412   interactions. *Current Biology* **19**:1415–1419.
413 Francis CD, Paritsis J, Ortega CP, Cruz A. 2011. Landscape patterns of avian habitat use and
414   nest success are affected by chronic gas well compressor noise. *Landscape Ecology* **26**:1269–
415   1280.
416 Gausland, I. 2000. Impact of seismic surveys on marine life. *The Leading Edge* **19:**903-905.
417 Gerhardt HC. 1992. Multiple messages in acoustic signals. *Seminars in Neuroscience* **4**:391–400.
418 Gerhardt HC, Schul J. 1999. A quantitative analysis of behavioral selectivity for pulse rise-time
419   in the gray treefrog, *Hyla versicolor*. J*ournal of Comparative Physiology A* **185**:33–40.
420 Goodwin SE, Shriver WG. 2011. Effects of traffic noise on occupancy patterns of forest birds.
421   Conservervation Biology **25**:406–411.
422 Grueninger WE, Pribram KH. 1969. Effects of spatial and nonspatial distractors on performance
423   latency of monkeys with frontal lesions. *Journal of Comparative and Physiological*
424   *Psychology* **68**:203–209.
425 Hardy JT, Hardy SA. 1969. Ecology of *Tridacna* in Palau. *Pacific Science*. **23**: 467-472.
426 Hawkins AJS, Klumpp DW. Nutrition of the giant glam *Tridacna gigas* (L.). II. Relative
427   contributions of filter-feeding and the ammonium-nitrogen acquired and recycled by
428   symbiotic alga towards total nitrogen requirements for tissue growth and metabolism.
429   *Journal of Experimental Marine Biology and Ecology*. **190**: 263-290.
430 Hazlett BA. 1999. Responses to multiple chemical cues by the crayfish *Orconectes virilis*.
431   *Behaviour* **136**:161-171.
432 Hildebrand JA. 2009. Anthropogenic and natural sources of ambient noise in the ocean. *Marine*
433   *Ecology Progress Series* **395**:5–20.
434 Hirst AG, Rodhouse PG. 2000. Impacts of geophysical seismic surveying on fishing success.
435   *Reviews in Fish Biology and Fisheries*. **10**:113–118.
436 Johnson GC, Karajah MT, Mayo K, Armenta TC, Blumstein DT. The bigger they are the better
437   they taste: size predicts predation risk and anti-predator behavior in giant clams. *Journal of*
438   *Zoology* **301**:102–107.

439 Jordan TE, Valiela I. 1982. A nitrogen budget of the ribbed mussel, *Geukensia demissa,* and its
440   significance in nitrogen flow in a New England salt marsh. *Limnology and Oceanography*
441   **27**:75–90.
442 Karp DS, Root TL. 2009. Sound the stressor: How hoatzins (*Opisthocomus hoazin*) react to
443   ecotourist conversation. *Biodiversity and Conservation* **18**:3733–3742.
444 Kershenbaum A, Blumstein DT, Roch MA, Acay C, Backus G, Bee MA, Bohn K, Cao Y, Carter
445   C, Casar C, Coen M, DeRuiter SL, Doyle L, Edelman S, Ferrer-i-Cancho R, Freeberg TM,
446   Garland EC, Gustison M, Harley HE, Huetz C, Hughes M, Bruno JH, Ilany A, Jin DZ,
447   Johnson M, Ju C, Karnowski J, Lohr B, Manser MB, McCowan B, Mercado III, E, Narins
448   PM, Piel A, Rice M, Salmi R, Sasahara K, Sayigh L, Shiu Y, Taylor C, Vallejo EE. Waller S,
449   Zamora-Guiterrez V. 2016. Acoustic sequences in non-human animals: a tutorial review and
450   prospectus. *Biological Reviews* **91**:13–52.
451 Kirschner S, Kleineidam CJ, Zube C, Rybak J, Grunewald B, Rossler W. 2006. Dual olfactory
452   pathway in the honeybee, *Apis mellifera. Journal of Comparaative Neurology* **499**: 933-952.
453 Krause J, Godin JGJ. 1996. Influence of prey foraging posture on flight behavior and predation
454   risk: Predators take advantage of unwary prey. *Behavioural Ecology* **7**:264–271.
455 Kunc HP, Schmidt R. 2019. The effects of anthropogenic noise on animals: A meta-analysis.
456   *Biology Letters* **15**: 20190649.
457 Kuznetsova A, Brockhoff PB, Christensen RHB. 2017. Tests in linear mixed effects models.
458   *Journal of Statistical Software* **82**:1-26.
459 Land MF. 2002. The spatial resolution of the pinhole eyes of giant clams (*Tridacna maxima*).
460   *Proceedings of the Royal Society B: Biological Sciences* **270**:185–188.
461 Lenth R, Singmann H, Love J, Buerkner P, Herve M. 2020. Estimated marginal means, aka least-
462   squares means. URL https://CRAN.R-project.org/package=emmeans.
463 Levinton JS. 1991. Variable feeding behavior in three species of *Macoma* (Bivalvia: Tellinacea)
464   as a response to water flow and sediment transport. *Marine Biology* **110**:375–383.
465 Lima SL, Dill LM. 1990. Behavioral decisions made under the risk of predation: a review and
466   prospectus. *Canadian Journal of Zoology* **68**: 619-640.
467 Maes JHR, de Groot G. Effects of noise on the performance of rats in an operant discrimination
468   task. *Behavioural Processes*. **61**:57-68.
469 McHenry MJ, Feitl KE, Strother JA, Van Trump WJ. 2009. Larval zebrafish rapidly sense the
470   water flow of a predator's strike. *Biology letters*. **5**: 477-479.
471 Morley EL, Jones G, Radford AN. 2014**.** The importance of invertebrates when considering the
472   impacts of anthropogenic noise. *Proceedings of the Royal Society B: Biological Sciences*
473   **281**:20132683.
474 Mosher JI. 1972. The responses of *Macoma balthica* (Bivalvia) to vibrations. *Journal of*
475   *Molluscan Studies* **40**: 125-131.
476 Munoz NE, Blumstein DT. 2012. Multisensory perception in uncertain environments.
477   *Behavioural Ecology* **23**:457–462.
478 Munoz NE, Blumstein DT. 2019. Optimal multisensory integration. *Behavioural Ecology* **23**:
479   457-462.
480 Naguib M, Wiley RH. 2001. Review: Estimating the distance to a source of sound: mechanisms
481   and adaptations for long-range communication. *Animal Behavior* **62**:825–837.
482 National Research Council. 1994. *Low-Frequency Sound and Marine Mammals: Current*
483   *Knowledge and Research Needs.* Washington DC: The National Academies Press.

484 Nedelec SL, Campbell J, Radford AN, Simpson SD, Merchant ND. 2016. Particle motion: the
485   missing link in underwater acoustic ecology. *Methods in Ecology and Evolution* **7**:836–842.
486 Neo ML, Eckman W, Vicentuan K, Teo SLM, Todd PA. 2015. The ecological significance of
487   giant clams in coral reef ecosystems. *Biological Conservation* **181**:111–123.
488 Neo ML, Todd PA. 2011. Predator-induced charges in fluted giant clam (*Tridacna squamosa*)
489   shell morphology. *Journal of Experimental Marine Biology and Ecology.* **397**: 21-26
490 Nowacek DP, Thorne LH, Johnston DW, Tyack PL. 2007. Responses of cetaceans to
491   anthropogenic noise. *Mammal Review* **37**:81–115.
492 Partan SR, Fulmer AG, Gounard MAM, Redmond JE. 2010. Multimodal alarm behavior in
493   urban and rural gray squirrels studied by means of observation and a mechanical robot.
494   *Current Zoology* 56:313-326.
495 Partan SR, Marler P. 2005. Issues in the classification of multimodal communication signals. *The*
496   *American Naturalist* **166**:231–245.
497 Peng C, Zhao X, Liu S, Shi W, Han Y, Guo C, Jiang J, Wan H, Shen T, Guangxu L. 2016.
498   Effects of anthropogenic sound on digging behavior, metabolism, Ca 2+ /Mg 2+ ATPase
499   activity, and metabolism-related gene expression of the bivalve *Sinonovacula constricta*.
500   *Science Reports* **6**:24266.
501 Pilditch CA, Grant J. 1999. Effect of variations in flow velocity and phytoplankton concentration
502   on sea scallop (*Placopecten magellanicus*) grazing rates. *Journal of Experimental Marine*
503   *Biology and Ecology* **240**:111–136.
504 Rabin LA, Coss RG, Owings DH. 2006. The effects of wind turbines on antipredator behavior in
505   California ground squirrels (*Spermophilus beecheyi*). *Biological Conservation* **131**:410–420.
506 R Core Team. 2019. R: A language and environment for statistical computing. R
507   Foundation for Statistical Computing, Vienna, Austria. URL https://www.R-project.org/.
508 Roberts L, Cheesman S, Breithaupt T, Elliott M. 2015. Sensitivity of the mussel *Mytilus edulis* to
509   substrate-borne vibration in relation to anthropogenically generated noise. *Marine Ecology*
510   *Progress Series* **538**:185–195.
511 Shi W, Han Y, Guan X, Rong J, Du X, Zha S, Tang Y, Liu G. 2019. Anthropogenic noise
512   aggravates the toxicity of cadmium on some physiological characteristics of the blood clam
513   *Tegillarca granosa*. *Frontiers in Physiology* **10**:3389.
514 Sievert C, Parmer C, Hocking T, Chamberlain S, Ram K, Corvellec M, Despouy P. 2020. Create
515   Interactive WebGraphics via 'plotly.js'. URL https://CRAN.R-project.org/package=plotly.
516 Simpson SD, Radford AN, Nedelec SL, Ferrari MCO, Chivers DP, McCormick MI, Meekan
517   MG. 2016. Anthropogenic noise increases fish mortality by predation. *Nature*
518   *Communications* **7**:10544.
519 Skilleter GA, Peterson CH. 1994. Control of foraging behavior of individuals within an
520   ecosystem context: the clam *Macoma balthica* and interactions between competition and
521   siphon cropping. *Oecologia* **100**:268-278.
522 Slabbekoorn H, Bouton N, van Opzeeland I, Coers A, ten Cate C, Popper AN. 2010. A noisy
523   spring: the impact of globally rising underwater sound levels on fish. *Trends in Ecology and*
524   *Evolution* **25**:419–427.
525 Solan M, Hauton C, Godbold JA, Wood CL, Leighton TG, White P. 2016. Anthropogenic
526   sources of underwater sound can modify how sediment-dwelling invertebrates mediate
527   ecosystem properties. *Scientific Reports* **6**: 20540
528 Soo P, Todd PA 2014. The behaviour of giant clams (Bivalvia: Cardiiadae: Triadcninae). *Marine*
529   *Biology* **161**: 2699-2717.

530   Stasek CR. 1965. Behavioural adaptation of the giant clam *Tridacna maxima* to the presence of
531   grazing fishes. *Veilger* **8**:29-35.
532   Taghon GL, Nowell ARM, Jumars PA. 1980. Induction of suspension feeding in spionid
533   polychaetes by high particulate fluxes. *Science* **210**:562-564.
534   Trager GC, Hwang JS, Strickler JR. 1990. Barnacle suspension-feeding in variable flow. *Marine*
535   *Biology* **105**:117-127.
536   Wale MA, Simpson SD, Radford AN. 2013. Noise negatively affects foraging and antipredator
537   behaviour in shore crabs. *Animal Behavior* **86**:111–118.
538   Warner RR, Dill LM. 2000. Courtship displays and coloration as indicators of safety rather than
539   of male quality: the safety assurance hypothesis. *Behavioural Ecology* **11**:444-451.
540   Weary DM. 1990. Categorization of song notes in great tits: which acoustic features are used and
541   why? *Animal Behavior* **39**:450–457.
542   Weilgart LS. 2007. A brief review of known effects of noise on marine mammals. *International*
543   *Journal of Comparative Psychology* **20**:159-168.
544   Weilgart LS. 2018. The impact of ocean noise pollution on fish and invertebrates. Report for
545   OceanCare, Switzerland.
546   Wilkens LA. 1986. The visual system of the giant clam *Tridacna*: behavioral adaptations.
547   *Biological Bulletin* **170**:393-408.
548   Wright AJ, Hatch LT, Aguilar de Soto N, Baldwin AL, Bateson M, Beale CM, Clark C, Deak T,
549   Edwards EF, Fernandez A, Godinho A, Hatch LT, Kakuschke A, Lusseau D, Martineau D,
550   Romero ML, Weilgart LS, Wintle BA, Notarbartolo-di-Sciara G, Cortes Martin V. 2007.
551   Anthropogenic noise as a stressor in animals: A multidisciplinary perspective *Journal of*
552   *Comparative Psychology* **20**:250-273.
553   Ydenberg RC, Dill LM. 1986. The economics of fleeing from predators. *Advances in the Study*
554   *of Behavior* **16**:229–249.
555   Zhou Y, Radford AN, Magrath RD. 2019. Why does noise reduce response to alarm calls?
556   Experimental assessment of masking, distraction and greater vigilance in wild birds.
557   *Functional Ecology* **33**:1280–1289.

Results from linear mixed effects model on variation in clam response

This model explains variation in giant clam responses (a—number of retractions in 60 s, b—latency to close, c—latency to reemerge) following a multimodal stimulus experiment. The reference level was no stimulus.

1   **Table 1: Results from linear mixed effects model on variation in clam response**
2   This model explains variation in giant clam responses (a—number of retractions in 60 s, b—
3   latency to close, c—latency to reemerge) following a multimodal stimulus experiment. The
4   reference level was no stimulus.
5
6   **(a)** Number of retractions in 60 s

Fixed Effects	Estimate ± SE	P
Intercept	0.535 ± 0.144	<0.001
Trial Number	0.0198 ± 0.347	0.569
After Plume Presence	-0.111 ± 0.077	0.150
Size	-0.323 ± 1.18	0.787
Treatment		
Flow	0.280 ± 0.05	<0.001
Sound	0.169 ± 0.05	0.001
Flow + Sound	0.277 ± 0.05	<0.001
Subject ID	0.010	0.101

8   **(b)** Latency to close

Fixed Effects	Estimate ± SE	P
Intercept	2.490 ± 1.125	0.032
Trial Number	-0.081 ± 0.248	0.743
After Plume Presence	0.188 ± 0.550	0.733
Size	19.685 ± 9.437	0.046
Treatment		
Flow	0.177 ± 0.348	0.613
Sound	0.054 ± 0.361	0.882
Flow + Sound	0.749 ± 0.353	0.037
Subject ID	0.753	0.867

10   **(c)** Latency to reemerge

Fixed Effects	Estimate ± SE	P
Intercept	1.465 ± 0.271	<0.001
Trial Number	0.012 ± 0.051	0.816
After Plume Presence	0.368 ± 0.112	0.002
Size	0.331 ± 2.346	0.889
Treatment		
Flow	-0.114 ± 0.074	0.127
Sound	0.014 ± 0.079	0.859
Flow + Sound	0.024 ± 0.074	0.752
Subject ID	0.056	0.236

Figure 1

Figure 1

A comparison of estimated marginal means of (a) Log_{10}(number of partial retractions) (b) latency to close (c) Log_{10}(Latency to reemerge) between all treatments. Bars indicate one standard error from the means. Letters above bars indicate treatments that are significantly different from each other.